

Impact of using Tobacco Heating System (THS) on indoor air quality in a night club

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Abstract

Heating tobacco instead of burning offers a potentially lower risk of delivering nicotine compared to conventional cigarettes. The majority of heat-not-burnt tobacco product studies have been related to the mainstream aerosol, but the data on the second hand aerosol is yet limited to several chamber studies, impact on indoor air quality in real life settings has not been yet reported. Here we assess pollution generated by tobacco heating system in a hospitality venue. Volunteers used tobacco heating system in a night club during non-operation hours. Additionally, indoor air quality was assessed while club in operation. Real-time aerosol particle concentration and off-line carbonyl, nicotine and 3-ethenylpyridine concentration was measured. Observed levels of particle number concentration was as follows: background at $1\text{E}+4 \text{ \# cm}^{-3}$, 10 users at $5\text{E}+4 \text{ \# cm}^{-3}$, 30 users at $1\text{E}+5 \text{ \# cm}^{-3}$ and club in operation at $1\text{E}+6 \text{ \# cm}^{-3} - 1\text{E}+7 \text{ \# cm}^{-3}$. This represents an increase by an order of magnitude in each subsequent scenario. The club featured relatively high background concentrations of gaseous pollutants presumably due to the third hand smoke. The impact of the usage of THS in non-operating club to the gaseous pollutant variation was much lower if at all significant compared to particle concentrations. The usage of tobacco heating system resulted in increasing particle number and mass concentrations from the background by an order of magnitude, but yet were an order of magnitude lower compared to club in operation. The majority of gaseous pollutant concentration were not affected, except nicotine.

Keywords: indoor air quality, heat not burnt, environmental tobacco aerosol, nicotine, hospitality venue.

39 INTRODUCTION

40

41 Tobacco use is the leading cause of global preventable morbidity and mortality. In spite of much
42 evidence on the adverse health effects of tobacco use, many people continue to smoke (Xiangyu
43 et al., 2018). Tobacco cigarette use is a highly addictive habit, making cessation a difficult and
44 challenging task (Farsalinos et al., 2017). There is an increasing focus on developing harm
45 reduction strategies to address the health risks of conventional cigarettes (CCs) (Lüdicke et al.,
46 2018). Electronic cigarettes (ECs) are currently the most popular harm-reduction products, but
47 many smokers do not find them sufficiently effective or satisfactory in fully substituting smoking
48 (Farsalinos et al., 2017). Heating tobacco instead of burning can offer a potentially lower risk of
49 delivering nicotine compared to CCs because it creates a less complex aerosol than burned
50 tobacco (Jaccard et al., 2017). Moreover, research on heat-not-burnt (HnB) tobacco products (or
51 alternatively called Tobacco Heating System (THS), Novel Tobacco Product (NTB)) has
52 consistently demonstrated that harmful and potentially harmful constituents (HPHCs) are reduced
53 or absent in the aerosols of heated tobacco. Simonavicius et al. reviewed 31 publication on HnB
54 products and concluded that compared with CCs, HnB products delivered up to 83% of nicotine
55 and reduced levels of harmful and potentially harmful toxicants by at least 62% and particulate
56 matter (PM) by at least 75% (Simonavicius et al., 2018). Most of studies however were related to
57 the mainstream aerosol. The data on the second hand HnB aerosol is yet limited to several studies
58 (Mottier et al. (2016); Mitova et al. (2016); Protano et al. (2016); Ruprecht et al. (2017);
59 Ichitsubo et al. (2018); Meišutovič-Akhtarjeva et al. (2019)).

60 The concentrations of most investigated indoor air constituents during the use of HnB products in
61 an environmentally controlled environments are similar to background levels (with the exception
62 of aldehydes) and an order of magnitude lower than in the CC environmental aerosol (Mitova et
63 al., (2016); Ichitsubo et al. (2018); Ruprecht et al. (2017); Meišutovič-Akhtarjeva et al. (2019)).

64 The review by Kaunelienė et al. (2018) provided a comparative analysis of THS generated
65 pollution against general indoor air quality (IAQ) in various micro-environments, especially with
66 combustion-based pollution sources present. The use of THS (as well as EC) in the controlled
67 environment was found to result in the lowest concentrations of formaldehyde, benzene, toluene,
68 and PM_{2.5} (particles with aerodynamic diameter lower than 2.5 μm) among majority researched
69 pollution sources (CC, water-pipe, incense, mosquito coils). The exposure to significantly higher
70 pollution levels of benzene, toluene and formaldehyde may occur in public environments,
71 especially transport micro-environments.

72 Following the adoption of WHO Framework Tobacco Control Convention in 2003, smoking bans
73 in public places have been introduced in many countries (WHO, 2017). A wide range of
74 regulatory responses ranging from no regulation to complete bans were also applied to the harm-
75 reduction products (Etter et al., 2011). Bans against the use of nicotine containing products in
76 public spaces at many countries is possibly the main reason that the evidence on the passive
77 exposure to the vapour released or exhaled from the harm-reduction products under real
78 conditions is still scarce. Only four studies on the impact of EC to IAQ performed using human
79 volunteers in natural settings were reported in a systematic review by Abidin et al. (2017).

