

1 **Health benefit assessment of China's National Action Plan on Air**
2 **Pollution in the Beijing-Tianjin-Hebei area**

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8
9 **Abstract**

10
11 To evaluate the effect of China's National Action Plan on Air Pollution (NAPAP), we assessed
12 the health benefits of PM_{2.5} remediation under the NAPAP from 2013 to 2017 in the Beijing-Tianjin-
13 Hebei (BTH) area using a relative risk model with real PM_{2.5} monitoring data and recent statistical
14 research data. The results revealed that the PM_{2.5} concentration in the BTH area decreased by 36 µg
15 m⁻³ (34.0%) under the NAPAP. PM_{2.5}-related mortality resulting from all causes, cardiovascular
16 disease, respiratory disease and lung cancer decreased to 58.1-65.2% of that in 2013; 102,133 PM_{2.5}-
17 related deaths were avoided, indicating a greater efficacy than the U.S. Cross-State Air Pollution
18 Rule. These results demonstrated that the NAPAP is effective and can be used a reference for other
19 countries to enact similar statutes.

20
21 **Keywords:** Health benefit assessment, PM_{2.5}, National Action Plan on Air Pollution, Beijing-Tianjin-
22 Hebei area

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23 INTRODUCTION

24

25 The definition of particulate matter with an aerodynamic diameter less than 2.5 μm ($\text{PM}_{2.5}$) was
26 established in 1997 by the U.S. Environmental Protection Agency to protect public health (Liang et
27 al., 2016). Since China began monitoring the ambient air $\text{PM}_{2.5}$ concentration nationwide in 2013,
28 $\text{PM}_{2.5}$ rapidly replaced PM_{10} as the most common chief pollutant (Pui et al., 2014); the 2013 annual
29 mean $\text{PM}_{2.5}$ concentration of the 74 cities required to monitor ambient air $\text{PM}_{2.5}$ concentration in
30 phase one (phase two for other cities began in 2016) was $72 \mu\text{g m}^{-3}$ (ranging from 26-160 $\mu\text{g m}^{-3}$),
31 and 95.9% of the cities failed to meet the secondary annual mean standard ($35 \mu\text{g m}^{-3}$) for ambient
32 air quality in China (Ministry of Ecology and Environment of the People's Republic of China, 2012).
33 $\text{PM}_{2.5}$ pollution was especially serious in the Beijing-Tianjin-Hebei (BTH) area (Zhang et al., 2017;
34 Cai et al., 2018), where none of the 13 cities met the standard and 7 were listed among the 10 cities
35 with the worst ambient air quality (Ministry of Ecology and Environment of the People's Republic of
36 China, 2014).

37 To improve the ambient air quality, as characterized by $\text{PM}_{2.5}$, the Chinese government enacted a
38 statute called the National Action Plan on Air Pollution (NAPAP, 2013-2017), which required that
39 the $\text{PM}_{2.5}$ concentration in the BTH area be reduced by 25.0% by 2017 compared with that in 2013.

40 Furthermore, Beijing was asked to reduce the $PM_{2.5}$ concentration to approximately $60 \mu g m^{-3}$ (The
41 State Council of the People's Republic of China, 2013). Many studies assessed the NAPAP by
42 forecasting the health benefits of $PM_{2.5}$ remediation attributed to the NAPAP (Lei et al., 2015; Fang
43 et al., 2016; Chen et al., 2017b; Maji et al., 2018). However, most studies selected the $PM_{2.5}$
44 concentration target of the NAPAP and utilized outdated population or mortality data as the
45 experimental data. Therefore, the real effect of the NAPAP is still not clear.

46 The aim of this study is to evaluate the effect of the NAPAP by assessing the health benefits of
47 $PM_{2.5}$ remediation resulting from the NAPAP in the BTH area between 2013 and 2017. The work is
48 based on a relative risk model using real monitoring data and recent statistical research data. We
49 hope the results will be useful for ambient air quality improvement and health risk control.

