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2 **A Cost-effective, Miniature Electrical Ultrafine Particle Sizer (mini-**
3 **eUPS) for Ultrafine Particle (UFP) Monitoring Network**
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Abstract

A cost-effective, miniature electrical ultrafine particle sizer (mini- eUPS) has been developed for the future UFP (ultrafine particle) monitoring network in cities with high traffic density and in communities close to either freeways, airports or stationary combustion sources. The mini- eUPS primarily consists of a mini-plate unipolar particle charger, mini-plate differential mobility classifier and a mini- particle Faraday cage with a sensitive electrometer. A custom-made circuit board was designed for the mini-eUPS operation. Additional to the primary function of measuring the UFP size distribution, the mini- eUPS records the temperature, relative humidity, pressure, altitude, timing and location for each measured data set. The operational status of mini-eUPS was also registered for the data quality assurance. The built-in features of self-diagnosis, malfunction reporting and wireless networking make mini- eUPS suitable for the future UFP monitoring network. The prototype mini- eUPS, having 6.5" (L) × 5.0" (W) × 4" (H) in the overall package ~1.0 kg in the total weight, measures the particles in sizes ranging from 5 to 200 nm. The performance of mini- eUPS was evaluated by comparing its measured data with those measured by a scanning mobility particle sizer (SMPS) for lab-generated particles in both unimodal and bimodal distributions, and UFPs emitted from a low-cost 3D printer and diesel engine. Reasonable agreement between mini- eUPS and SMPS data was obtained in the comparison.

Keywords: Miniature particle sizer, ultrafine particles, cost-effective sensor

57 **1. Introduction**

58 The concern of adverse impact of ultrafine particles (UFPs, defined as particles in the
59 sizes less than 100 nm) on the environment and health has been increasing in recent years as the
60 number of scientific evidences published in the literature increased. From the toxicological and
61 health perspective, UFPs, due to their high surface-to-volume ratio and high number
62 concentration, often possess high bio-availability and toxicity (Nel et al., 2006; Heinlaan et al.,
63 2008; Park et al., 2011) Peer-reviewed literature reports that UFPs could easily enter the human
64 lung and deposit in the alveolar region, even entering in the blood stream and transporting to
65 vital organs (Paur et al., 2001; Takenaka et al., 2001; Kreyling et al., 2002; Oberdörster et al.,
66 2004). Epidemiologic studies have also shown that UFPs are particularly relevant to pulmonary
67 diseases, cancer and mortality because of their high diffusion coefficient and greater
68 accumulation ability (Delfino et al., 2005; Samet et al., 2007; Hoek et al., 2010; Steward et al.,
69 2010; Ostro et al., 2015; Aguilera et al., 2016; Du et al., 2016; Heusinkveld et al., 2016; Li et al.,
70 2017). The increased asthma prevalence has been often found to occur in the areas having high
71 UFP level in the ambient or high motor vehicle traffic density and residence communities in
72 close proximity to freeways (Samet et al., 2000; Holguin, 2008; Salam et al., 2008; Patel and
73 Miller, 2009; Mazaheri et al., 2014; Kim et al., 2016). In spite of the published health data, a
74 causal relationship between the short-term UFP exposure and cardiovascular effects was
75 suggestive, but not sufficient, by EPA 2009 PM Integrated Science Assessment, ISA (US EPA
76 Report 2009; Baldauf et al., 2016). Limited data on the spatial/temporal variability in UFP
77 concentration and on the UFP composition, and incomplete information on the spatial/temporal
78 evolution of UFP size distribution are some of the limitations on the published health studies for
79 the above causality determination.

80 UFPs are either derived directly from combustion and vehicle emission or formed when
81 organic compounds with low vapor pressure spontaneously nucleate/condense on other small
82 particles (Kulmala et al., 2004). UFPs are major sources for fine particles produced through the
83 agglomeration and/or vapor condensation. In addition to their omni-presence existence in the
84 ambient, UFPs of various chemical compositions have been manufactured for industrial
85 applications. Example applications of industrial UFPs are silver nanoparticles in apparel for their
86 antibacterial and anti-odor properties, the use of titanium dioxide and zinc oxide nanoparticles in
87 cosmetics to effectively block the UV rays in the sunlight, and the inclusion of carbon
88 nanoparticles in rubbers to lengthen the lifetime of vehicle tires (Lee et al., 2007; Vassiliou et al.,
89 2008; Somasundaran et al., 2010; Madani et al., 2011; Smijs and Pavel, 2011).

90 However, the monitoring of ultrafine particles is challenging because of the rapid
91 variation of size and concentration of UFPs both in time and in the distance from their sources,
92 especially for the communities in close proximity to freeways/airports or in cities where a high
93 density of motor vehicles are present (Zhu et al., 2002; Karner et al., 2010; Padró-Martínez et al.,
94 2012; Patton et al., 2014; Manigrasso et al., 2017; Hudda et al., 2018). UFP monitoring networks
95 in the resolution of 10-20 meters are required to collect the data needed for epidemiologic studies
96 (Zhu et al., 2002). Establishing a UFP monitoring network thus requires multiple monitors as the
97 network nodes. Scientific instruments for the UFP characterization have been developed and
98 applied in various aerosol particle studies for decades. The review of instruments capable of
99 measuring UFPs could be found in the literature (Chen and Pui, 2008). Instruments for the
100 integral measurement of UFPs (i.e. total number, surface area and mass concentrations of UFPs),
101 are available in market. Examples of integral instruments are condensation particle counters
102 (CPCs), for total number concentration measurement; nanoparticle surface area monitors

103 (NSAMs), for sensing total surface area concentration of particles deposited in a human lung;
104 and electrical aerosol detectors (EADs), for measuring the total mass concentration of UFPs. In
105 addition, particle-electrical-mobility-based instruments (e.g. Electrical Aerosol Analyzer, EAA,
106 and Differential Mobility Analyzer, DMA) have been widely applied in measuring the size
107 distributions of UFPs. Electrical-mobility-based instruments in the portable version (listed in the
108 Table 1) have also been made available in recent years. The above UFP sizers are mainly
109 designed for laboratory use and operated alone. They are typically expensive in list price and
110 bulky/heavy in final packages, limiting the application in the UFP monitoring network. Cost-
111 effective UFP sizers in much compact packages are highly in demand for setting up a UFP
112 monitoring network.

113 In this article, we report our effort in the development of a cost-effective, miniature
114 electrical ultrafine particle sizer (mini- eUPS). Electrical-mobility-based techniques are the
115 backbone of this mini- eUPS because of their capability of measuring the size distribution of
116 UFPs. A typical electrical-mobility-based particle sizer consists of a particle charger which
117 imposes a well-defined charge distribution on particles to be measured, an electrical-mobility
118 classifier which either sizes particles or changes size distribution of particles based on their
119 electrical mobility, and a particle concentration detector which measures the concentration of
120 particles being sized/changed. Typical particle concentration detectors are either CPCs or aerosol
121 electrometers. In the following sections, we describe the configuration of mini- eUPS and its
122 core components. The evaluation of core components and additional features of the mini- eUPS
123 are also discussed. Examples are at last given to demonstrate the overall performance of the
124 mini- eUPS.

125 **2. Materials and Methods**

126 **2.1 Design of mini- eUPS and its Core Components**

127 Shown in Figure 1 is the schematics of miniature electrical ultrafine particle sizer (mini-
128 eUPS). The mini- eUPS consists of three core components: a unipolar mini-plate particle charger,
129 a mini-plate differential mobility classifier, and a mini- particle Faraday cage with a sensitive
130 electrometer. Ambient particles are sampled into the mini- eUPS via a miniature cyclone, to
131 remove particles with the sizes larger than 200 nm. Prior to entering the mini-plate charger, the
132 temperature and relative humidity of the particle stream are measured. The used mini- cyclone
133 was designed and calibrated according to the works of Hsiao et al (2009) and Liu et al (2015).