80 The impact of HnB products on IAQ in real life settings has not been investigated so far. The aim
81 of this study was to simulate the variation of IAQ following the usage of HnB product (namely,
82 THS) in a non-operating night club with varying amount of users, and compare it against the
83 typical IAQ in a night club under operation.

84

85

86 **METHODS**

87 ***Test product***

88 Tobacco heating system THS 2.2 (further referred to as IQOS) with HEETS Bronze Label fillings
89 (Philip Morris International Inc., USA) were used.

90

91 ***Premises***

92 The experiment was conducted in a night club in Kaunas, Lithuania, with the floor area of approx.
93 160 m² and the volume of approx. 880 m³. The club premises may be subjectively divided to
94 several areas, including wardrobe, bar area, DJ scene, the main dancing floor, and seating areas
95 (Fig. S1). The facility was equipped with a mechanical ventilation system, supplying the air via
96 air handling unit equipped with rotary heat exchanger, 9100 m³ h⁻¹ maximum air flow. However,
97 this air handling unit was operated manually, while most of the time the club has been operated in
98 the natural ventilation regime (0.5 ACH as determined by the CO₂ concentration decay after the
99 closure). The indoor of club was a non-smoking area following a smoking ban in public premises
100 in Lithuania. The smoking room was located outdoors with the entrance from the main club area.
101 During operation the club was attended by 200 – 400 people at the time, while the smoking room
102 may accommodate approx. 30 humans. No cooking activities took place in a club, only drinks
103 and occasionally cold appetizers were served.

104

105 ***Experimental procedure***

106 The experiment was performed in February 2018 during the course of two campaigns. The first
107 campaign was carried out in non-working hours of the club to evaluate THS generated pollution
108 while avoiding presence of other active pollution sources. Three days of experiment, each

109 consisting of five measurement sessions (Campaign #1): background; background with 10
110 humans present but without THS use; 10 humans simultaneously using THS; background with 30
111 humans present but without THS use; 30 humans THS used by scattered all over the main club
112 area. Each measurement session took 30 minutes. The experiments were conducted with only
113 natural ventilation present (air handling unit off), since this reflects usual condition of a nightclub.
114 After the measurement session, particle and CO₂ concentration were reduced to the background
115 levels after both IQOS usage sessions via purging indoor air by running air handling unit at 75%
116 capacity (equivalent to 7-8 ACH) for 30 min. Indoor air was sampled at the breathing height (1.5
117 m from the floor) at two locations, representing different type of occupancy: the main (dance
118 floor) area of the club (further referred to as Zone 1) and sitting area (Zone 2) (see Fig.S1). Zone
119 2 was located under a balcony thus restricted airflows around it, presumably decreasing the
120 dispersion rate of the exhaled aerosol.

121 The second campaign was carried out for three days (Campaign #2). During each day, one hour
122 before the club opening (representing background) and 3 hours of club in operation were sampled
123 in order to obtain the profile of IAQ occurring during typical club activities. During the third day,
124 the real-time measurements were extended for 4 additional hours to obtain variation of pollutant
125 concentrations until and after the closure of the club. Similar sampling locations were employed
126 as in Campaign #1. The ventilation during Campaign #2 was adjusted manually running air
127 handling unit at 25% (equiv. of 2.6 ACH) capacity with no attendees to 100% (equiv. of 10 ACH)
128 capacity with the maximum attendees (~400 people).

129

130 ***Analytics***

131 The real-time particle number concentration (PNC), CO₂ concentration, relative humidity, and
132 temperature, as well as off-line carbonyls (acetaldehyde and formaldehyde), off-line nicotine and
133 3-ethenylpyridine concentration were analysed.

134 The real-time size-segregated particle concentration have been measured using the electrical low
135 pressure impactor (ELPI+, Dekati, Inc., Finland), at a flow rate of 10 l min⁻¹. Real-time
136 concentrations of aerosol samples have been registered in 1 Hz or 10 Hz intervals. The sampling
137 airflow was being switched between Zone 1 and Zone 2 using a 3-way switching valve every 5
138 minutes. The sample was delivered to particle sampler using Tygon® formula E-3603 laboratory
139 tubing. Such tubing has been confirmed as the best compromise considering particle losses
140 (Asbach et al., 2016).

141 The values of carbon dioxide (CO₂), temperature, and relative humidity have been continuously
142 recorded by air quality meters (7545 IAQ-CALC, TSI Inc., USA).