50 51 **MATERIALS AND METHODS**

52 53 *Study area*

54 The BTH area ($36^{\circ}05'N-42^{\circ}40'N$, $113^{\circ}27'E-119^{\circ}50'E$) is the largest urban agglomeration in
55 northern China and includes Beijing municipality, Tianjin municipality and 11 prefecture-level cities
56 of Hebei Province. In 2016, the area accounted for only 2.2% of the land area but 8.1% of the
57 population and 10.2% of the GDP of China (National Bureau of Statistics of the People's Republic

58 of China, 2017). The important position, dense population and poor ambient air quality caused
59 PM_{2.5}-related health problems to be very severe in the BTH area (Zheng et al., 2014; Song et al.,
60 2016, 2017).

61

62 ***Health risk model***

63 Because the exposure-response relationship between the level of exposure to PM_{2.5} and the
64 mortality was linear or near linear (Krewski et al., 2000; Pope III, 2000; National Research Council
65 (U.S.) Committee on Estimating the Health-Risk-Reduction Benefits of Proposed Air Pollution
66 Regulations, 2002), we selected a widely used epidemiological relative risk model (Broome et al.,
67 2015; Gao et al., 2016; Chen et al., 2017a) to quantify the association between long-term exposure to
68 ambient air PM_{2.5} and the risk of PM_{2.5}-related disease. The model is as follows:

69

$$RR = \exp[\beta \cdot (C - C_0)] \quad (1)$$

$$I = \left(\frac{RR - 1}{RR} \right) \cdot E \cdot P \quad (2)$$

70 In Eq. (1), C_0 represents the critical PM_{2.5} concentration, below which the health impact of PM_{2.5} is
71 considered to be approximately 0. C represents the exposed PM_{2.5} concentration. β is the exposure-
72 response coefficient, indicating changes in health effects. RR represents the relative risk.

73 In Eq. (2), P represents the exposed population. E represents the mortality of disease under PM_{2.5}

74 concentration C . I represents the $PM_{2.5}$ -related deaths, which we used to assess the health benefits of
75 $PM_{2.5}$ remediation resulting from the NAPAP.

76 The mortalities in this study are from all causes, cardiovascular disease, respiratory disease and
77 lung cancer, as the diameter of $PM_{2.5}$ is small enough to penetrate the bronchioles and alveoli,
78 affecting respiratory function (Turner et al., 2011). Therefore, long-term exposure to $PM_{2.5}$ mainly
79 contributes to the risk of developing cardiovascular disease, respiratory disease and lung cancer
80 (Englert, 2004; Franklin et al., 2007; Kim et al., 2015).

81
82 ***Scenario settings and data sources***

83 In this study, we denoted two scenarios—the baseline and the NAPAP—in the BTH area. The
84 baseline scenario represented the health risk in 2013, whereas the NAPAP scenario represented the
85 health risk under the NAPAP in 2017. The difference in I between the two scenarios was used to
86 assess the health benefits from the NAPAP.

87 The values of β used in this study, given in Table 1, were obtained by meta-analysis. To obtain
88 appropriate values, studies used in the meta-analysis were based on the following criteria: (1)
89 focusing on the long-term association of ambient $PM_{2.5}$ with mortality; (2) including at least one of
90 all-cause mortality, cardiovascular disease mortality, respiratory disease mortality and lung cancer

91 mortality; (3) providing effect estimates such as β with standard error or RR with 95% confidence
92 interval; and (4) studies about China or conducted in recent 3 years were given priority.

93 The values of E and P were mainly obtained from statistical yearbooks, reports and relative studies
94 of each city. For the 2017 data that remain unpublished, we estimated the values of E and P from
95 existing data.

96 C_0 is defined as the lowest $PM_{2.5}$ concentration monitored during the exposure time or a threshold
97 recommended by authoritative standards or studies (Apte et al., 2015; Fang et al., 2016). The C_0 of
98 each city in this study is the air quality guideline ($10 \mu\text{g m}^{-3}$) of the WHO (2006), which is a
99 worldwide guideline and close to the primary annual mean standard ($15 \mu\text{g m}^{-3}$) for ambient air
100 quality in China (Ministry of Ecology and Environment of the People's Republic of China, 2012).

