Temporal Dynamics of Air Bacterial Communities in a University Health Centre Using Illumina MiSeq Sequencing

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ABSTRACT

Bacterial contamination of air may have human health implications by the transmission of potential human pathogens. Therefore, assessment of air bacterial abundance and composition in different built environment is essential. Jawaharlal Nehru University health centre (UHC) is a primary healthcare setting providing need-based medication to university students. Using active air sampling method, we collected eight air samples from the indoor and outdoor area of UHC across four different seasons. The total genomic DNA was extracted from the air samples and subjected to 16S rRNA gene-based next-generation sequencing. We performed the taxonomic classification along with comparative analysis of air bacterial communities. This study revealed that *Proteobacteria*, *Actinobacteria*, *Bacteroidetes* and *Firmicutes* are the dominant phyla in the sampled air. Overall, the air bacterial composition in our studied samples was comparatively simple; only ten taxonomic families accounting for ~75% of the total sequences determined. We also observed ESKAPE pathogens in the air metagenomes in a low percentage (4.42%), which were dominated by *Pseudomonas*, *Acinetobacter* and *Staphylococcus*. *Proteobacteria*, *Actinobacteria* and *Firmicutes* showed significant correlation with PM_{2.5}. We suggest that routine air monitoring and microbiological survey is essential for air quality standards and potential human pathogens detection in health care settings. It is the first report from India to uncover the temporal dynamics of air bacterial communities in UHC using Illumina MiSeq (PE300) sequencing and Quantitative Insights into Microbial Ecology (QIIME).

Keywords: Air sampling; Particulate matter; Bacterial diversity; ESKAPE pathogens.

INTRODUCTION

University Health Centre (UHC) provides primary health care services to university students and residents of the university campus of Jawaharlal Nehru University (JNU). During unhealthy conditions, people visit UHC to get various medical facilities like consultation and lab investigation. The UHC microenvironment may possess aerosolised microorganisms which may have a potential health impact on patients as well as healthcare personnel. Therefore, the study of airborne bacterial diversity has a significant effect on public health assessment and infection control measures (Lax and Gilbert, 2015). The most-reported microorganisms involved in nosocomial infections are *Enterococcus faecium*, *Staphylococcus aureus*, *Klebsiella pneumoniae*, *Acinetobacter baumannii*, *Pseudomonas aeruginosa*, and *Enterobacter* species (ESKAPE pathogens) and most of them showed multidrug resistance (Santajit and Indrawattana, 2016).

Previous studies have characterised the airborne microbes

based on traditional culture method in the hospital environment (Qudiesat *et al.*, 2009; Gilbert *et al.*, 2010; Sudharsanam *et al.*, 2012; Matouskova and Holy, 2013; Park *et al.*, 2013; Gaüzère *et al.*, 2014; Frías-De León *et al.*, 2016). The culture-based methods only estimate a low proportion of total bacterial diversity in a particular environment (Pace, 1997). Therefore, culture-based studies fail to assess actual bacterial diversity of the hospital environment. To overcome these limitations, we must rely on culture-independent methods to determine the total bacterial diversity prevailing in a particular environment.

In recent years, quantitative polymerase chain reaction (qPCR) and next-generation sequencing (NGS) have been widely used to characterise bacteria in environmental samples using 16S rRNA gene (Lee *et al.*, 2010; Bertolini *et al.*, 2013). The NGS has emerged as a better tool for estimation of bacterial diversity in different indoor and outdoor environments (Jiang *et al.*, 2015; Gao *et al.*, 2017). Previously, pyrosequencing has been applied to explore the bacterial diversity in a healthcare setting (Luna *et al.*, 2007; Poza *et al.*, 2012). Illumina MiSeq (PE300) sequencing provides a more accurate estimation of the bacterial community dynamics compared to previously used DNA sequencing techniques.