143 Concentration of nicotine and 3-ethenylpyridine (3-EP) has been estimated following the ISO
144 18145 procedure which is based on the collection of compounds by adsorption on a sorbent resin
145 (XAD®-4.7 x 70-mm size, 2 sections, 40/80 mg sorbent, 20/40 mesh, SKC Inc.), extraction and
146 determination by gas chromatography-mass spectrometry (GC-MS-QP2010 Ultra, Shimadzu
147 Corp., Japan). Concentration of formaldehyde and acetaldehyde has been estimated following the
148 ASTM D5197-03 procedure based on sample collection on a silica gel coated with 2,4-
149 dinitrophenylhydrazine (DNPH) reagent (SKC Inc.). The DNPH derivatives were analysed for
150 parent aldehydes utilizing high performance liquid chromatography (HPLC-DAD (UV)).

151

152 ***Data processing and analysis***

153 The measured distributions were based on PNC (unit particles/cm³ or # cm⁻³). The particle size
154 distribution (PSD) was based on the number density (concentration) distribution function, which

155 represented particle concentration normalized to the particle size bin where it was measured, i.e.,
156 the number of particles per volume of air with sized between D_p and dD_p . This is usually
157 expressed mathematically as $D_p = dN/d\log D_p$ ($\# \text{ cm}^{-3}$). The PSD data was adjusted for particle
158 losses in the sampling lines due to diffusion and inertial/gravitational losses for each channel
159 separately.

160

161 **RESULTS AND DISCUSSION**

162 *Time resolved variation of aerosol concentrations*

163 Highly time resolved aerosol concentration has been measured across the entire experiment,
164 including both controlled usage of IQOS (Campaign #1), as well as the uncontrolled operation of
165 the club (Campaign #2). The time series of PNC during the two measurement campaigns are
166 shown in Fig. 1.

167 The controlled experiment (Campaign #1) started with the background, having the median
168 concentration of particles $9.4\text{E}+3 \# \text{ cm}^{-3}$ and $9.8\text{E}+3 \# \text{ cm}^{-3}$ in Zones 1 and 2, respectively. This is
169 a moderate concentration of particles observable in many general indoor living environments,
170 usually ranging within $1\text{E}+3$ to $9\text{E}+4 \# \text{ cm}^{-3}$ (Fromme, 2012). This background was measured in
171 the premises where no continuous activities were present after a several days of workweek
172 closure before the opening in the weekend. The entry of 10 volunteers did not affect the
173 background concentration, with the median concentrations remaining at $9.2\text{E}+3 \# \text{ cm}^{-3}$ and
174 $9.9\text{E}+3 \# \text{ cm}^{-3}$ in Zones 1 and 2, respectively. Moreover, no major variation of particle
175 concentration has been registered. Human movement may have resulted into resuspension of dust,
176 but such super-micrometer particles were not reflected well in total particle concentration
177 variation, possibly due to losses in the sampling line transport.

178 Once the ten volunteers started using IQOS devices, the concentration of particles has increased,
179 as reflected by the first peak (Fig.1(A). The maximum has been registered at $1.2E+5 \text{ \# cm}^{-3}$,
180 while the median concentrations were at $3.6E+4 \text{ \# cm}^{-3}$ and $3.5E+4 \text{ \# cm}^{-3}$ in Zones 1 and 2,
181 respectively. This is a statistically significant increase ($p<0.05$). Such level of particles is
182 comparable to our simulation measurements in a room-scaled chamber with one volunteer and
183 ventilation rate 0.5 (Meišutovič-Akhtarieva et al., 2019). At the same time, this is relatively low
184 level compared to other hospitality environments even with smoking absent. In Germany very
185 high median PNCs of $2.2E+5 \text{ \# cm}^{-3}$ were measured during four main visiting hours in non
186 smoking areas in four cafés/restaurants, $1.1E+5 \text{ \# cm}^{-3}$ in two bars and $2.9E+5 \text{ \# cm}^{-3}$ in seven
187 discos (Bolte et al., 2008).

188 Once usage of 10 IQOS stopped, the concentration started decaying due to natural ventilation and
189 further due to purging with forced ventilation to level of $4E+4 \text{ \# cm}^{-3}$ (CO_2 level between 460-
190 480 ppm), which served as a background concentration for 30 volunteers present in the premises.

191 Thirty IQOS users resulted in another significant increase of PNC to maximum value of $1.5E+5 \text{ \#}$
192 cm^{-3} , and median of $1.2E+5 \text{ \# cm}^{-3}$ in Zone 1 and $1.3E+5 \text{ \# cm}^{-3}$ in Zone 2. Such levels are
193 comparable of the one registered in room chamber with 5 volunteers using IQOS (Meišutovič-
194 Akhtarieva et al., 2019) and in the higher range of hospitality environments such as German
195 restaurants, pubs and discotheques, ranging from $1.2E+5 \text{ \# cm}^{-3}$ to $2.1E+5 \text{ \# cm}^{-3}$ (Bolte et al.,
196 2008). The concentration has not decreased quickly to the background. This did not agree with
197 our findings from the simulations in a room chamber, where we observed a rapid decay of
198 concentrations to background levels, and attributed it to the short life span of volatile particles.
199 Such difference from chamber findings is probably due to a much more intensive emission
200 having 30 simultaneous users, as well as lower ventilation, having air handling unit off (air
201 exchange rate approx. 0.5 ACH) during the test.