134 After taking the temperature and relative humidity measurement, the particle stream is
135 directed to a mini-plate unipolar charger (shown in Fig. 2a) to electrically charged particles (Liu,
136 2015; Chen et al., 2018). The charger is the modified version of the mini-plate particle charger
137 studied in the dissertation work of Liu (2015), in which the detail optimization and performance
138 evaluation of the mini-plate charger can be found. The particle charger included in the mini-
139 eUPS is constructed by sandwiching one metal plate and one metal block. The particle flow
140 channel (in the rectangular cross-section and with the tapered expansion inlet and contraction
141 outlet) is embedded in the metal block with the inlet tube. The metal plate with the outlet tube is
142 used as the channel cover for the flow channel. The ion generation chamber is located near the
143 flow channel exit and built in the metal block. A perforated plate is used to partition the ion
144 chamber from the flow channel. The particle-charging zone is defined in the channel space
145 adjacent to the perforated plate. In the ion generation chamber, a Tungsten wire of 50 μm in
146 diameter is installed in the width direction to produce unipolar ions via positive DC-corona
147 discharging. Once produced, a small portion of unipolar ions is diffused into the particle-
148 charging zone. Particles enter the charger from the left-top inlet tube, pass through the particle

149 charging zone where they acquire electrical charges by the random collision with unipolar ions,
150 and exit the charger via the right-bottom outlet tube. The corona discharge is operated at 2.0 μA
151 in order to achieve the optimal performance of the charger. For the simplicity, no ion-driving
152 voltage is applied in the operation of this charger. Note that an additional feature of sheath flow
153 is also designed in the charger, offering a clean flow to reduce the possible contamination of
154 corona wire.

155 A mini-plate differential mobility classifier (mini-plate DMC; shown in Fig. 2b) is placed
156 at the exit of the mini-plate particle charger to size particles according to their electrical mobility
157 (Liu and Chen 2016). The mini-plate DMC is in general constructed by two metal plates installed
158 in parallel and separated by the spacing of 1.6 mm (i.e., 1/16"): the height of particle
159 classification zone). An aerosol flow contraction/expansion channel (for transporting either
160 polydisperse or classified particles) is designed in each metal plate to minimize the electrostatic
161 loss of charged particles. Polydisperse charged particles flow in the mini-plate DMC from the
162 top tube, through a flow expansion channel, to the particle classification zone via the entrance
163 slit. The dimensions of aerosol entrance and exit slits are 28.6 mm (i.e., 1 1/8" in length), which
164 is ~75% of the full width of the classification zone. The reason for the reduced slit length
165 (compared with the full channel width) is to minimize the wall effect on the DMC performance.
166 The distance between the aerosol entrance and exit slits, i.e., the particle classification length, is
167 52.4 mm (i.e., 2 1/16"). Classified particles are extracted out the classification zone from the exit
168 slit, moved through a flow contraction channel, and exited the DMC from the bottom tube.
169 Particle-free sheath gas is directed into the particle classification zone via the left inlet and a flow
170 laminarizer. The excess flow exits the classification zone from the outlet located on the other end
171 of the DMC. A high DC voltage is applied to the top plate while the bottom one is on the

172 electrical ground, establishing a uniform electrical field in the particle classification zone. For
173 safety, the metal plates are electrically insulated. The overall size of this prototype mini-plate
174 DMC is comparable to the size of an iPhone 6. Instead of the flowrate ratio of 1:10, the mini-
175 plate DMC is in default operated at the aerosol and sheath flowrates of 0.3 and 1.5 lpm (i.e., the
176 flowrate ratio of 1:5) in the mini- eUPS.

177 Instead of condensation particle counters (CPCs), a mini- particle Faraday cage integrated
178 with a sensitive electrometer is applied in the mini- eUPS for measuring the concentration of
179 classified particles. The selection of an aerosol Faraday cage over a CPC is based on the
180 consideration of energy consumption, working fluid handling, miniaturization and the cost of
181 CPCs. Heating and cooling of the working fluid are required in a CPC to enlarge the sizes of
182 classified particles for optically counting. These thermal processes consume much energy. Extra
183 accessory is also required to accomplish the tasks of feeding, storage and refilling for the
184 working fluid in a CPC. The extra accessory will increase the cost and package size of the mini-
185 eUPS. Notice that the trade-off for using a particle Faraday cage over a CPC is on the low
186 detection limit of particle concentration. A CPC would measure the particle in the number
187 concentration down to 1 \#/cm^{-3} . Particles in the number concentration more than 300 \#/cm^{-3} for
188 each size bin (when operated at the flowrate of 0.6 lpm) is required for sensing via a particle
189 Faraday cage, because of the lower detection limit of 0.5 fA of the current electrometer.
190 Although the low current limit will be reduced to 0.2 fA in the later version of electrometer to be
191 incorporated in the mini- eUPS, classified particles in the concentration more than 100 \#/cm^{-3} for
192 each size bin are required for concentration sensing (assuming unit charge on individual
193 classified particles).

194 Fig. 2c shows the schematics of mini- particle Faraday cage used in mini- eUPS.
195 Charged particles are directed into the cage through the top inlet tube, and trapped in a small disk
196 filter located in the metal cup which is electrically isolated from the cage and grounded through a
197 wire. A sensitive electrometer connected to the wire is used in order to measure the induced
198 current once the charged particles are trapped. After passing through the disk filter, the carry
199 flow exits the cage from the outlet tube.

200 **2.2 Integration of mini-eUPS**

201 Two small high-voltage power supplies (EMCO C50 and Q50-5) are included in the
202 mini- eUPS for the operations of the unipolar mini-plate particle charger and mini-plate DMC.
203 For the charger operation, a feedback control loop is designed to keep the corona discharge
204 current constant (at the default setting of 2.0 μ A). One small air pump (Schwarzer Precision
205 FZ135) and a micro-bridge flowmeter (Honeywell AWM 43600) are used to sample and monitor
206 particle flows, respectively. For the mini-plate DMC operation, the 2nd small pump (Schwarzer
207 Precision FZ140) is applied as the driver to extract the excess flow for its port, pass it through
208 two HEPA-grade tube filters (located at the upstream and downstream of the pump), and inject it
209 into the sheath flow port. The 2nd micro-bridge flow meter (Honeywell flow meter-AWM 3300)
210 is used to monitor the circulating flow rate.

211 A custom-made circuit board and an A/D card were used to acquire the readings from
212 high-voltage supplies, micro-bridge flowmeters, and temperature and humidity sensors, and to
213 control high-voltage power supplies and air pumps. In addition, the wireless communication and
214 networking modules, timer, global positioning system (GPS) and pressure sensor are also
215 included in the same circuit board. The Raspberry Pi micro-processor with the 7" touch screen is

216 selected as the user interface, and for running the code for the mini- eUPS operation and data
217 analysis/display.

218 The prototype mini- eUPS is in default operated at the stepping voltage mode in order to
219 simplifying the coding. The scanning voltage operation mode will be later included in the code.
220 As the default setting, 20 size bins covering the sizes ranging from 5 to 200 nm are set in the
221 mini- eUPS. Because of the response time of electrometer varies at different current levels, the
222 measuring cycle is set at 60 seconds (i.e., 3 seconds for each size bin on average). The measuring
223 cycle could be reduced when measuring UFPs in high concentration. It is because of the fast
224 response time of electrometer at the high readout compared with that at the low signal.

225 The assembled mini- eUPS with all the accessory is packed in the commercially available
226 box having the size of 6.5" (L) \times 5.0" (W) \times 4" (H) and with the total weight of \sim 1.0 kg (without
227 the battery). Note that the package size could be further reduced if it was custom-made. Either a
228 lithium-ion battery or an AC-to-DC power adapter can power the mini- eUPS. Two USB ports
229 are also available for the communication between the mini- eUPS microprocessor and a laptop
230 computer. For the reference, the design specification of mini- eUPS is summarized in Table 2.

231 To retrieve the particle size distribution from measured raw data, a data inversion scheme
232 is required for mini- eUPS because of the presence of multiple charges on particles of the same
233 size after passing through the particle charger. Mathematical techniques have been proposed for
234 the data inversion (Kandlikar and Ramachandranst, 1999). The choice of a particular algorithm
235 for the size distribution retrieval is always a trade-off among the algorithmic complexity,
236 accuracy of reconstruction and available calculation power (Voutilainen et al., 2000). Under the
237 consideration of limited computational power and memory available, we selected the constrained
238 least-square method for the near *in-situ* screening of ultrafine particle size distributions measured.

239 The constrained least-square scheme was first developed by Philips (1962) and later introduced
240 into the aerosol measurement by Twomey (1965, 1975). For retrieving agreeable size distribution,
241 the constrained method, with a maximum error tolerance of 5% in the measurements was
242 reported by Rizzi et al. (1982) and Markowski (1987). Please note that the built-in data
243 inversion is only for the quality screening of measured data. The final size distribution of UFPs
244 could be obtained by offline data inversion schemes if the on-board data inversion was
245 unsatisfactory (Wolfenbarger and Seinfeld, 1990; Talukdar and Swihart, 2003; Dubey and
246 Dhaniyala, 2013).