101 The C under the baseline scenario is 2013 annual mean $PM_{2.5}$ concentration of each city. The C
102 under the NAPAP scenario is the $PM_{2.5}$ concentration of each city under the NAPAP. The annual
103 mean $PM_{2.5}$ concentration was obtained from the environmental status bulletin of each city.

104 To assess the health benefits from the NAPAP alone, we excluded the effects of meteorology
105 variations and estimated the $PM_{2.5}$ concentration under the NAPAP. For the three key cities (Beijing,
106 Tianjin and Shijiazhuang) in the BTH area, the average relative humidity decreased, and the average

107 wind speed overall increased between 2013 and 2017. Lower average relative humidity results in a
108 weaker secondary formation of $PM_{2.5}$, and higher average wind speed is conducive to the spread of
109 pollutants; thus, the diffusion conditions of the BTH area were considered to be better in 2017 on the
110 whole. In fact, according to official statements (Ministry of Ecology and Environment of the
111 People's Republic of China, 2018), the meteorological conditions for pollutant diffusion in the BTH
112 area in 2017 were slightly better than those in 2013; the meteorological variations contributed
113 approximately 5.0% to the $PM_{2.5}$ concentration reduction ratio (39.6%), and the NAPAP contributed
114 the remaining 34.6%, accounting for approximately 87.4% (34.6%/39.6%) of the reduction in $PM_{2.5}$
115 concentration. Thus, the $PM_{2.5}$ concentration under the NAPAP could be calculated based on the
116 2017 annual mean $PM_{2.5}$ concentration, the $PM_{2.5}$ concentration variation from 2013 to 2017 and the
117 NAPAP contribution.

118 The details and sources of the above parameters are listed in the Supplementary Material.

119 120 **RESULTS AND DISCUSSION**

121 122 ***$PM_{2.5}$ concentration***

123 We found that under the NAPAP scenario, the $PM_{2.5}$ concentration of each city in the BTH area
124 improved significantly compared with that under the baseline scenario, suggesting that the NAPAP

125 achieved its target. Fig. 1 shows the $PM_{2.5}$ concentration of each city under the two scenarios. The
126 $PM_{2.5}$ concentration in the BTH area under the NAPAP scenario was $70 \mu g m^{-3}$ (ranging from 32-95
127 $\mu g m^{-3}$), which is $36 \mu g m^{-3}$ (34.0%) lower than that under the baseline scenario (ranging from 40-
128 $160 \mu g m^{-3}$), thus achieving the target of a 25.0% decrease in $PM_{2.5}$ concentration. The decrease in
129 each city ranged from 20.0-43.8%. Every city except Zhangjiakou and Chengde outperformed the
130 25.0% target, and Xingtai decreased the most. The low decreases observed in Zhangjiakou and
131 Chengde may be attributed to their already low $PM_{2.5}$ concentrations in the BTH area, which were 40
132 and $49 \mu g m^{-3}$, respectively, under the baseline scenario. Beijing's $PM_{2.5}$ concentration under the
133 NAPAP scenario was $62 \mu g m^{-3}$, and this city also achieved its goal.

134 135 ***PM_{2.5}-related mortality***

136 The $PM_{2.5}$ -related mortality and $PM_{2.5}$ -related mortality ratio (the contribution of the $PM_{2.5}$ -related
137 mortality to the overall mortality) in the BTH area under the NAPAP scenario also notably decreased
138 from that under the baseline scenario.