202 The preparations for the club opening (Campaign #2) already resulted in high background values
203 (of $2.7\text{E}+5 \text{ \# cm}^{-3}$ in Zone 1 and $2.9\text{E}+5 \text{ \# cm}^{-3}$ in Zone 2, Fig. 1(B)). Yet the club operation
204 resulted in the highest PNC (max $1.7\text{E}+7 \text{ \# cm}^{-3}$, median $7.3\text{E}+5 \text{ \# cm}^{-3}$ in Zone 1 and $8.0\text{E}+5 \text{ \#}$
205 cm^{-3} in Zone 2), which is statistically significantly higher compared to both background of the
206 same day, but more importantly, all the scenarios involving IQOS usage. Such levels and
207 variation is a result of the local particle sources, among most important were the operation of
208 artificial fog machine as well as fugitive emissions from the adjacent smoking room, as
209 confirmed further by the concentrations of CC markers. The particle levels were generally high,
210 comparable to those registered in a chamber during the use of CC (Meišutovič-Akhtarjeva et al.,
211 2019), and higher in German hospitality environments involving smoking of tobacco (Bolte et al.,
212 2008).

213 A general trend comparing results of PNC increase from both measurement campaigns was as
214 follows: background at $1\text{E}+4 \text{ \# cm}^{-3}$, 10 IQOS at $5\text{E}+4 \text{ \# cm}^{-3}$, 30 IQOS at $1\text{E}+5 \text{ \# cm}^{-3}$ and club
215 in operation at $1\text{E}+6 \text{ \# cm}^{-3} - 1\text{E}+7 \text{ \# cm}^{-3}$. This represents an increase by an order of magnitude
216 in each subsequent scenario.

217

218 ***Particle mass concentrations***

219 The particle mass concentration (PMC) as calculated from real-time ELPI+ measurements was
220 grouped based on the measurement scenarios (Fig. 2). Generally, $\text{PM}_{2.5}$ and PM_{10} displayed
221 similar variations between the measurement scenarios, except that PM_{10} indicated longer tails of
222 distribution (represented by both 5th and 95th percentile whiskers), suggesting that there was
223 higher variation of concentration. This is due to the fact that PM_{10} is more influenced by
224 resuspended dust, associated with human activities, while $\text{PM}_{2.5}$ is associated with primary and

225 secondary particles, resulting from thermal aerosol release (such as fog machines, cigarette
226 smoking, or exhalation of cigarette aerosol).

227 The scenarios from background to using of 30 IQOS devices did not result in significant increase
228 in PMC, with the median $PM_{2.5}$ ranging from $2.7 \mu\text{g m}^{-3}$ at Zone 1 and $2.8 \mu\text{g m}^{-3}$ at Zone 2 in
229 case of background to $11.4 \mu\text{g m}^{-3}$ (Zone 1) and $12.3 \mu\text{g m}^{-3}$ (Zone 2) in case of 30 IQOS. Such
230 comparatively low impact of IQOS usage to indoor PMC may have been resulted due to the fact
231 that the majority of particles was located in sub-micrometer size range (as discussed in the
232 following subchapter) thus not carrying much mass. Another issue may be associated with the
233 particle mass loss during the transport in sampling lines, since particles may have partially
234 evaporated.

235 The operating club resulted in significantly higher concentrations compared to 30 IQOS users.
236 Preparations to the club opening already resulted higher $PM_{2.5}$ concentrations, this time having
237 spatial variations between Zone 1 ($59.0 \mu\text{g m}^{-3}$) and Zone 2 ($35.0 \mu\text{g m}^{-3}$). This may be associated
238 with more intensive staff moving and re-suspending particles from surfaces, as well as potential
239 secondary tobacco aerosol brought by staff from smoking room.

240 Further on, club in operation resulted almost several orders of magnitude higher median $PM_{2.5}$
241 ($3715 \mu\text{g m}^{-3}$ at Zone 1 and $1770 \mu\text{g m}^{-3}$ at Zone 2) and PM_{10} ($4590 \mu\text{g m}^{-3}$ at Zone 1 and 2470
242 $\mu\text{g m}^{-3}$ at Zone 2) concentrations.

243 These levels are significantly higher than $PM_{2.5}$ concentration in hospitality venues ($36\text{-}869 \mu\text{g}$
244 m^{-3}) as reviewed by Fromme et al. (2012) in smoking areas of Californian casinos during the
245 periods of the highest occupancy (median $44\text{-}110 \mu\text{g m}^{-3}$) (Klepeis et al., 2012) or German
246 discotheques (median $599.2 \mu\text{g m}^{-3}$) during the principal business hours before implementation of
247 partial smoking ban (Gleich et al., 2011) and are closer to the mean $PM_{2.5}$ and PM_{10}
248 concentrations ($619.1 \mu\text{g m}^{-3}$ and $1156.6 \mu\text{g m}^{-3}$) during rush working hours (17–21) at the

249 combined water-pipe/CC smoking cafes in Tehran (Heydari et al., 2019). It must be noted that
250 such differences in case of our measurements possibly result from the deriving of mass
251 concentration from the real-time PNC measurements, as opposed to the filter based collection. In
252 the latter method, filters are conditioned before the gravimetric analysis, thus losing significant
253 portion of the mass in case of volatile aerosol, which seems to be the case in our measurements,
254 as indicated by the PSD analysis in the following subchapter.