247 **3. Result and Discussion**

248 **3.1. Core components of mini- eUPS:**

249 Prior to the mini- eUPS integration, the performances of core components were
250 individually evaluated. The calibration method and detail performance of mini-plate DMC used
251 in mini- eUPS could be found in the work of Liu and Chen (2016). The performance of the mini-
252 plate particle charger and the mini- particle Farage cage are given in this section.

253 The calibration of mini-plate particle charger was performed by the experimental setup
254 and method detailed in the work of Chen et al. (2018). Figure 3 shows the basic performance of
255 mini-plate unipolar particle charger, i.e., intrinsic and extrinsic charging efficiencies as a
256 function of particle size, when it was operated at the 2 μ A corona current and at the flow rate of
257 0.6 lpm. The intrinsic charging efficiency is defined as the percentage of particles receiving
258 electrical charges in a particle charger and the extrinsic charging efficiency is the percentage of
259 charged particles existing a charger. The difference between the extrinsic and extrinsic efficiency
260 is due to the particle loss in a charger. The charging efficiency of the charger increased as the

261 particle size became larger. The extrinsic charging efficiency of 10 nm particles was about 12%,
262 evidencing a good charging performance of the mini-plate charger. For particles in sizes larger
263 than 100 nm, the intrinsic charging efficiency reached the maximal value of 100% while the
264 extrinsic charging efficiency stays ~ 85%. Also, the ratio of extrinsic charging efficiency to
265 intrinsic charging efficiency increased with the increase of particle size, from 38% (for 20 nm
266 particles) to 78% (for 60 nm particles), indicating that smaller particles were more likely lost in
267 the charger. For the inversion scheme to retrieve the size distribution of UFPs, the charge
268 distributions of particles at various sizes are also required. The charge distributions of particles in
269 given sizes, which is required for the data inversion scheme applied in the mini- eUPS, were also
270 measured and shown in Figure 4. To re-produce measured charge distribution for particles in
271 different size a Gaussian distribution function was proposed to fit the measured charge
272 distribution data. It is found that the three parameters in the Gaussian distribution equation are
273 functions of average charge and particle size. By best fitting the experimental charge
274 distributions, a data fitted Gaussian distribution equation for calculating the extrinsic charge
275 distribution of particles can be expressed as:

$$276 \quad y = 8.3691D_p^{-0.655} \times \exp\left\{\frac{[x-(0.0281D_p - 0.1847)]^2}{(0.0039D_p + 0.5502)^2}\right\} \quad (1)$$

277 where D_p is particle size (unit: nm). As shown in Fig. 4, the general trend of the experimental
278 and fitted charge distribution agrees well.

279 The performance of the mini- particle Faraday cage was calibrated, and the data are
280 shown in Figure 5. DMA-classified particles of 80 nm in diameter were used in the calibration of
281 the mini- Faraday cage. The calibration was performed at both aerosol flow rates of 0.3 and 1.5
282 lpm. The current (I) measured by the aerosol electrometer can be expressed as

283
$$I = ne(NQ_a) \quad (2)$$

284 where I is the current measured by mini- particle Faraday cage, N is the number concentration of
285 single-charged particles, measured by a UCPC (TSI 3776); Q_a is the flow rate of particle stream,
286 n is the number of elementary charges on individual particles ($n = 1$ in this case), and e is the
287 elementary charge, 1.6×10^{-19} C). Accordingly, the correlation between the measured current I
288 and the eNQ_a should be a straight reference line with the slope of 1.0 (or the 45 degree in the
289 plot). The excellent linear relationship between I and eNQ_a is shown in Fig. 5 and the slopes of
290 two lines are close to one (since all the test particles carried single charge).

291 **3.2 Comparison of mini- eUPS and TSI SMPS**

292 Four examples are provided herein to demonstrate the overall performance of the
293 prototype mini- eUPS by comparing the mini- eUPS data with those measured by TSI SMPS
294 (Scanning mobility Particle Sizer, TSI Inc.). The TSI SMPS (scanning mobility particle
295 spectrometer) used in this comparison consisted of a particle neutralizer (i.e., either the Kr⁸⁵ or
296 soft X-ray neutralizer), a long DMA (Differential Mobility Analyzer; TSI model 3081) with the
297 DMA platform (TSI model 3080) and UCPC (Ultrafine Condensation Particle Counter; TSI
298 model 3776). The SMPS was operated at the aerosol-to-sheath flow rate ratio of 1:10. All the
299 raw SMPS data were measured in the highest resolution offered by the SMPS. Since the mini-
300 eUPS has less sizing bins than TSI SMPS, the raw size bins of SMPS data were merged to the
301 number of size bins equivalent to mini- eUPS's for the comparison.

302 Sub-micrometer particles in both unimodal and bi-modal size distributions were
303 generated in the lab for this part of testing. The detail of experimental setup used to generate test
304 particles could be found in the Supplementary information (SI) (Figure S1). Particles in

305 unimodal size distribution were generated via the nebulization of NaCl solutions in 0.1%
306 concentration by a custom-made Collison atomizer. Once produced, the particle stream was
307 directed through a diffusion dryer with silica gel as the descendant to remove the water vapor
308 and a Po^{210} neutralizer to minimize the electrical charge level on resultant particles. Particles in
309 the bi-modal size distribution were generated by mixing the unimodal NaCl particles produced
310 above with DEHS (Di-Ethyl-Hexyl-Sebacat) oil particles produced by condensation aerosol
311 generator (TSI model 3475).

312 The comparison of size distributions measured by mini- eUPS and TSI SMPS for lab-
313 generated particles in unimodal and bimodal distributions is shown in Figure 6 (6a for unimodal
314 size distribution; 6b for the bi-modal size distribution). For each particle size distribution
315 measurement, at last five runs were taken and the average of measured distributions was reported.
316 The particle size distributions measured by mini- eUPS is in reasonable agreement with those
317 obtained by TSI SMPS, indicating that the assembled mini- eUPS is in the working condition.
318 The slight variation in each size bin of measured particle size distributions are primarily because
319 of the use of different particle concentration detectors and data reduction schemes used in mini-
320 eUPS and TSI SMPS

321 Both mini- eUPS and TSI SMPS were also applied to measure the diesel particles
322 generated from EPA Tier 2 diesel engine (Cummins Onan quiet diesel 8KW max Genset). Figure
323 S2 shows the facility and experimental setup for this measurement (in SI). Figure 7 shows the
324 size distributions of diesel particles measured by both instruments. More than three runs were
325 taken in this measurement and the average of measured distributions was reported. The
326 reasonable agreement between two measured particle size distributions was again obtained in
327 this measurement.

328 Last example of the performance comparison between mini-eUPS and SMPS is on the
329 measurement of particulate emission from a low-cost 3D printer (Lulzbot mini, Aleph Objects Inc.).
330 The schematic diagram and description of the setup for this part of the measurement is included
331 in the SI (Figure S3). A typical comparison of printer-emitted particle size distributions
332 measured by both mini- eUPS and TSI SMPS is shown in Figure 8. The data shown in Fig. 8
333 were the averages of five measured particle size distributions. The agreement between two
334 measured particle size distributions shall be considered reasonable although the SMPS-measured
335 concentration of particles in the range near the peak size was slightly higher than that measured
336 by mini- eUPS. It is again possibly due to the detector difference used in mini- eUPS and TSI
337 SMPS.

338 In the future, a small-scale wireless UFP monitoring network, based on multiple mini-
339 eUPS, will be set up for field testing of mini- eUPSs and their long-term network function.

340

341

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346 particles.

347

348 **COI statement**

349 Chen, one of the authors, holds the licensed IP, which is similar in name, but unrelated in
350 configuration, to this project.