139 Fig. 2 shows the $PM_{2.5}$ -related mortality and the ratio from different causes of deaths in the BTH
140 area under the two scenarios. The $PM_{2.5}$ -related all-cause mortality was 1.32‰ under the NAPAP
141 scenario, accounting for 58.1% of that under the baseline scenario. The $PM_{2.5}$ -related mortality of

142 three specific diseases, i.e., cardiovascular disease, respiratory disease and lung cancer, were 0.83,
143 0.15 and 0.11‰ respectively, under the NAPAP scenario, accounting for 60.6-65.2% of that under
144 the baseline scenario.

145 The ratio of PM_{2.5}-related all-cause mortality was 23.6% under the NAPAP scenario, 11.0
146 percentage points lower than that under the baseline scenario. The PM_{2.5}-related mortality ratios of
147 cardiovascular disease, respiratory disease and lung cancer were 26.7, 33.4 and 25.0% under the
148 NAPAP scenario, which were 12.1, 13.9 and 11.5 percentage points lower, respectively, than those
149 under the baseline scenario.

150 151 ***Health benefits***

152 We found that the NAPAP had significant health benefits. Fig. 3 shows the deaths and PM_{2.5}-
153 related deaths in the BTH area under the two scenarios. For the NAPAP scenario, there were 621,488
154 deaths, 146,941 (23.6%) of which were related to PM_{2.5}. For the baseline scenario, there were
155 719,084 deaths, 249,074 (34.6%) of which were related to PM_{2.5}. Compared with the baseline
156 scenario, 102,133 PM_{2.5}-related deaths were avoided in the BTH area due to the NAPAP.

157 The PM_{2.5}-related deaths of the three specific diseases listed above also decreased under the
158 NAPAP scenario, accounting for 60.2-65.4% of that under the baseline scenario. For cardiovascular

159 disease, there were 344,611 deaths under the NAPAP scenario, 91,985 (26.7%) of which were
160 related to PM_{2.5}, which was 58,085 (12.1 percentage points) fewer deaths than that under the baseline
161 scenario. For respiratory disease, there were 49,233 deaths under the NAPAP, 16,448 (33.4%) of
162 which were related to PM_{2.5}, which was 8,693 (13.9 percentage points) fewer deaths than that under
163 the baseline scenario. For lung cancer, there were 46,829 deaths under the NAPAP scenario, 11,713
164 (25.0%) of which were related to PM_{2.5}, which was 7,740 (11.4 percentage points) fewer deaths than
165 that under the baseline scenario.

166 Fig. 4 shows the spatial distribution of PM_{2.5}-related deaths in the BTH area under the two
167 scenarios. The average number of PM_{2.5}-related deaths in each city in the BTH area under the
168 NAPAP scenario was 11,303 (ranging from 2,479-21,022), 7,856 fewer than that under the baseline
169 scenario (ranging from 4,171-34,718). Shijiazhuang, Beijing and Tangshan were the top 3 most-
170 improved cities, with an average of 13,413 avoided deaths (ranging from 12,149-14,393), while
171 Qinhuangdao, Chengde and Zhangjiakou were the bottom 3 cities, with an average of 1,596 avoided
172 deaths (ranging from 1,476-1,692).

173 The NAPAP was shown to be more effective than the U.S. Cross-State Air Pollution Rule. The
174 latter rule led to approximately 23,500 avoided deaths per year (Lei et al., 2015), approximated to

175 117,500 avoided deaths in 5 years for the whole U.S., which is 15,367 (15.0%) more than the
176 avoided deaths in the BTH area under the NAPAP scenario. However, considering the U.S.
177 population was approximately 3 times that of the BTH area, the NAPAP was obviously more
178 efficient.

179 Economic benefits arising from improved health in the BTH area of the NAPAP were also
180 significant and worthwhile. According to related studies (Zeng and Jiang, 2010; Huang and Zhang,
181 2013), the value of statistical life in health costs attributable to China's air pollution was
182 approximately 1 million CNY per person. Therefore, the economic benefits brought about by
183 improved health alone reached approximately 102.1 billion CNY, which accounted for 41.0% of the
184 NAPAP direct investment in the BTH area (Chinese Academy for Environmental Planning, 2015).