255

256 *Particle size distributions*

257 The PSDs based on PNC indicated that the major part of particles were of sub-micrometer size
258 range, and more importantly, sub 100 nm range (Fig. 3(A)). It may be expected that aerosol in
259 premises with thermal sources would have a mode at 80-100 nm range, indicating accumulating
260 particles (Bolte et al., 2008). However, the ultrafine mode indicates that major portion of particles
261 were at nucleation size range for all measurement scenarios. We have observed such phenomena
262 in chamber for the exhaled THS aerosol and attributed such distribution to the rapidly
263 evaporating volatile matter of the exhaled aerosol, which may reach 95% of total mass. Moreover,
264 the measured RH levels during IQOS experiment averaged at 36-43%, which has been indicated
265 as favourable range for the particle evaporation (Meišutovič-Akhtarjeva et al., 2019). Club in
266 operation yielded similar PSD, possibly reflecting the presence of highly volatile aerosol from
267 fog machine, while RH stayed relatively low (29-35%).

268 The PSD based on the mass concentration (Fig. 3(B)) indicated the mode at 200-300 nm during
269 the use of IQOS and at $\sim 1 \mu\text{m}$ during club in operation. The latter is associated with the fact that
270 particle number concentration was significantly higher in that range, compared to IQOS. The
271 PSD in the range of 2-10 μm had much uncertainty associated with particle losses during
272 sampling thus could not be estimated quantitatively with high accuracy, but expectedly it should

273 feature a mode in particle mass caused by the re-suspended dust from the moving personnel and
274 visitors.

275

276 ***Gaseous pollutants***

277 The club featured relatively high background concentrations of gaseous pollutants (Fig. 4 and
278 Table S1). The mean background concentration of formaldehyde ($37.4 \mu\text{g m}^{-3}$) and acetaldehyde
279 ($41.1 \mu\text{g m}^{-3}$) were higher than typically found in residential and public environments
280 (Kaunelienė et al., 2018). The mean background concentration of nicotine ($3.4 \mu\text{g m}^{-3}$) and 3-EP
281 ($1.7 \mu\text{g m}^{-3}$) were similar to those measured in non-smoking areas of Finnish nightclubs and
282 discos during peak hours (geometric mean of $2.9 \mu\text{g m}^{-3}$ and $0.8 \mu\text{g m}^{-3}$, respectively) (Johnsson
283 et al., 2006). Presumably, fugitive emissions from the smoking room as well as exhaled cigarette
284 smoke, emissions from clothing and hair from returning smokers to the main club area resulted in
285 substantial adsorption of tobacco combustion products and subsequent re-emission, so called
286 third hand smoke.

287 The impact of the usage of THS in non-operating club to the gaseous pollutant variation was
288 much lower if at all significant compared to particle concentrations (Fig. 4). The usage of 30
289 IQOS caused slight increase of concentrations as compared to 10 IQOS, however increased
290 significantly only for formaldehyde and acetaldehyde in Zone 1 ($p < 0.05$). Hypothetically, the
291 usage of IQOS in Zone 2 should have been resulted in higher concentrations in comparison to
292 Zone 1 due to restricted dispersion. This hypothesis has been partially confirmed for
293 formaldehyde, acetaldehyde and nicotine (10 IQOS), but the differences did not appear to be
294 statistically significant.

295 The club in operation resulted in either the similar levels of pollutants to 30 IQOS (formaldehyde,
296 nicotine in Zone 2) or higher pollutant levels (3-EP, statistically significantly acetaldehyde).

297 Although the club is a non-smoking environment, Zone 2 was most likely affected by fugitive
298 emissions of tobacco smoke from the adjacent smoking room. This is evidenced by increased
299 concentrations of tobacco-specific markers – nicotine and 3-EP. Elevated levels of acetaldehyde
300 may be attributed to the consumption of alcohol-containing beverages, as acetaldehyde is a
301 product of ethanol metabolism and is present in the exhaled breath (Bagnardi et al., 2001).
302 The observed levels of nicotine during the use of 30 IQOS and club in operation were at same
303 range compared to the concentrations in the smoking areas of Finnish nightclubs and discos
304 during peak hours (geometric mean $11.0 \mu\text{g m}^{-3}$ and $2.4 \mu\text{g m}^{-3}$, respectively, Johnsson et al.
305 (2006), Las Vegas casinos where smoking is still not banned ($6.7 \mu\text{g m}^{-3}$, Achutan et al. (2011)),
306 New York City hookah bars during hours of the highest occupancy ($4.2 \mu\text{g m}^{-3}$, Zhou et al.
307 (2015)).
308 Penetration of pollutants from smoking areas to non-smoking areas has been reported in multiple
309 earlier research campaigns. South Australian pubs, clubs, and cafes having separately enclosed
310 smoking areas had lower mean levels of nicotine $4.4 \mu\text{g m}^{-3}$ compared to the unenclosed ones
311 ($7.5 \mu\text{g m}^{-3}$), while air of smoking areas contained nicotine concentration of $15 \mu\text{g m}^{-3}$ measured
312 over a period of 2 – 4 h during a normal to busy dinner or lunch sessions (Cenko et al., 2004).
313 Pollutant concentrations were significantly higher in Spanish hospitality venues with outdoor
314 smokers close to the entrance than in those without outdoor smokers measured during the same
315 range of hours (morning/afternoon) (López et al., 2013).