In India, various research groups have studied the microbial contaminants present in air samples of different indoor and outdoor environments. In previous studies carried

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out in Rajasthan (Yadav *et al.*, 2007), Mumbai (Gangamma *et al.*, 2011), Nagpur (Jagzape *et al.*, 2013), Tamil Nadu (Srikanth *et al.*, 2008; Sudharsanam *et al.*, 2012), Gwalior (Yadav *et al.*, 2015), Agra (Mamta *et al.*, 2015), Chennai (Valsan *et al.*, 2015; Priyamvada *et al.*, 2018), Munnar (Valsan *et al.*, 2015) and Kolkata (Debasmita, 2011), researchers have studied biogenic materials by traditional culture, microscopy, PCR and MALDI techniques. Some researchers also characterised microbes from different indoor areas in hospital settings (Sudharsanam *et al.*, 2012; Bajpai *et al.*, 2014; Paul *et al.*, 2015). In addition, various research groups in Delhi (Yadav *et al.*, 2007; Srivastava *et al.*, 2012; Balyan *et al.*, 2017; Lal *et al.*, 2017; Balyan *et al.*, 2019) also tried to characterise different biogenic components (pollens, bacteria and fungi) of aerosol using culture-based and morphological identification methods. Srivastava and colleagues collected air samples from JNU health centre and used culture-based method to detect the Gram-positive bacteria and Gram-negative bacteria. As authors identified cultured bacteria based on their morphology and cell shape, they failed to report actual bacterial abundance and diversity (Srivastava *et al.*, 2012). As the NGS and metagenomics are emerging fields in the aerobiological research, its application will provide more valuable information, which might otherwise skip in these culture-based studies.

As in the Indian scenario, no data are available to study the air bacterial communities in health care setting using advanced 16S rRNA gene-based Illumina MiSeq sequencing, we have attempted to characterise the air bacterial communities from the indoor and outdoor area of the UHC using the same. For this purpose, air samples were collected using microbial air sampler during four different seasons, spring, monsoon, winter and summer. Next, the abundance and composition of bacterial communities present in the collected air samples were estimated. Further to gain knowledge on air bacterial diversity, α - and β -diversity was estimated. The abundance and composition of aerosolised nosocomial pathogens with special focus to ESKAPE pathogens were also investigated. Finally, correlations between the dominant phyla and the environmental factors (Temperature, Relative Humidity, PM_{2.5} and PM₁₀) were analysed.

METHODS

Air Sample Collection and Meteorological Conditions

We have collected air samples from the indoor and outdoor area of JNU Health Centre (28.54°N, 77.16°E) from March 2018 to June 2019 in different seasons, namely spring, monsoon, winter and summer (Fig. S1). We collected the air samples from indoor (Spring Indoor [SPI], Monsoon Indoor [MI], Winter Indoor [WI], and Summer Indoor [SUI]) and outdoor area (Spring Outdoor [SPO], Monsoon Outdoor [MO], Winter Outdoor [WO], and Summer Outdoor [SUO]) of the health centre during each season. Detailed information of sampling date, and corresponding Temperature (T) and Relative humidity (RH) using Fisherbrand Enviro-Meter, are presented in Table S1 (in supplementary material). We have also recorded the PM_{2.5} and PM₁₀ concentrations during sampling using Aerocet 831 Handheld Particle Counter, Met

One Instruments, Inc. (Table S1 in supplementary material).

We have collected the air samples in 12 ml of sterile collection liquid (PBS buffer, pH 7.4) using the microbial air sampler (Coriolis micro air sampler, Bertin Technologies, France) at an average flow rate of 300 L min⁻¹ for two hours and for each sampling, 12 runs of 10 mins each with a one-minute pause for refilling the collection cone was performed. All the sampler fittings were sterilised before sample collection to avoid potential contamination. After collection, the samples were maintained at 4°C and transported to the laboratory within an hour. The longer sampling time and high flow rate of the sampler combined with Illumina MiSeq sequencing could minimise the effect of single air sampling per season on our determination of bacterial abundance and community composition because air is continuously moving.

DNA Extraction and PCR Amplification

The collected air samples (now hydrosol) were concentrated by centrifugation at 12000 × g for 60 mins at 4°C (Jiang *et al.*, 2015) and the resulting pellet was resuspended in 4 mL of PBS buffer (pH 7.4). The genomic DNA was extracted using a DNeasy PowerLyzer kit as per the manufacturers' protocol (Qiagen, India). The concentration and quality of extracted DNA were measured by NanoDrop Spectrophotometer ND-2000c (Thermo Fischer Scientific, USA).