185 186 *Comparison with related studies*

187 Previous studies have focused on the same question addressed here. By assuming the exact
188 NAPAP target, (Lei et al., 2015) estimated that the avoided deaths in the BTH area should be 89,000,
189 which was 13,133 (12.9%) fewer deaths than our results. The difference arose from two reasons. The
190 first reason was the difference in β . The β in the previous study, which took the lower limit of that
191 used in the U.S., was 3.0%/10 $\mu\text{g m}^{-3}$ (ranging from 2.0-4.0%/10 $\mu\text{g m}^{-3}$), 1.3%/10 $\mu\text{g m}^{-3}$ lower than

192 our value. The second reason was the uncertainty in $PM_{2.5}$ concentrations. The $PM_{2.5}$ concentrations
193 in the previous study were from Community Multiscale Air Quality, an air quality forecasting model,
194 which may cause uncertainty in comparison with real monitoring data. Furthermore, the previous
195 study showed a 34.0% decrease in the $PM_{2.5}$ concentrations but provided no detailed figures.

196 Under the same assumptions, (Chen et al., 2017b) estimated that the deaths prevented in the BTH
197 area should be 70,255 (14,051 per year over 5 years), which was 31,878 (31.2%) fewer deaths than
198 in our study. The difference was mainly attributed to the difference in the $PM_{2.5}$ concentration. The
199 previous study obtained 2012 $PM_{2.5}$ monitoring data from Beijing, Tianjin and Shijiazhuang and
200 selected 2012 as the baseline year. However, the official monitoring and reporting of $PM_{2.5}$
201 concentrations in China began in 2013, and to the best of our knowledge, no $PM_{2.5}$ concentrations
202 were publicly available in 2012. The 2012 concentration in the previous study was $84 \mu g m^{-3}$, which
203 was $22 \mu g m^{-3}$ lower than that in our baseline year. Moreover, the concentration in the previous study
204 decreased by approximately $21 \mu g m^{-3}$ (25%) from the baseline year to 2017, which was $15 \mu g m^{-3}$
205 lower than the actual result.

206 According the research results of (Yang et al., 2013), the ambient air $PM_{2.5}$ -related mortality
207 ranked fourth in causes of mortality after diet (30.6%, among all reported deaths), high blood

208 pressure (24.6%), and tobacco (16.4%). Our PM_{2.5}-related mortality ratio was 34.6% under the
209 baseline scenario and 23.6% under the NAPAP scenario, both of which are much higher than that of
210 tobacco. However, the study period in the previous study was from 1990 to 2010, and the lack of
211 PM_{2.5} monitoring data and the rapid change in China in the years after may have led to the
212 differences. Nonetheless, the previous study provided a reference; that is, the NAPAP reduced PM_{2.5}-
213 related health risks from a level exceeding the risk factor for diet to a level near that of high blood
214 pressure.

215 Though our results were robust and realistic, some limitations should be noted. Since certain data
216 were unavailable, some of the populations and mortalities were estimated rather than drawn from
217 public statistics. One key index with which to assess the health benefit—the years of life lost
218 (YLL)—could not be obtained, and the exposure-response coefficient was taken from a meta-
219 analysis of the latest study data rather than authoritative results. Thus, we suggest that further studies
220 should focus on local exposure-response coefficients and improving assessments by incorporating
221 the YLL.

222 223 **CONCLUSIONS**

224
225 We verified the effect of the NAPAP by assessing the health benefits of PM_{2.5} remediation in the

226 BTH area under the NAPAP. To the best of our knowledge, this is the first and the most accurate
227 study assessing the NAPAP in the BTH area using real PM_{2.5} monitoring data and recent statistical
228 research data.