316

317 **CONCLUSIONS**

318

319 The usage of THS in the night club hospitality environment resulted in increasing particle number
320 and mass concentrations from the background by the order of magnitude, but were an order of
321 magnitude lower compared to club in operation, having fugitive emissions from adjacent

322 smoking room and fogging machines. Particles generated by THS usage as well as generated in
323 operating club demonstrated high volatility as indicated by significant nucleation mode in particle
324 distribution. At the same time, the use of IQOS did not significantly affect the concentration of
325 the majority of gaseous pollutants in the club having high background concentration values of
326 tobacco smoke related pollutants.

327 While the potential observed impact of intensive IQOS usage (such as 30 simultaneous users) in
328 indoor hospitality environments may be associated with the deteriorating indoor air quality and
329 increase exposure to nicotine and acetaldehyde, as well as particles, care should be taken that
330 despite the partial ban smoking of traditional cigarettes may still be the most important factor
331 affecting the exposure of visitors in the night club hospitality environment.

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339

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Figure Captions

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446 **Figure 1.** Variations of PNC ($\# \text{ cm}^{-3}$) during controlled use of simultaneous 10 and 30 IQOS devices (above) and
447 during uncontrolled operation of the club (below).

448

449 **Figure 2.** Particle mass ($\text{PM}_{2.5}$ and PM_{10}) and number (PN_{10}) concentrations in Zone 1 and Zone 2 under controlled
450 conditions (10 IQOS and 30 IQOS) and club in operation (Club OP). Backgr.+10p. – background with 10 humans
451 present but without THS use; Backgr.+30p. - background with 30 humans present but without THS use.

452

453 **Figure 3.** Particle size distribution of particle number concentration (A) and particle mass concentration (B) during
454 the use of THS under controlled conditions (10 IQOS and 30 IQOS) and club in operation.

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456 **Figure 4.** Concentration of gaseous pollutants under different scenarios: 10 volunteers using IQOS (10 IQOS), 30
457 volunteers using IQOS (30 IQOS) and club in operation (CLUB OP); Z1 - Zone 1, Z2 - Zone 2. Background
458 subtracted values are indicated as dots.

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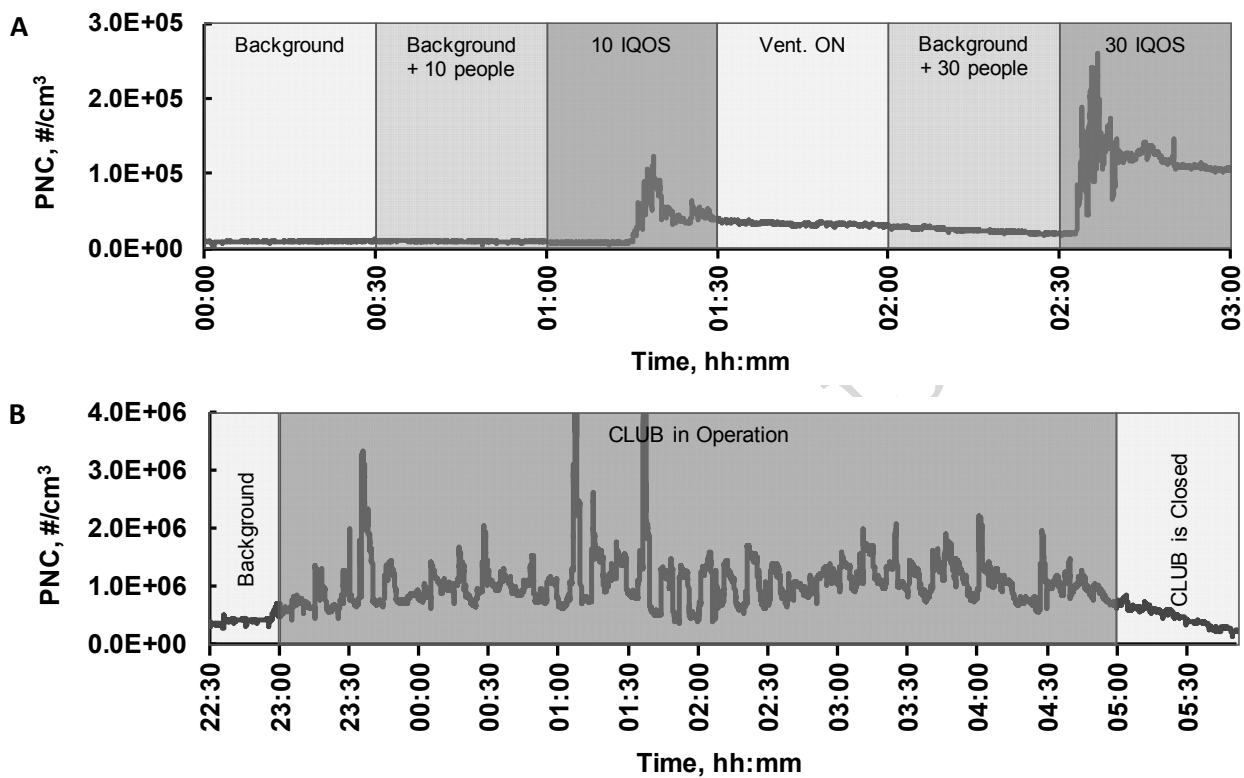
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Fig.1.