Extracted genomic DNA was amplified using 16SrRNA gene universal bacterial primers, 27F and 1492R (Devereux and Wilkinson, 2004), on a Thermocycler (BioRad, technologies, USA). The 50 µL of PCR mixture contained aseptic water (36.5 µL), 10X Taq polymerase buffer (5 µL), dNTP mix (1.0 µL), forward and reverse primers (1.0 µL), template DNA (5 µL) and Taq polymerase enzyme (0.5 µL). The PCR conditions were as follows: 95°C for 10 min, 25 cycles of 30 s at 95°C, 120 s at 45°C, 60 s at 72°C, and finally 10 min at 72°C. We carried all the PCR assay (samples, positive template control, negative template control) in triplicate. The size and quality of resulting amplicon were checked on 1% agarose gel using 100 bp gene ruler (Thermo Fischer Scientific, USA) and visualised on a UV transilluminator (Scientific Systems, New Delhi, India). Triplicates of PCR products were pooled and sent for 16S rRNA gene next-generation sequencing to Eurofins Genomics India Pvt. Ltd., Bengaluru, India.

Illumina MiSeq Library Preparation, Cluster Generation and Sequencing

The quality of PCR amplified DNA was determined using NanoDrop spectrophotometer. The amplicon libraries were prepared using the Nextera XT Index Kit (Illumina Inc.). The 16S rRNA gene-specific forward primer (5'-GCCTACGGGNGGCWGCAG-3') and 16S rRNA reverse primer (5'-ACTACHVGGGTATCTAATCC-3') were used to amplify the bacterial V3-V4 regions (Klindworth *et al.*, 2013). The paired-end libraries (2 × 300) were sequenced following the 16S metagenomic sequencing library preparation protocol (Part# 15044223 Rev.B). All the unassembled, high throughput sequencing reads obtained were submitted at NCBI Sequence Read Archive (SRA) as BioProject under the accession number of PRJNA589998.

Trimmomatic v0.38 (Bolger *et al.*, 2014) and FLASH (v1.2.11) (Magoč and Salzberg, 2011) were used to obtain the high-quality clean-reads and stitched the PE data into single reads based on the overlap. The merged sequences after removing adapter sequences, ambiguous reads and low-quality sequences were defined as ‘trimmed sequences’, which were filtered out the chimeras using the UPPARSE (Edgar, 2013).

We used the Quantitative Insights into Microbial Ecology (QIIME) pipeline for post-sequencing analysis (Caporaso *et al.*, 2010). We have picked Operational Taxonomic Units (OTUs) based on 97% sequence similarity within reads and chose a representative sequence from each OTU against the Greengenes database (v13_8) (McDonald *et al.*, 2012) for downstream analysis. This representative sequence was used for taxonomic identification of the OTU. OTU table was generated having representative sequences of OTUs with corresponding taxonomic ranks assigned against Greengenes database. We used `alpha_diversity.py` and `alpha_rarefaction.py` workflow script to calculate the alpha diversity (Shannon, Simpson and Chao1) and generate rarefaction curves, respectively. The QIIME script `beta_diversity_through_plots.py` was used to calculate the beta diversity to compare the different bacterial communities, which could be illustrated as the PCoA plots using weighted UniFrac diversity metrics (Lozupone and Knight, 2005). We have also calculated the Spearman’s rank correlation coefficient (SRCC), ρ , between the different bacterial phyla and the environmental factors (Temperature, Relative Humidity, $PM_{2.5}$ and PM_{10}).

RESULT AND DISCUSSION

Meteorological Conditions and Particulate Matter (PM) Concentration

The average temperature (T), relative humidity (RH), $PM_{2.5}$ and PM_{10} varied during the sampling across different seasons. The average T and RH during sampling periods were $(28.05 \pm 8.14)^\circ\text{C}$ and $(49.5 \pm 17.26)\%$, respectively (Table S1 in Supplementary material). The concentrations of $PM_{2.5}$ and PM_{10} ranged from $13.53 \mu\text{g m}^{-3}$ to $477.6 \mu\text{g m}^{-3}$ and $80.17 \mu\text{g m}^{-3}$ to $739.02 \mu\text{g m}^{-3}$, respectively (Table S1 in Supplementary material). The $PM_{2.5}$ concentration was highest in SUI and lowest in MI. On the other hand, the PM_{10} concentration was highest in WI due to the thick and persistent fog conditions during winter, which facilitates the prolonged suspension of these particulates in the air. However, the rain in monsoon brought down the concentration of particulate matter, including both $PM_{2.5}$ and PM_{10} .