229 The NAPAP was effective, as the PM_{2.5} concentration under the NAPAP in the BTH area
230 decreased by 36 $\mu\text{g m}^{-3}$ (34.0%) from 2013 to 2017, achieving the target set for the area. The PM_{2.5}-
231 related mortality caused by all causes, cardiovascular disease, respiratory disease and lung cancer in
232 the BTH area under the NAPAP was 1.32, 0.83, 0.15 and 0.11%, respectively, accounting for 58.1-
233 65.2% of the values under the baseline scenario. The PM_{2.5}-related mortality ratio under the NAPAP
234 was 23.6%, 11.0 percentage points lower than that under the baseline scenario.

235 Under the NAPAP, 102,133 deaths were avoided in the BTH area and 102.1 billion CNY was
236 saved. The NAPAP was more efficient in improving health than the U.S. Cross-State Air Pollution
237 Rule. Though the PM_{2.5}-related mortality in the BTH area decreased from a value exceeding diet-
238 related mortality and reaching the same mortality risk as that of high blood pressure under the
239 NAPAP, the PM_{2.5}-related mortality was still high.

240 This study demonstrated the effect of the NAPAP and proved that the ambient air quality could be
241 improved through effective state planning. We believe that the NAPAP can provide a reference for

242 countries to improve the ambient air quality and, furthermore, promote the improvement of ambient
243 air quality worldwide.

244

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246

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250

251 **DECLARATIONS OF INTEREST**

252

253 None.

254

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342 **Table 1.** Values of the exposure-response coefficients (β).

Causes of death	β (%/10 $\mu\text{g m}^{-3}$)
All causes	4.30
Cardiovascular disease	4.97
Respiratory disease	6.69
Lung cancer	4.66

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Figure Captions

345 **Fig. 1.** PM_{2.5} concentration ($\mu\text{g m}^{-3}$) of each city in the BTH area under the two scenarios

346 **Fig. 2.** PM_{2.5}-related mortality and PM_{2.5}-related mortality ratio in the BTH area under the two
347 scenarios

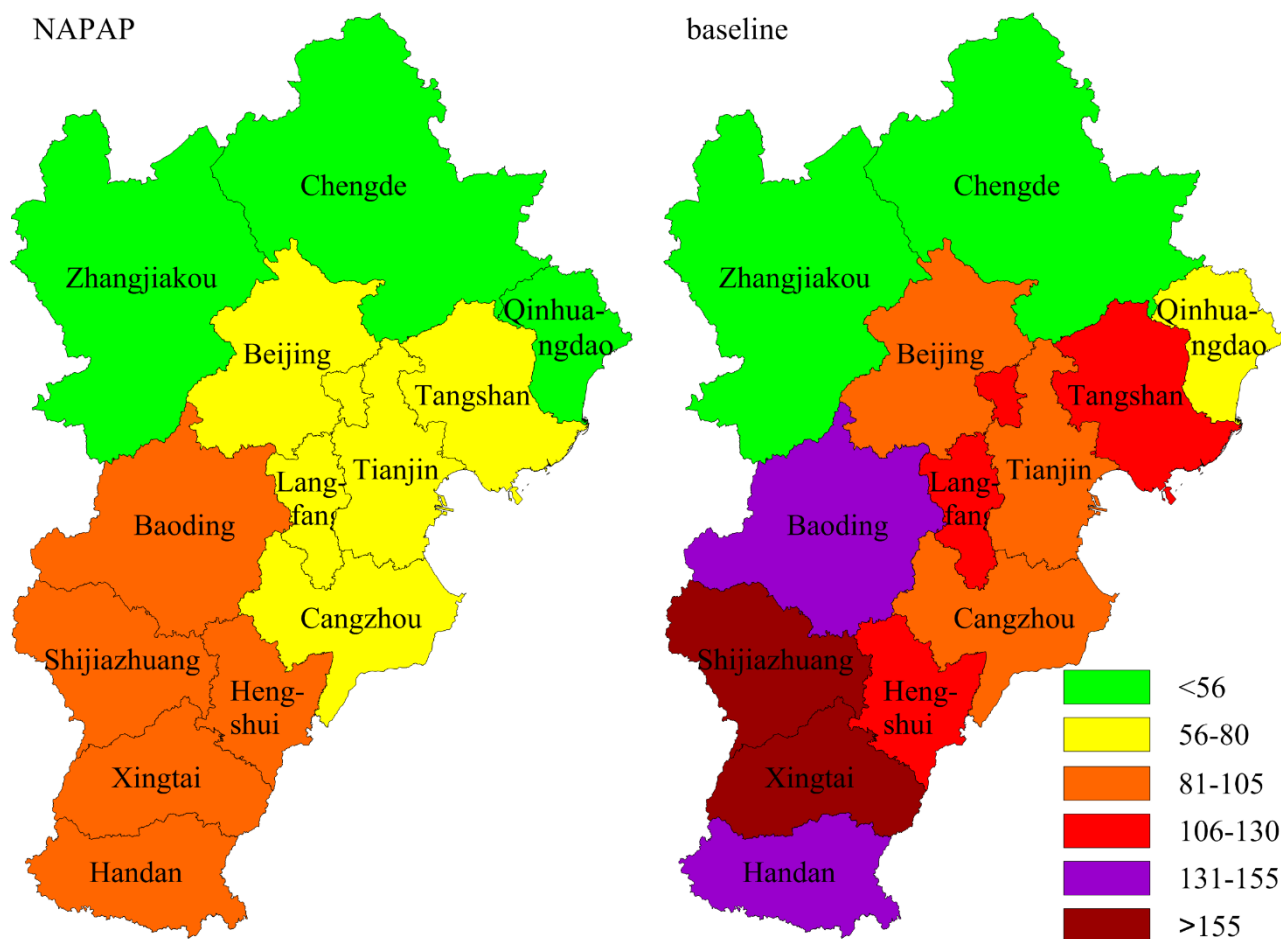
348 **Fig. 3.** Deaths and PM_{2.5}-related deaths in the BTH area under the two scenarios