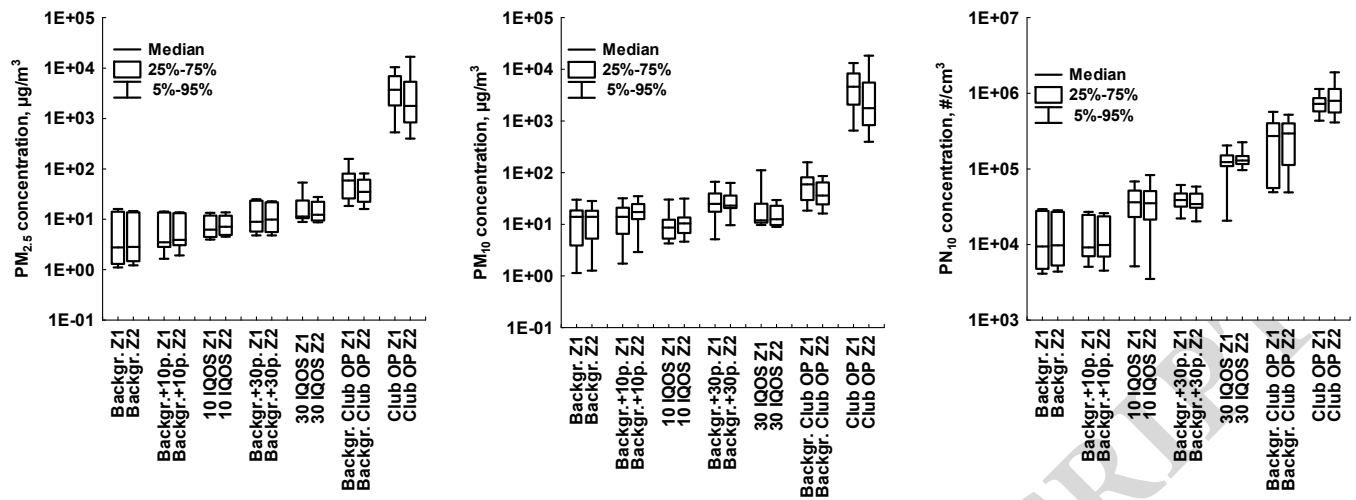
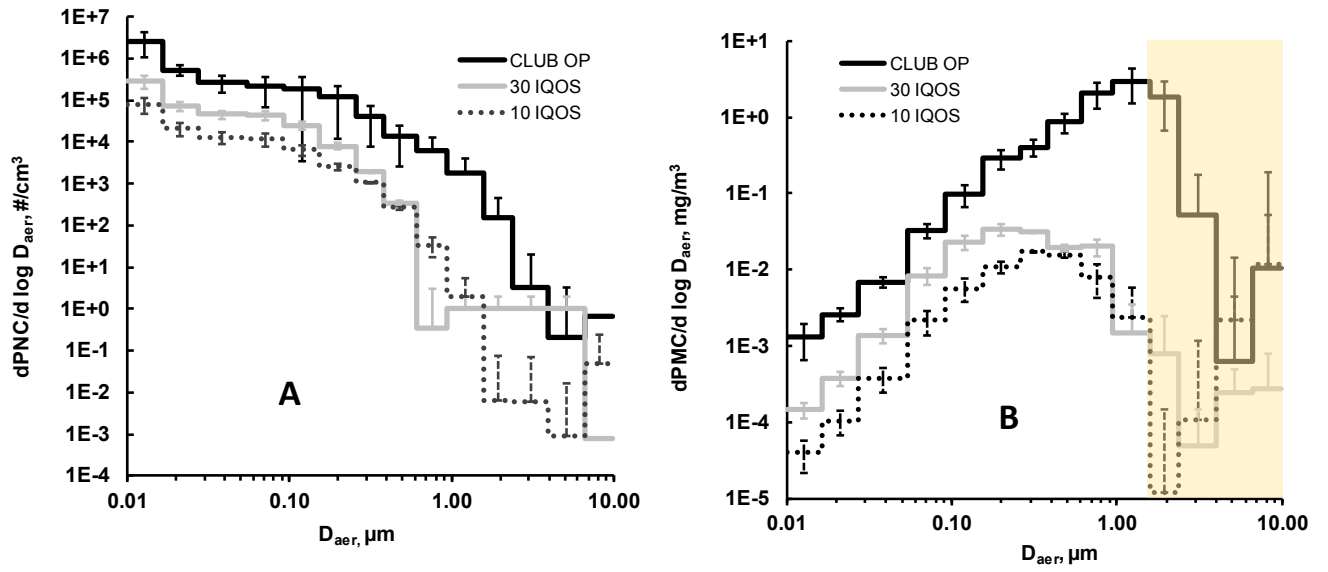