Richness and Diversity Analysis across four Different Seasons

Based on the Shannon and Simpson diversity index, the bacterial diversity of indoor air was highest in SUI and lowest in MI. Furthermore, the bacterial diversity of outdoor air was highest in SUO and lowest in SPO (Table S3 in Supplementary material), which was also depicted by the rarefaction curves (Fig. S2 in Supplementary material). The Shannon indices of collected air samples were ranged from 3.2 to 9.1. The Shannon indices of SPO and MI were in proximity and also lower among the eight studied samples.

Similarly, the Shannon indices of WO and SPI samples are close to each other (Table S3 in Supplementary material). The rarefaction curves showed saturation level, indicating that most of the bacterial community was covered by the sequences obtained by MiSeq PE300 (Fig. S2 in Supplementary material).

We classified the obtained sequences in various taxa ranging from phylum to genus. A total of 34 phyla, 96 classes, 182 orders, 339 families, and 749 genera were retrieved (Table S4 in Supplementary material). For different air samples, the numbers of taxa determined at phylum, class, order, family and genus were in the range of 16–27, 39–78, 56–146, 115–283, 220–604, respectively. A proportion of sequences which could not be assigned to any known group, showing the occurrence of novel sequences. The proportions of unclassified taxa were 0.003%–0.162%, 0.045%–1.559%, 0.795%–9.353%, 10.789%–43.634% for class, order, family and genus, respectively (Table S5 in Supplementary material).

Bacterial Community Composition at Different Taxonomic Levels

The *Proteobacteria* was the most dominant phylum throughout all air samples, ranging from 39.21% to 78.04% of the sequences in the respective library of each air sample. *Actinobacteria*, *Bacteroidetes* and *Firmicutes* were also found in all air samples, but at the variable relative abundance in the sequences of each library (Fig. 1). Our findings are consistent with previous studies shown that *Proteobacteria*, *Actinobacteria* and *Firmicutes* were the most abundant microbes in the hospital air (Du *et al.*, 2018; Gao *et al.*, 2018). *Proteobacteria*, *Actinobacteria* and *Firmicutes* were also the dominant phyla in the urban $PM_{2.5}$ samples collected across the seasons (Du *et al.*, 2018; Li *et al.*, 2019).

Out of 96 classes detected across all samples, 20 of them were dominant (> 1% abundance at least one sample), as shown in Fig. 2, *Actinobacteria*, *Bacilli*, *Alphaproteobacteria*, *Gammaproteobacteria*, were shared by the eight samples with higher relative abundance than *Sphingobacteriia*, *Flavobacteriia*, *Betaproteobacteria*, *Planctomycetia*, *Bacteroidia*, *Oscillatoriothricaceae*. *Gammaproteobacteria* accounted for 38.64% of the total sequences. *Gammaproteobacteria* (44.41% of the total sequences) were dominated in the phylum *Proteobacteria*, followed by *Alphaproteobacteria* (8.79% of the total sequences) and *Betaproteobacteria* (4.27% of the total sequences). The other abundant classes with relative abundance more than 1% of the total sequences were *Actinobacteria* (14.15%), *Bacilli* (13.06%), *Sphingobacteriia* (6.51%), *Flavobacteriia* (5.09%), *Saprospirae* (1.53%), *Planctomycetia* (1.40%).

At the order level, 27 of total 185 orders were dominant (> 1% abundance at least one sample) and shared by all the eight air samples. As shown in Fig. 3, *Pseudomonadales*, *Actinomycetales* and *Bacillales* showed higher abundance with an average abundance of 36.8%, 14.15%, 12.09%, respectively. Another 8 orders with average abundance more than 1% of the total sequences were *Sphingobacteriales* (6.51%), *Flavobacteriales* (5.09%), *Burkholderiales* (4.24%), *Rhizobiales* (3.38%), *Rhodobacterales* (2.24%), *Caulobacterales* (1.89%), *Saprospirales* (1.53%), *Xanthomonadales* (1.46%).

Fig. 1. Distribution of different phyla in air samples collected from indoor and outdoor area of JNU Health Centre in each season: Winter Outdoor (WO), Winter Indoor (WI), Summer Outdoor (SUO), Summer Indoor (SUI), Spring Outdoor (SPO), Spring Indoor (SPI), Monsoon Outdoor (MO), and Monsoon Indoor (MI).