351

ACCEPTED MANUSCRIPT

List of References

- 352
353
- 354 Aguilera, I.; Dratva, J.; Caviezel, S.; Burdet, L.; de Groot, E.; Ducret-Stich, R.E.; Eeftens, M.;
355 Keidel, D.; Meier, R.; Perez, L.; et al. (2016) Particulate matter and subclinical
356 atherosclerosis: Associations between different particle sizes and sources with carotid intima-
357 media thickness in the SAPALDIA study. *Environ. Health Perspect.*, 124(11):1700-1706.
- 358 Baldauf, Richard W., Robert B. Devlin, Peter Gehr, Robert Giannelli, Beth Hassett-Sipple,
359 Heejung Jung, Giorgio Martini, Joseph McDonald, Jason D. Sacks and Katherine Walker
360 (2016) Ultrafine Particle Metrics and Research Considerations: Review of the 2015 UFP
361 Workshop. *Int J Environ Res Public Health*, 13, 1054.
- 362 Chen, Da-Ren and D. Y. H. Pui (2008). Nanoparticle and Ultrafine Aerosol Measurement, in Air
363 Sampling Instruments for Evaluation of Atmospheric Contamination, 10th Ed., *ACGIH*.
- 364 Chen, Xiaotong, Qiaoling Liu, Jingkun Jiang and Da-Ren Chen (2018). Performance of Small
365 Plate and Tube Unipolar Particle Chargers at Low Corona Current, *Aerosol Air Qual.*
366 *Res.*, April 19, 2018. DOI: 10.4209/aaqr.2018.02.0060
- 367 Delfino, R.J., Sioutas, C., and Malik, S. (2005) Potential role of ultrafine particles in associations
368 between airborne particle mass and cardiovascular health. *Environ Health Perspect.*, 113:
369 934-946.
- 370 Du, Y.; Xu, X.; Chu, M.; Guo, Y.; Wang, J. (2016) Air particulate matter and cardiovascular
371 disease: The epidemiological, biomedical and clinical evidence. *J. Thorac. Dis.*, 8, E8–E19.
- 372 Dubey, P. and Dhaniyala, S. (2013). Improved Inversion of Scanning Electrical Mobility
373 Spectrometer Data Using a New Multiscale Expectation Maximization Algorithm. *Aerosol*
374 *Sci Technol.*, 47:69–80
- 375 Heinlaan, M., Ivask, A., Blinova, I., Dobourguier, H.C., and Kahru, A. (2008) Toxicity of
376 nanosized and bulk ZnO, CuO and TiO₂ to bacteria *Vibrio fischeri* and crustaceans *Daphnia*
377 *magna* and *Thamnocephalus platyurus*. *Chemosphere*, 71: 1308-1316
- 378 Heusinkveld, H.J.; Wahle, T.; Campbell, A.; Westerink, R.H.; Tran, L.; Johnston, H.; Stone, V.;
379 Cassee, F.R.; Schins, R.P. (2016) Neurodegenerative and neurological disorders by small
380 inhaled particles. *Neurotoxicology*, 56, 94–106.
- 381 Hoek, G.; Boogaard, H.; Knol, A.; de Hartog, J.; Slottje, P.; Ayres, J.G.; Borm, P.; Brunekreef,
382 B.; Donaldson, K.; Forastiere, F.; et al. (2010) Concentration response functions for ultrafine
383 particles and all-cause mortality and hospital admissions: Results of a European expert panel
384 elicitation. *Environ. Sci. Technol.*, 44, 476–482.
- 385 Holguin, F. (2008) Traffic, outdoor air pollution, and asthma. *Immunol Allergy Clin North Am*
386 2008, 28: 577-588, viii–ix.
- 387 Hsiao, T.-C., DA-REN CHEN and S. Y. Son. (2009) Development of Mini-cyclone as the Size-
388 selective Inlet for Miniature Particle Detectors, *J Aerosol Sci.*, 40:481-491.
- 389 Hudda, H., M.C. Simon, W. Zamore and J. L. Durant (2018). Aviation-Related Impacts on
390 Ultrafine Particle Number Concentrations Outside and Inside Residences near an Airport.
391 *Environ. Sci. Technol.*, 52, 1765-1772.
- 392 Kandlikar, M.; and Gurumurthy Ramachandranst (1999) Inverse Methods for Analyzing Aerosol
393 Spectrometer Measurements: A CRITICAL REVIEW. *J. Aerosol Sci.*, 30(4): 413-437.

394 Karner, A.A.; Eisinger, D.S.; Niemeier, D.A. (2010). Near-roadway air quality: Synthesizing the
395 findings from real-world data. *Environ. Sci. Technol.*, 44, 5334–5344.

396 Kim N. Dirks, Judith Y. T. Wang (2016). Amirul Khan and Christopher Rushton. Air Pollution
397 Exposure in Relation to the Commute to School: A Bradford UK Case Study. *Int J Environ*
398 *Res Public Health*,13, 1064

399 Kreyling, W.G., Semmler, M., Erbe, F., Mayer, P., Takenaka, S., and Schulz, H. (2002).
400 Translocation of ultrafine insoluble iridium particles from lung epithelium to extrapulmonary
401 organs is size dependent but very low. *J Toxicol Environ Health*, 166: 998-1004

402 Kulmala, M., Vehkamäki, H., Petaja, T., Dal Maso, M., Lauri, A., Kerminen, V.-M., Birmili, W.,
403 and McMurry P.(2004). Formation and growth rates of ultrafine atmospheric particles: A
404 review of observations, *J. Aerosol Sci.*, 35: 143-176.

405 Lee, Y.H., Park, H.K., Lee, Y.M., Kim, K., and Park, S.B.(2007). A practical procedure for
406 producing silver nanocoated fabric and its antibacterial evaluation for biomedical
407 applications. *Chem, Commun.*, 2959-2961

408 Li, Y., Kevin J. Lane, Laura Corlin, Allison P. Patton, John L. Durant, Mohan Thanikachalam,
409 Mark Woodin, Molin Wang and Doug Brugge (2017). Association of Long-Term Near-
410 Highway Exposure to Ultrafine Particles with Cardiovascular Diseases, Diabetes and
411 Hypertension. *Int J Environ Res Public Health*, 14, 461.

412 Liu, Di, Ta-Chih Hsiao, and Da-Ren Chen (2015). Performance Study of a Miniature Quadro-
413 inlet Cyclone, *J. Aerosol Sci.*, 2015, 90:161-168.

414 Liu, Qiaoling (2015). Ultrafine Particle Generation and Measurement, PhD dissertation, Virginia
415 Commonwealth University

416 Liu, Qiaoling and Da-Ren Chen (2016). Experimental Evaluation of Miniature Plate DMAs
417 (mini-plate DMAs) for Ultrafine Particle Sizing and Classification, *Aerosol Sci Technol.*,
418 50(3):297-307.

419 Madani, S.Y., Naderi, N., Dissanayake, O., Tan, A., and Seifalian, A.M. (2011). A new era of
420 cancer treatment: carbon nanotubes as drug delivery tools. *Int. J. Nanomedicine*, 2963-2979

421 Manigrasso, M., Claudio Natale, Matteo Vitali, Carmela Protano and Pasquale Avino (2017).
422 Pedestrians in Traffic Environments: Ultrafine Particle Respiratory Doses. *Int J Environ Res*
423 *Public Health*, 14, 288.

424 Markowski, G. R.(1987). Improving Twomey’s Algorithm for Inversion of Aerosol
425 Measurement Data. *Aerosol Sci. Technol.*, 7, 127-141.

426 Mazaheri, Mandana, Sam Clifford, Rohan Jayaratne, Megat Azman Megat Mokhtar, Fernanda
427 Fuoco, Giorgio Buonanno, and Lidia Morawska (2014). School Children’s Personal
428 Exposure to Ultrafine Particles in the Urban Environment. *Environ. Sci. Technol.*, 48, 113-
429 120.

430 Nel A, Xia T, Madler L, Li N. (2006). Toxic potential of materials at the nanolevel. *Science*,
431 311:622-627

432 Oberdörster, G., Sharp, Z., Atudorei, V., Elder, A., Gelein, R., Kreyling, W., and Cox, C. (2004).
433 Translocation of Inhaled Ultrafine Particles to the Brain. *Inhal. Toxicol.*, 16(6-7): 437-45.

434 Ostro, Bart, Jianlin Hu, Debbie Goldberg, Peggy Reynolds, Andrew Hertz, Leslie Bernstein, and
435 Michael J. Kleeman (2015). Associations of Mortality with Long-Term Exposures to Fine

436 and Ultrafine Particles, Species and Sources: Results from the California Teachers Study
437 Cohort. *Environ. Health Perspect.*, 123(6), 549-556.