Fig.2.

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Fig.3.

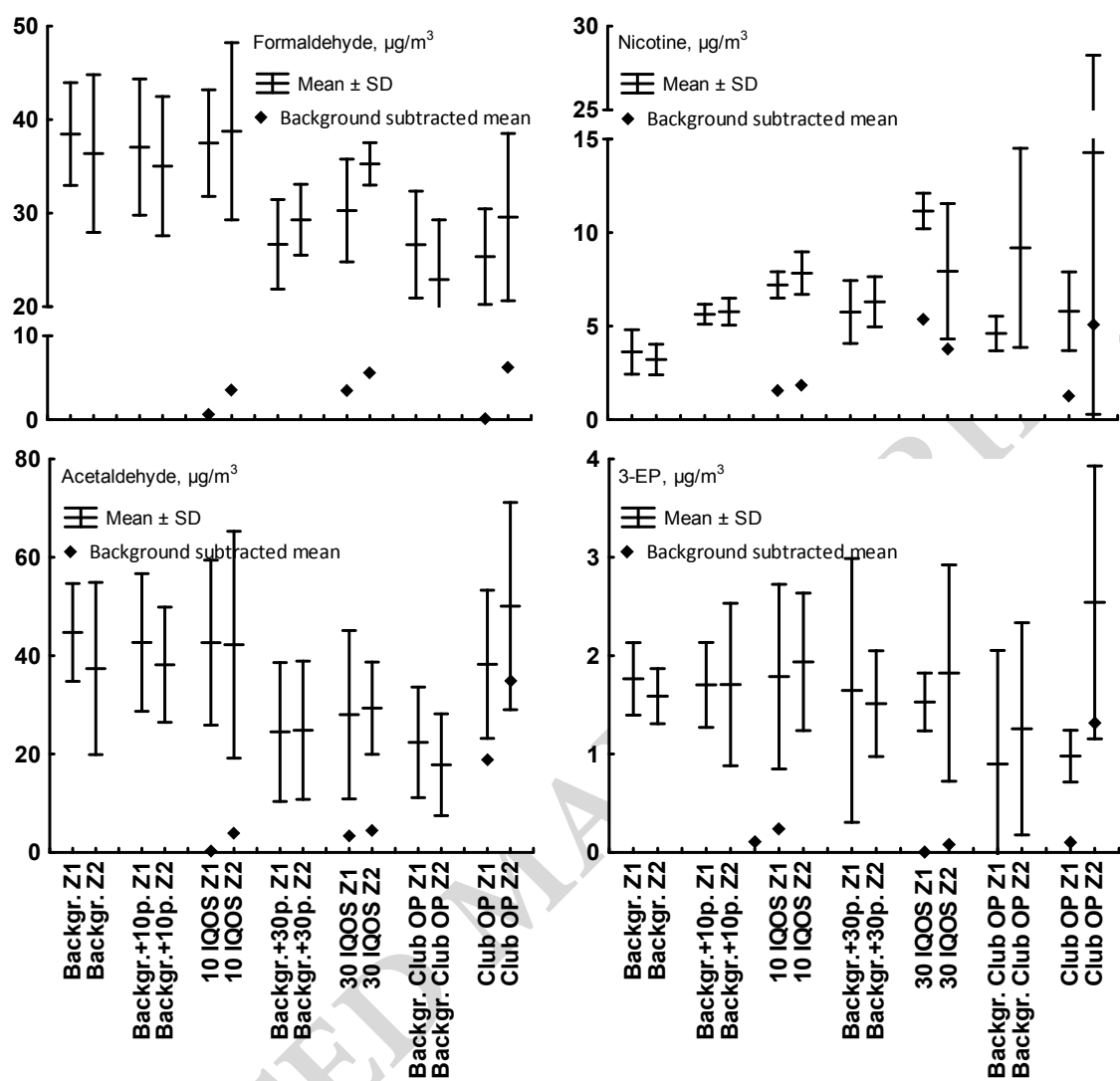


Fig.4.

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