Fig. 2. Distribution of different classes in air samples collected from indoor and outdoor area of JNU Health Centre in each season: Winter Outdoor (WO), Winter Indoor (WI), Summer Outdoor (SUO), Summer Indoor (SUI), Spring Outdoor (SPO), Spring Indoor (SPI), Monsoon Outdoor (MO), and Monsoon Indoor (MI).

At the family level, 38 of 339 families were dominant (> 1% abundance at least one sample), as shown in Fig. 4. *Moraxellaceae* (33.66%) showed the highest abundance followed by *Micrococcaceae* (8.14%) and *Sphingobacteriaceae* (6.44%). Another 11 families with average abundance more than 1% of the total sequences were *Planococcaceae* (5.13%), *Exiguobacteraceae* (4.86%), *Flavobacteriaceae* (4.52%), *Oxalobacteraceae* (4.01%), *Pseudomonadaceae* (2.92%), *Rhodobacteraceae* (2.25%), *Microbacteriaceae* (1.92%), *Caulobacteraceae* (1.87%), *Chitinophagaceae* (1.51%), *Xanthomonadaceae* (1.40%), *Bacillaceae* (1.07%).

The 52 dominant genera (> 1% abundance at least one sample) were selected from a total of 750 genera, as shown in Fig. 5. *Psychrobacter*, *Moraxellaceae unclassified* and *Arthrobacter* showed higher abundance with an average abundance of 24.66%, 7.69%, 6.40%, respectively. *Psychrobacter* was the most abundant genera in all samples except MI, MO and SUO. The genus *Psychrobacter*, a member of the family *Moraxellaceae*, is a widespread and evolutionarily successful bacterial group, the biology of which may provide better insights of environmental adaptation and survival (Bowman, 2006). *Moraxellaceae unclassified* was abundant in all samples except SUO. *Arthrobacter* was abundant in all samples except WI. The genus *Arthrobacter*, a member of the family *Micrococcaceae*, have been isolated from soil and sediments, but some clinical isolates were also reported (Busse and Wieser, 2014). Another 16 genera with average abundance more than 1% of the total sequences were *Exiguobacterium* (4.69%), *Gillisia* (3.96%), *Planococcus* (3.93%), *Pedobacter* (3.39%), *Oxalobacteraceae unclassified* (2.72%), *Pseudomonas* (2.69%), *Sphingobacterium* (2.18%), *Caulobacteraceae unclassified* (1.74%), *Xanthomonadaceae unclassified* (1.21%), *Flavisolibacter* (1.20%), *Paracoccus* (1.16%), *Ralstonia* (1.08%), *Micrococcaceae unclassified* (1.01%). In similar line, it was reported that the dominant airborne bacterial genera identified in the hospital environment were *Staphylococcus spp.* (50%), *Micrococcus spp.* (15–20%), *Corynebacterium spp.* (5–20%), and *Bacillus spp.* (5–15%) (Kim et al., 2010). Moreover, *Staphylococcus* (51%) and *Micrococcus* (37%) were dominant among the bacterial genera in Portuguese hospital (Cabo Verde et al., 2015). A culture-based study of tertiary care centre from Central India reported Coagulase-negative *Staphylococci* (CoNS) followed by *Bacillus*, *Staphylococcus aureus* and *Pseudomonas aeruginosa* in air samples (Bajpai et al., 2014). *Micrococci*, CoNS, *Enterobacter* and *Pseudomonas* were the predominant in the hospital air collected from west-Chennai, India using passive and active methods (Sudharsanam et al., 2012). The hospital indoor air samples showed the predominance of *Escherichia coli*, *Staphylococcus aureus*, *Pseudomonas aeruginosa*, *Klebsiella sp.* at Kalyani, West Bengal, India (Paul et al., 2015).

Principal coordinate analysis (PCoA) was applied based on the weighted UniFrac distance matrix to compare the bacterial communities across the air samples. The P1 and P2 explained 49.12% and 19.55% of the variation in overall community structure, respectively (Fig. 6). We found that samples WO and SPO tended to cluster together. However,

the remaining samples were far from this cluster and ungrouped. This plot showed that the majority of samples were independent and different from each other (Fig. 6).