438 Padró-Martínez, L.T.; Patton, A.P.; Trull, J.B.; Zamore, W.; Brugge, D.; Durant, J. L.(2012).
439 Mobile monitoring of particle number concentration and other traffic-related air pollutants in
440 a near-highway neighborhood over the course of a year. *Atmos. Environ.*, 2012, 61, 253–264.

441 Park, M.V., Neigh, A.M., Vermeulen, J.P., de la Fonteyne, L.J., Verharen, H.W., Briede, J.J.,
442 van Loveren, H., and de Jong, W. H. (2011). The effect of particle size on the cytotoxicity,
443 inflammation, developmental toxicity and genotoxicity of silver nanoparticles. *Biomaterials*,
444 32: 9810-9817

445 Patel, M.M. and Miller, R. L. (2009). Air pollution and childhood asthma: recent advances and
446 future directions. *Curr. Opin. Pediatr.*, 21: 235-242.

447 Patton, A.P.; Perkins, J.; Zamore, W.; Levy, J.I.; Brugge, D.; Durant, J.L. (2014). Spatial and
448 temporal differences in traffic-related air pollution in three urban neighborhoods near an
449 interstate highway. *Atmos. Environ.*, 99, 309–321.

450 Paur, H., Cassee, F.R., Teeguarden, J., Fissan, H., Diabate, S., Aufderheide, M., Kreyling, W.G.,
451 Hänninen, O., Kasper, G., Riediker, M., Rothen-Rutishauser, B., and Schmid, O. (2001). *J*
452 *Aerosol Sci.*, 42: 668-692

453 Philips, D. L. (1962). A technique for the numerical solution of certain integral equations of the
454 first kind, *J. Assoc. Comput. Mach.*, 9: 84–97.

455 Rizzi, R., Guzzi, R. and Legnani, R. (1982). Aerosol size spectra from spectral extinction data:
456 the use of a linear inversion method. *Appl. Opt.*, 21, 1578-1587.

457 Salam, M.T., Islam, T., and Gilliland, F.D. (2008). Recent evidence for adverse effects of
458 residential proximity to traffic sources on asthma. *Curr Opin Pulm Med.*, 14: 3-8.

459 Samet, J.M., Dominici, F., Curriero, F.C., Coursac, I., and Zeger, S.L. (2000). Fine particulate air
460 pollution and mortality in 20 US cities, 1987–1994. *N Engl J Med.*, 343: 1742–1749.

461 Samet J. M., D. Graff, J. Berntsen, A. J. Ghio, Yuh-Chin T. Huang and R. B. Devlin. (2007). A
462 Comparison of Studies on the Effects of Controlled Exposure to fine, Course and Ultrafine
463 Ambient Particulate Matter from a single Location. *Inhalation Toxicology*, 19(Suppl. 1):29-
464 32.

465 Smijs, T.G., and Pavel, S. (2011). Titanium dioxide and zinc oxide nanoparticles in sunscreens:
466 focus on their safety and effectiveness. *Nanotechnol Sci Appl.*, 2011, 4: 95-112.

467 Somasundaran, P., Fang, X., Ponnurangam, S., and Li, B.(2010). Nanoparticles: characteristics,
468 mechanisms and modulation of biotoxicity. *KONA Power Part J.*, 28: 38-48

469 Stewart, J., Chalupa, D.C., Devlin, R., Frasier, L.M., Huang, L.S., Little, E.L., Lee, S.M., Phipps,
470 R.P., Pietropaoli, A.P., Taubman, M.B., Utell, M.J., and Frampton, M.W.(2010). Vascular
471 effects of ultrafine particles in persons with type 2 Diabetes, *Environ. Health Perspect.*, 118:
472 1692-1698.

473 Takenaka, S., Karg, E., Roth, C., Schulz, H., Ziesenis, A., and Heinzamann U. (2001).
474 Pulmonary and systemic distribution of inhaled ultrafine silver particles in rats. *Environ.*
475 *Health Perspect.*, 2001, 109: 547-551

476 Talukdar, S. S., and Swihart, M. T.(2003). An Improved Data Inversion Program for Obtaining
477 Aerosol Size Distribution from Scanning Differential Mobility Analyzer Data. *Aerosol Sci.*
478 *Technol.*, 37:145-161.

479 Twomey, S. (1965). The Application of Numerical Filtering to the Solution of Integral Equations
480 Encountered in Indirect Sensing Measurements. *J FRANKLIN I*, 279: 95-109.

481 Twomey, S. (1975). Comparison of constrained linear inversion and an iterative nonlinear
482 algorithm applied to the indirect estimation of particle size distributions. *J. Comput.*
483 *Phys.*, 18(2):188-200

484 United States Environmental Protection Agency (EPA). Integrated Science Assessment for
485 Particulate Matter: Final Report; EPA/600/6-08/139F; National Center for Environmental
486 Assessment, Office of Research and Development: Research Triangle Park, NC, USA, 2009.

487 Vassiliou, A., Bikiaris, D., Chrissafis, K., Paraskevopoulos, K.M., Stavrev, S.Y., and Docoslis, A.
488 (2008). Nanocomposites of isotactic polypropylene with carbon nanoparticles exhibiting
489 enhanced stiffness, thermal stability and gas barrier properties. *Compos Sci Technol.*, 2008,
490 68: 933-943

491 Voutilainen, V., F. Stratmann and J. P. Kaipio. (2000). A Non-homogenous Regularization
492 Method for the Estimation of Narrow Aerosol Size Distribution. *J Aerosol Sci.*, 31(12):1433-
493 1445.

494 Wolfenbarger J. K. and Seinfeld, J. H. (1999). Inversion of Aerosol Size Distribution Data. *J.*
495 *Aerosol Sci.*, 21(2):227-247

496 Zhu, Y.; W. Hinds; S. Kim; S. Shen and C. Sioutas. (2002). Study of ultrafine particles near a
497 major highway with heavy-duty diesel traffic. *Atmos. Environ.*, 36:4323-4335.

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503 Table 1 Summary of compact/portable ultrafine particle sizers which are commercially
504 available. The prototype mini- eUPS is listed herein for the comparison only.

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506 Table 2 Summary of design specification of key components and prototype mini- eUPS.

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Table 1

Instrument	Dimensions	Weight
TSI Nanoscan 3910	45 x 23 x 39 (cm) (18"x9"x15.3")	~ 8 kg (w/o batteries); ~ 9 kg (with batteries)
Kanomax portable aerosol mobility spectrometer (PAMS)	23 x 23 x 15 (cm) (9"x 9"x 6")	4.5 kg
Naneum Nano-ID PMC500	30 x 33 x 26 (cm) 11.8"x 13"x 10.2"	6.25 kg
Grimm mini wide range aerosol spectrometer (WARS)	34 x 31x 12 (cm) (13.4"x 12.2"x 4.7")	7.6 kg
miniature electrical ultrafine particle sizer (mini- eUPS)*	16.5x12.7x10 (cm) (6.5"x 5.0"x 4")	~1 kg (w/o battery)

*it is listed for the comparison only (not commercially available).

Table 2**Specification of prototype mini- eUPS***

Final package	
Dimensions	6.5" (L) × 5.0" (W) × 4" (H) ^a
Weight	~ 1.0 kg (w/o battery)
mini-plate particle charger	
discharge current	0.2 μ A
voltage range	2-5 kV
flow rate	0.3-0.6 lpm
mini-plate differential mobility classifier	
particle flow rate	0.3-0.6 lpm
sheath flowrate	1.5-3.0 lpm
voltage range	0-5 kV
particle size range	5-200 nm at 1.5 lpm
mini- particle Faraday cage	
lower current limit	0.5 fA ^b
particle flow rate	0.3-0.6 lpm

* for the prototype mini- eUPS only (subject to change).

^a commercially available box

^b ~ 300 #/cm³ (singly-charged particles) at 0.6 lpm. The detection limit would be ~0.2 fA in the near future.

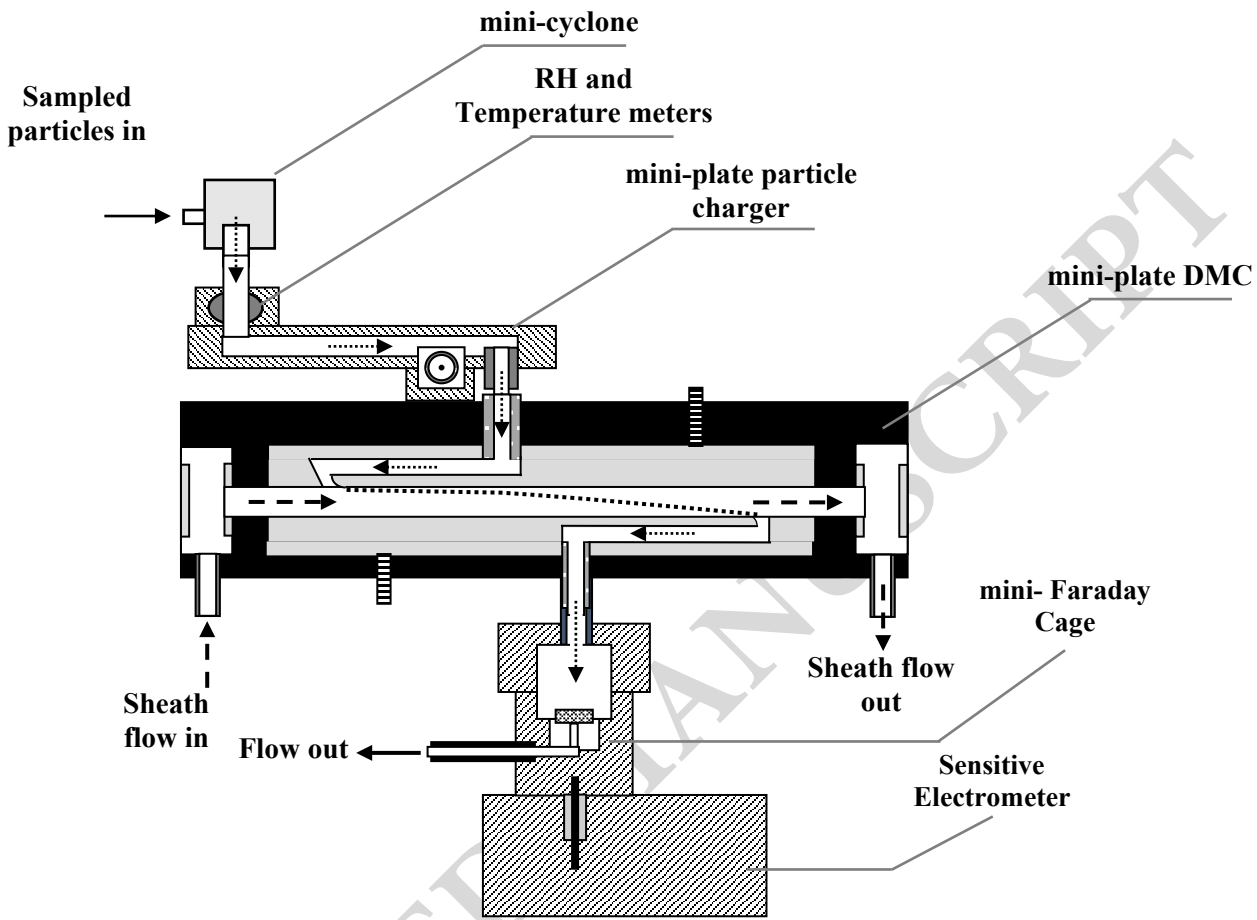
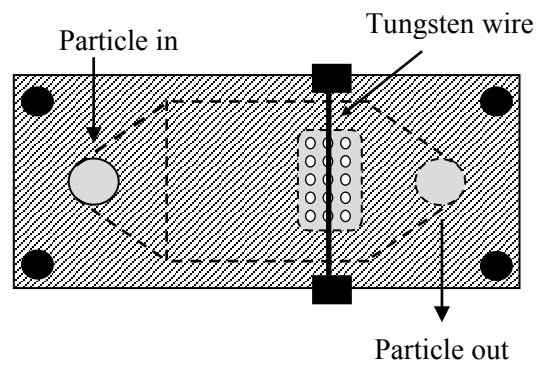
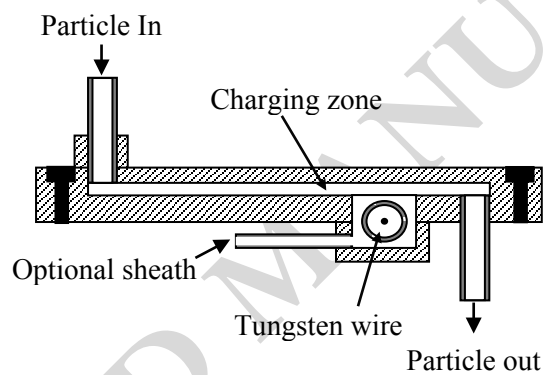


Figure 1



Top View



Side View

Figure 2a

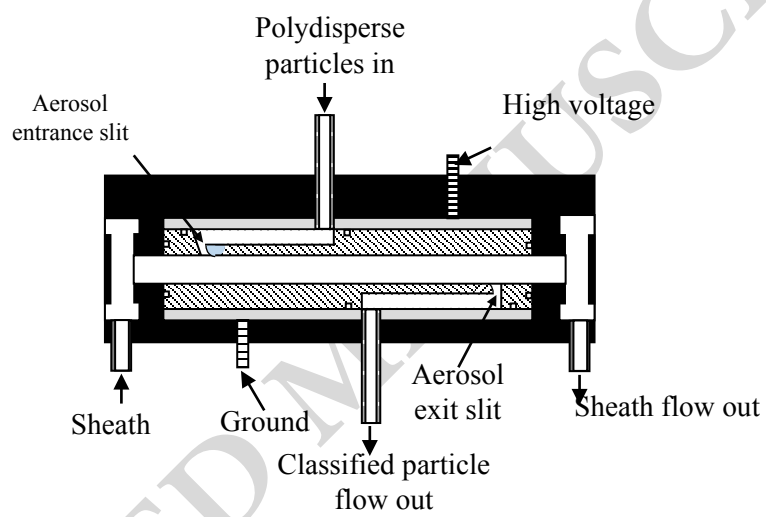
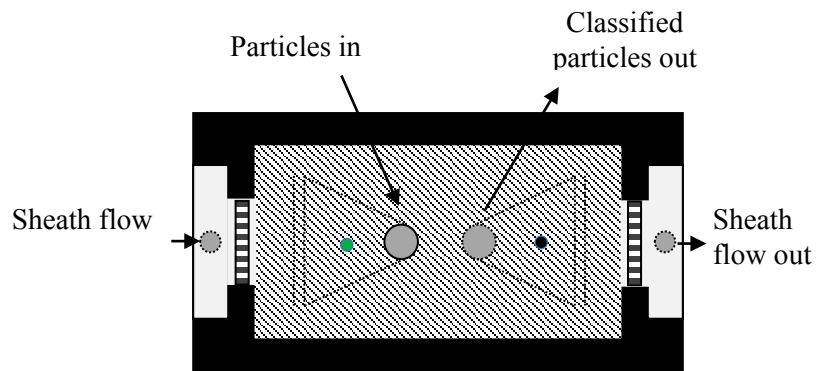