Correlation between Bacterial Community Structure, Meteorological Parameters (T, RH) and Particulate Matter (PM_{2.5} and PM₁₀)

Airborne microbes are the vital component of particulate matter and have a potential impact on human health (Zhai et al., 2018). Due to their small aerodynamics nature, PM_{2.5} and PM₁₀ loaded with airborne pathogens can directly entered into the human respiratory system, resulting in respiratory illnesses (Zhai et al., 2018). Here we tried to analyse the correlation between bacterial community structure (at phylum level), meteorological parameters (T, RH) and respirable particulate matter (PM_{2.5} and PM₁₀) concentration by calculating the SRCC across different seasons (in supplementary Table S2). *Actinobacteria* showed a positive correlation with temperature. *Proteobacteria* showed a positive correlation with PM_{2.5} concentration ($\rho = 0.857$, $p < 0.05$), suggesting that the relative abundance of *Proteobacteria* is in accordance with the concentration of PM_{2.5}. *Proteobacteria* has been a frequently reported dominant phylum of PM_{2.5} in China (Gao et al., 2017; Gao et al., 2018; Li et al., 2019; Pan et al., 2019). Interestingly, the *Actinobacteria* showed a negative correlation with both PM_{2.5} ($\rho = -0.893$, $p < 0.01$) and PM₁₀ ($\rho = -0.964$, $p < 0.01$). Firmicutes showed a negative correlation with PM_{2.5} concentration ($\rho = -0.893$, $p < 0.01$) only. These negative correlations can be explained by inter-phyla correlations as given in Table S7 in the supplementary material; where *Proteobacteria* showed a negative correlation with *Actinobacteria* ($\rho = -0.762$, $p < 0.05$). Since all the bacterial communities simultaneously existed in the air at the time of sample collection, the summation of relative abundance of each bacterial phyla in a given sample should be equal to 100. Therefore, the increase of the relative abundance of *Proteobacteria* with the increase in PM concentration might be indirectly resulting in the decrease of the abundance of *Actinobacteria* with the increase in PM concentration. All remaining phyla showed no significant correlation with meteorological parameters and particulate matter concentration. The SRCC analysis could explain the relationship between the bacterial community structure, meteorological parameters and particulate matter concentration; but failed to estimate the relative importance of these factors in shaping the bacterial community structure (Gao et al., 2017).

Distribution of ESKAPE Pathogens

Nosocomial infections are also known as hospital-associated infections (HAIs). The primary causative agents of HAIs are *Enterococcus faecium*, *Staphylococcus aureus*, *Klebsiella pneumoniae*, *Acinetobacter baumannii*, *Pseudomonas aeruginosa* and *Enterobacter spp.* (in short ESKAPE pathogens). Some of the nosocomial pathogens showed multi-drug resistance (MDR) due to improper use of broad-spectrum antibiotics in healthcare settings (Khan et al., 2015). In this study, we obtained 22056 ESKAPE sequences, accounting for 4.42%, of the total 498671 sequences. The number of ESKAPE sequences were ranged from 90 to

Fig. 4. Distribution of different families in air samples collected from indoor and outdoor area of JNU Health Centre in each season: Winter Outdoor (WO), Winter Indoor (WI), Summer Outdoor (SUO), Summer Indoor (SUI), Spring Outdoor (SPO), Spring Indoor (SPI), Monsoon Outdoor (MO), and Monsoon Indoor (MI).

Fig. 5. Distribution of different genera in air samples collected from indoor and outdoor area of JNU Health Centre in each season: Winter Outdoor (WO), Winter Indoor (WI), Summer Outdoor (SUO), Summer Indoor (SUI), Spring Outdoor (SPO), Spring Indoor (SPI), Monsoon Outdoor (MO), and Monsoon Indoor (MI).

Fig. 6. A Principal Coordinates plot (based on weighted UniFrac distances) showing variation in overall community structure Abbreviations. Winter Outdoor (WO), Winter Indoor (WI), Summer Outdoor (SUO), Summer Indoor (SUI), Spring Outdoor (SPO), Spring Indoor (SPI), Monsoon Outdoor (MO), and Monsoon Indoor (MI).

10323 across different air samples. The lowest sequences were reported in SPO and WO, WI and SPI, moderate in SUI and MO, and higher in SUO, MI (Table S6 in supplementary material). Reasonably, the climatic conditions during spring and winter are unfavourable for the bacterial growth, that is the reason we observed the lowest pathogen load during spring and winter seasons. In contrast, during summer and monsoon seasons, the pathogen load was higher due to the optimum climatic conditions for the bacterial growth.

Overall, we found 31 pathogenic species (known and unclassified) and their relative abundance were shown in Fig. 7(a). Among the 31 species, the relative abundance of unclassified species of *Pseudomonas*, *Acinetobacter* and *Staphylococcus* were high, which was 57.27%, 18.24% and 14.35% respectively. Airborne species of *Staphylococcus* and *Pseudomonas* showing multi-drug resistance were also reported from the neonatal intensive care unit (Morgado-Gamero et al., 2019). There were five ESKAPE pathogens, the relative abundances of which in the range of 1% to 3%, including *Acinetobacter schindleri* (2.98%), *Klebsiella unclassified* (1.37%), *Staphylococcus sciuri* (1.22%), *Enterococcus unclassified* (1.20%), *Acinetobacter lwoffii* (1.02%). Moreover, there were another 23 ESKAPE pathogens with the relative abundance less than 1%, i.e., *Acinetobacter johnsonii*, *Pseudomonas pseudoalcaligenes*, *Staphylococcus equorum*, *Pseudomonas alcaligenes*, *Staphylococcus epidermidis*, *Pseudomonas stutzeri*, *Acinetobacter rhizosphaerae*, *Enterococcus asini*, *Staphylococcus aureus*, *Enterobacter unclassified*, *Pseudomonas nitroreducens*, *Acinetobacter guillouiae*, *Enterobacter gergoviae*, *Staphylococcus saprophyticus*, *Staphylococcus pettenkoferi*, *Enterobacter cloacae*,

Enterococcus cecorum, *Pseudomonas viridiflava*, *Enterobacter cowanii*, *Enterobacter turicensis*, *Enterococcus casseliflavus*, *Pseudomonas fragi*, *Pseudomonas veronii*.

Fig. 7(b) depicted the relative abundance of 31 ESKAPE pathogens within each air sample. SPI and WO contained 13 ESKAPE pathogens. MI and MO contained 14 ESKAPE pathogens. There were 9, 15, 18 and 28 pathogens species in SPO, WI, SUI and SUO. *Pseudomonas unclassified*, *Staphylococcus unclassified*, *Acinetobacter unclassified*, *Acinetobacter lwoffii* were shared by all eight air samples. *Pseudomonas unclassified* was the dominant pathogenic species in MI, MO and SPI accounting for 97.50%, 88.33%, 56.70% of the pathogen sequences of each sample, respectively. *Staphylococcus unclassified* was dominant in SPO, WI and SUO, accounting for 60.00%, 39.28%, 34.44% of the pathogen sequences of each sample, respectively. *Acinetobacter unclassified* was the dominant pathogenic species in SUI and SUO, accounting for 59.24% and 41.11% of the pathogen sequences of each sample, respectively. *Staphylococcus sciuri* was shared by all samples except SPO sample. *Acinetobacter johnsonii* was shared by all samples except MO sample. *Enterococcus unclassified* was shared by all samples except MI and SPO. *Acinetobacter schindleri* was shared by all samples except SPO and WO. *Staphylococcus epidermidis* was shared by all samples except SPI, SPO and WO. *Staphylococcus aureus* and *Staphylococcus equorum* were shared by all samples except MI, MO and SUI. *Klebsiella unclassified* was shared by all samples except WI, WO and MO. *Pseudomonas stutzeri* was shared by all samples except WI, WO and SPO. *Enterococcus asini* was shared by all samples except SPI, SPO, MI and SUI. *Acinetobacter guillouiae* was shared by

Fig. 7(a). Relative abundance of ESKAPE pathogens of total pathogen sequences in air samples collected from indoor and outdoor area of JNU Health Centre.