Figure 2b

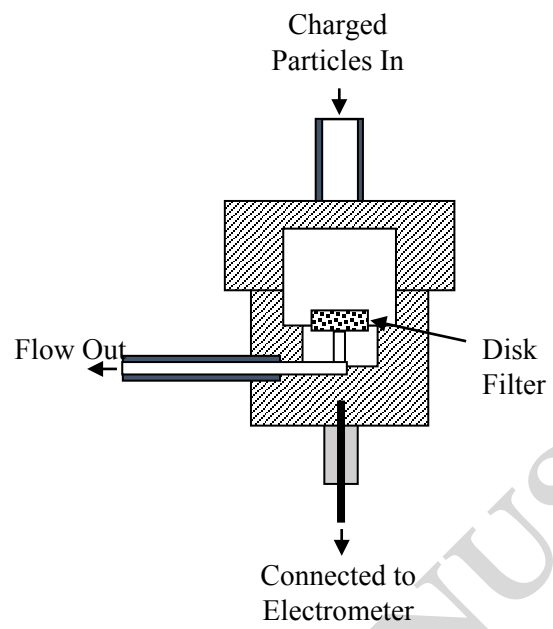


Figure 2c

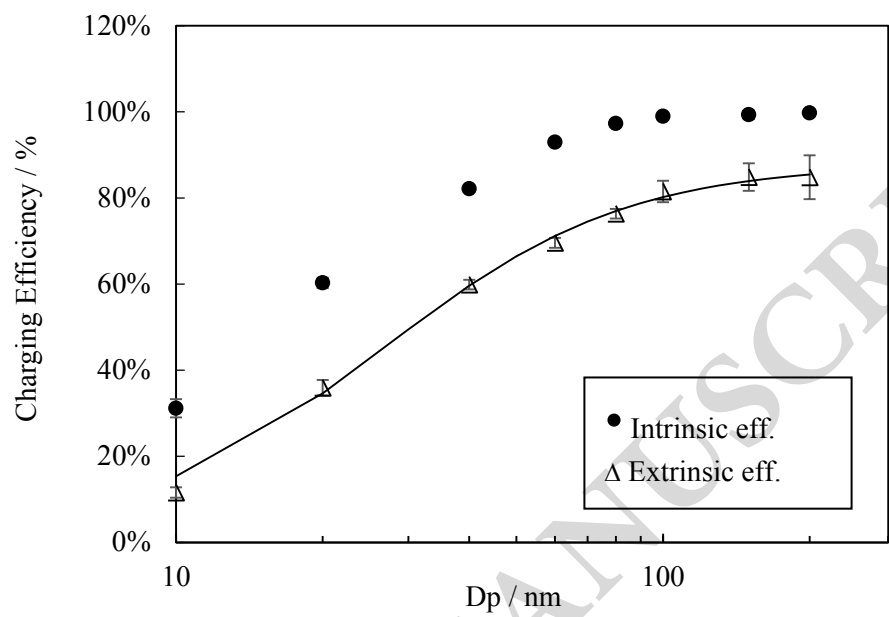
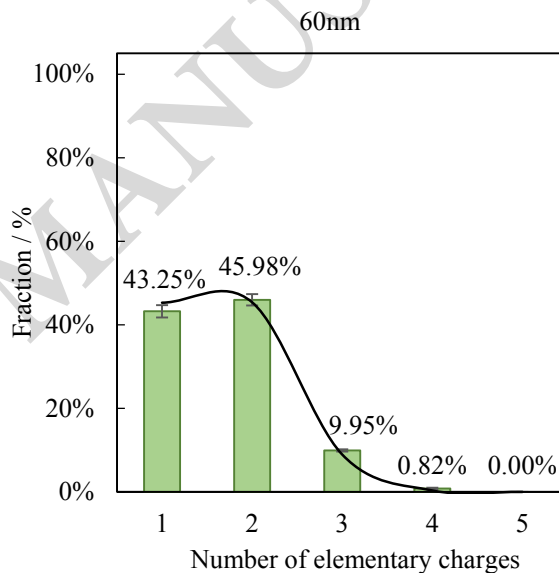
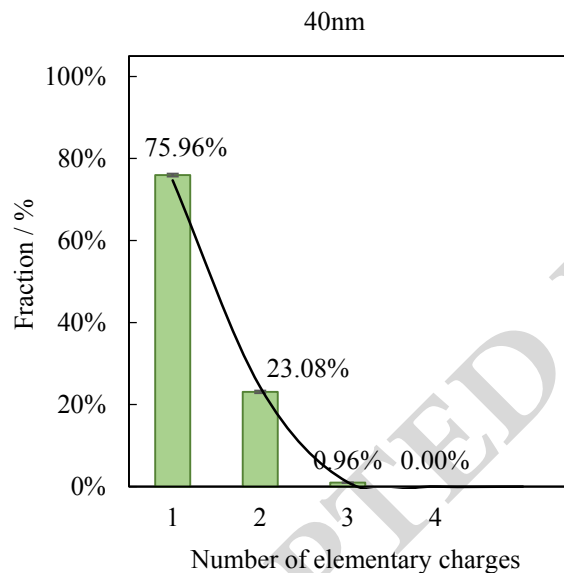
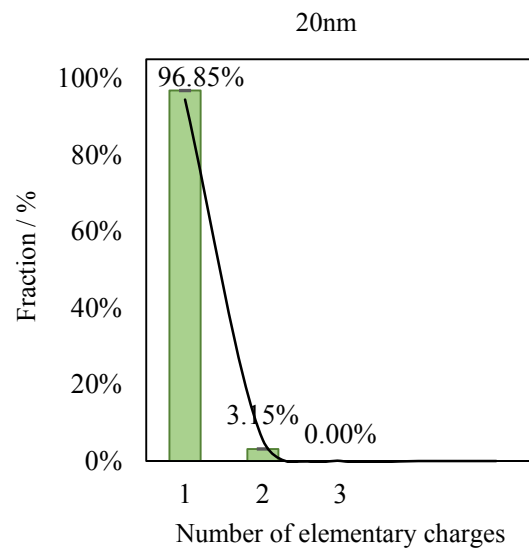
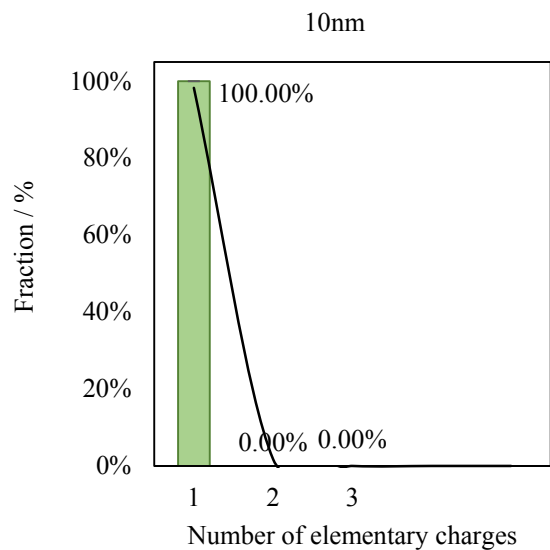