Fig. 7(b). Distribution of ESKAPE pathogens in air samples collected from indoor and outdoor area of JNU Health Centre in each season: Winter Outdoor (WO), Winter Indoor (WI), Summer Outdoor (SUO), Summer Indoor (SUI), Spring Outdoor (SPO), Spring Indoor (SPI), Monsoon Outdoor (MO), and Monsoon Indoor (MI).

all samples except SPI, SPO, SUI and WI. *Pseudomonas pseudoalcaligenes* and *Pseudomonas alcaligenes* were shared by all samples except SPI, SPO, WI and WO. *Acinetobacter rhizosphaerae* only occurred in SUI, SUO and SPI. *Enterococcus cecorum* and *Enterococcus casseliflavus* have only occurred in SUO and WI. *Pseudomonas viridiflava* has only occurred in MI and MO. *Pseudomonas nitroreducens*, *Enterobacter gergoviae*, *Enterobacter cloacae* and *Enterobacter unclassified* only occurred in SPI and SPO. *Staphylococcus pettenkoferi*, *Staphylococcus saprophyticus* and *Enterobacter turicensis* have only occurred in SUO. *Pseudomonas veronii* has only occurred in WO. Previous studies also revealed that *Staphylococcus aureus* (~16%), coagulase-negative Staphylococci (13%–17.2%) and *Micrococcus luteus* (10.7%–13.3%) were found to be the most common airborne pathogens in healthcare settings (Qudiesat et al., 2009). Higher prevalence of clinically relevant bacterial pathogens, Coagulase-negative staphylococci (29.6%), *Staphylococcus aureus* (26.3%), *Acinetobacter* species (9.5%), and *Pseudomonas aeruginosa* (5.3%) were also reported in a hospital air in Ethiopia (Solomon et al., 2017). However, some of the species assigned to the ESKAPE pathogens category in this report are not yet recognised as pathogenic to the humans. Moreover, as the short reads were generated during Illumina MiSeq sequencing, the accurate classification of pathogens at species-level has remained obscure (Jin et al., 2018).

CONCLUSIONS

We used next-generation sequencing to investigate the bacterial abundance and diversity in UHC during different seasons. *Proteobacteria*, *Actinobacteria*, *Bacteroidetes* and *Firmicutes* were the dominant phyla in the sampled air. Overall, air bacterial composition is comparatively simple in this study; only ten taxonomic families were accounting for ~75% of the total sequences determined. The longer sampling time and high flow rate of the sampler, combined with Illumina MiSeq sequencing could minimise the limitation of single sampling per season to determine the seasonal dynamics of air bacterial communities because air is continuously moving (Gao et al., 2017). However, multiple sampling in each season might have provided a better estimate of season-wise bacterial abundance and diversity. Also, the season-wise bacterial abundance in the air needs to be confirmed using qPCR of 16S rRNA gene (Lee et al., 2010).

We also found 31 ESKAPE pathogens in extremely low percentage (4.42%) of entire sequences reported and dominated by unclassified species of *Pseudomonas*, *Acinetobacter* and *Staphylococcus*.

Actinobacteria showed significant correlation with temperature, PM_{2.5} and PM₁₀. Whereas, *Proteobacteria* and *Firmicutes* showed significant correlation with PM_{2.5} only. Exposure to these airborne bacterial pathogens may result in the emergence of respiratory ailments in human beings. The holistic approach, including administrative and environmental control, as well as personal protective measures, may regulate the airborne bacterial infections in healthcare settings.

ACKNOWLEDGEMENTS

KM is thankful to UGC (UPE-II: ID 59) and DST-PURSE (Phase-II) for financial assistance. We acknowledge Dr. Sudesh Yadav for providing Coriolis Micro air sampler and Met One Digital Particle Counter 831 (brought under UGC-UPE-II: ID 98). We also acknowledge the JNU Administration for granting permission to collecting the air samples from the Health Centre. MKC is grateful to Council of Scientific & Industrial Research (CSIR), India for providing Senior Research Fellowship and Contingency (File No: 09/263(1073)/2015-EMR-I). MKC is also thankful to Ms. Sonali Pardhiya, SES, JNU, for helping in Air sampling and providing valuable suggestions for manuscript preparation. MS thanks DST for WOS-A research grant (SR/WOS-A/LS-36/2018).

DISCLAIMER

The authors declare no conflict of interest.

SUPPLEMENTARY MATERIAL

Supplementary data associated with this article can be found in the online version at <http://www.aaqr.org>.

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Received for review, November 28, 2019

Revised, April 14, 2020

Accepted, April 15, 2020