Figure 3



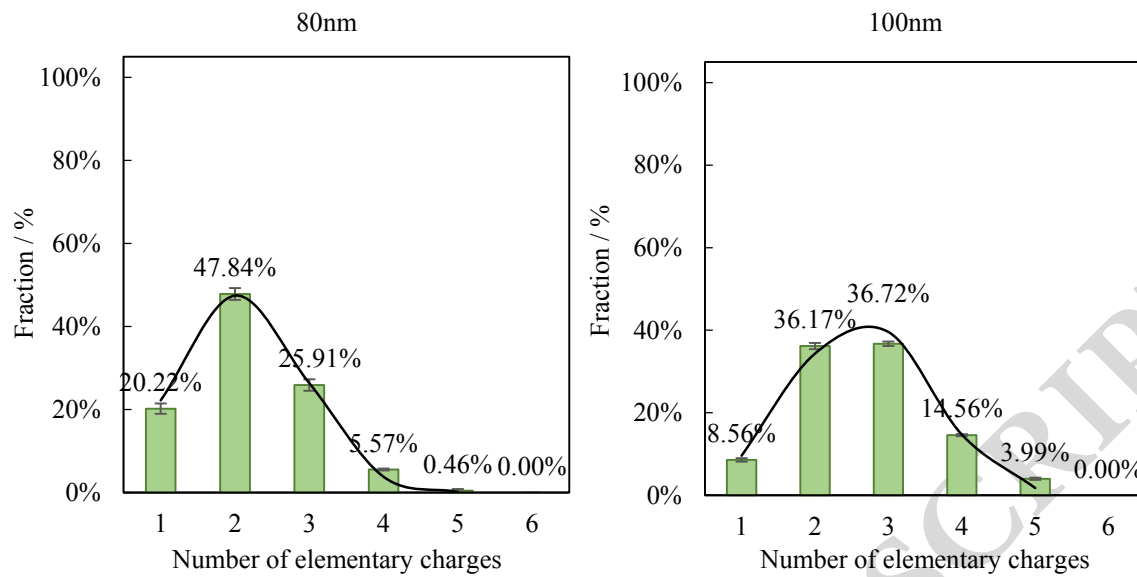


Figure 4

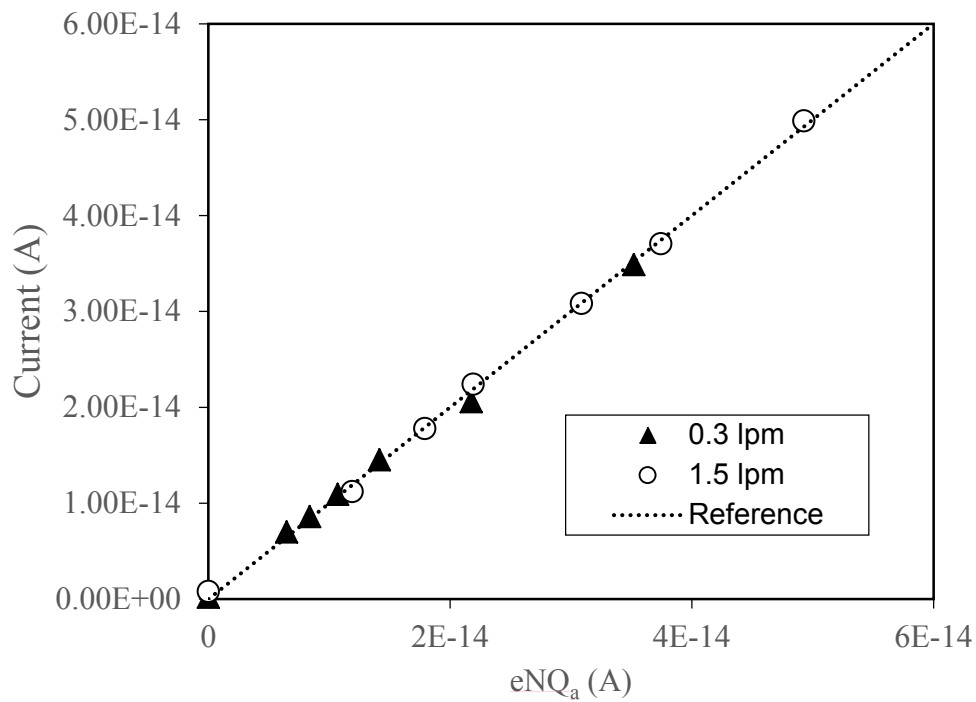
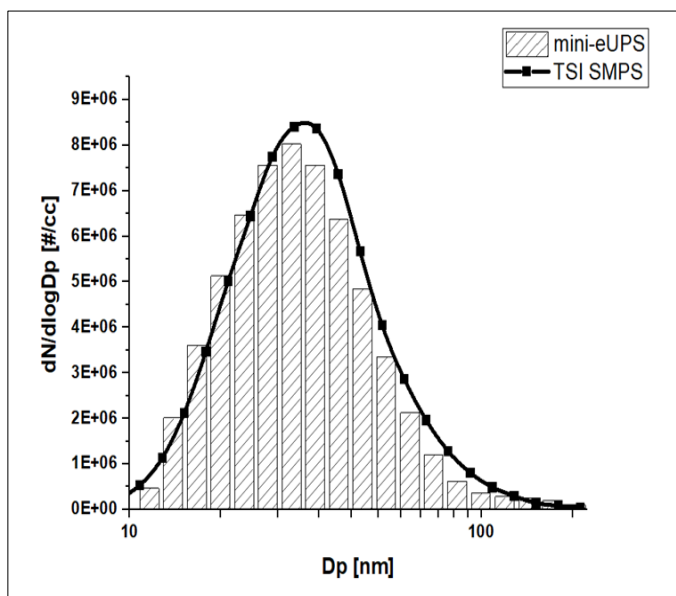
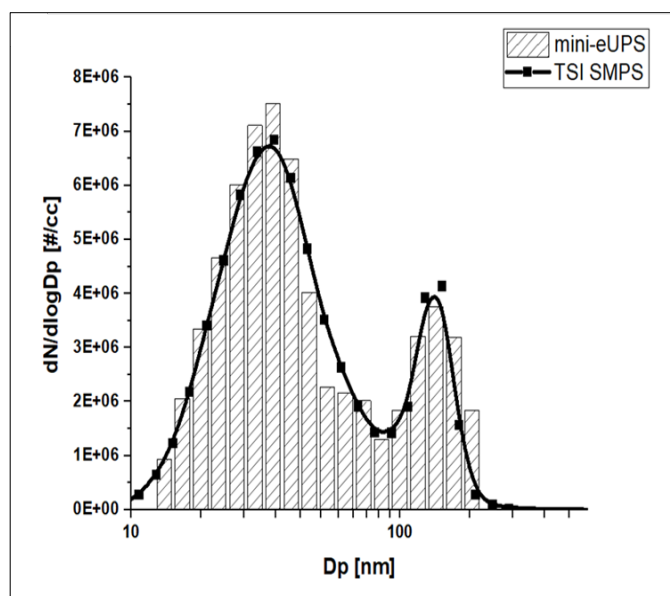


Figure 5



(a)



(b)

Figure 6

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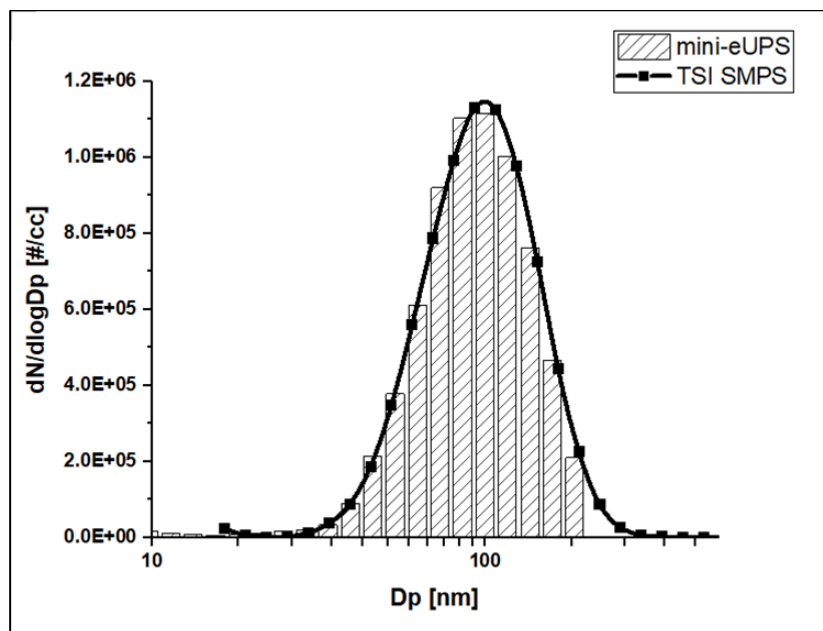


Figure 7

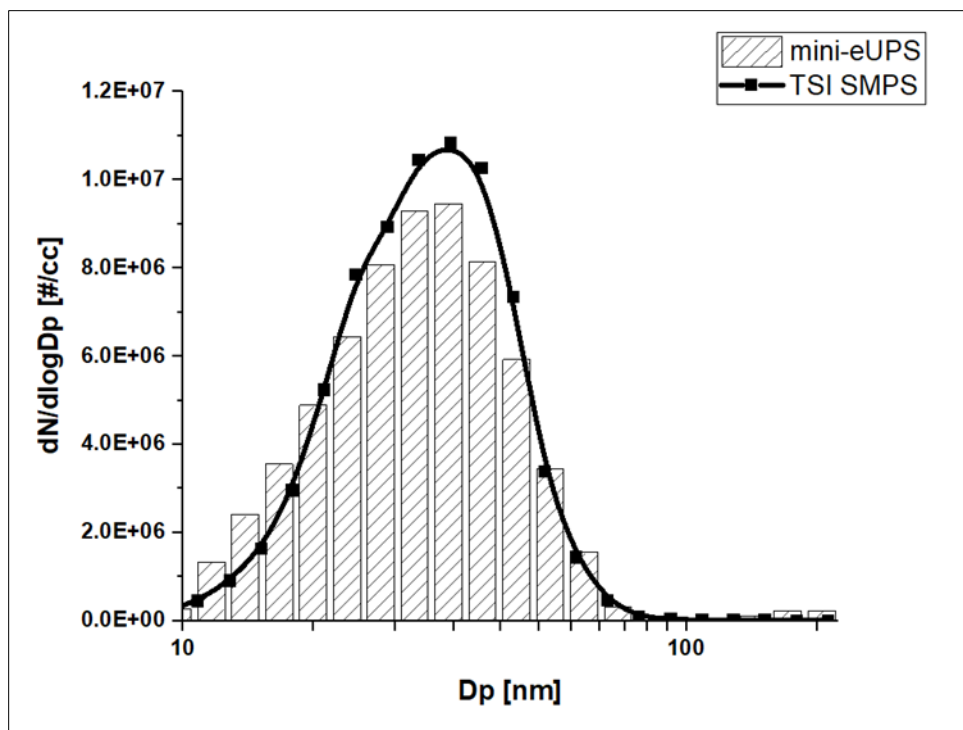


Figure 8