349 **Fig. 4.** PM_{2.5}-related deaths (thousands) in the BTH area under the two scenarios

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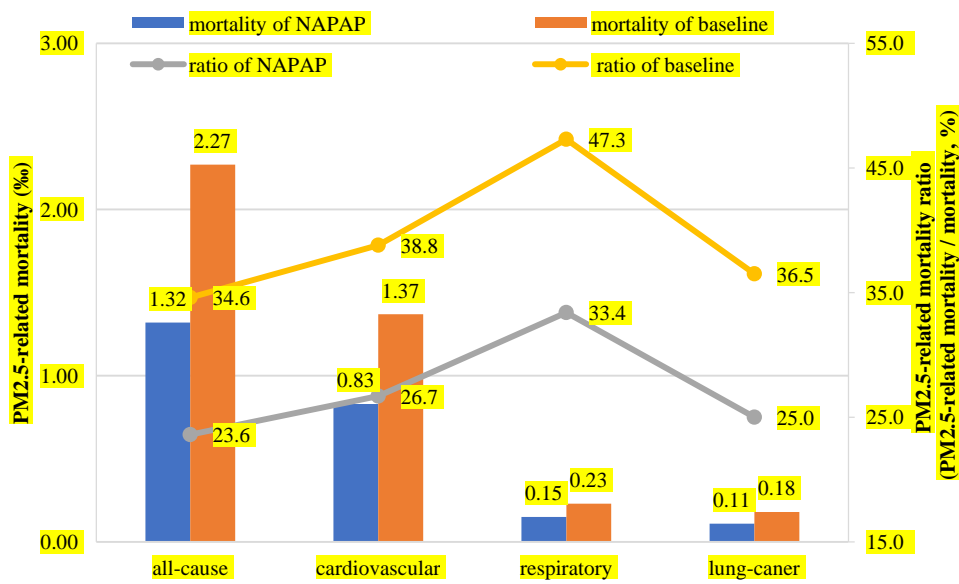


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Fig. 1. PM_{2.5} concentration (μg m⁻³) of each city in the BTH area under the two scenarios

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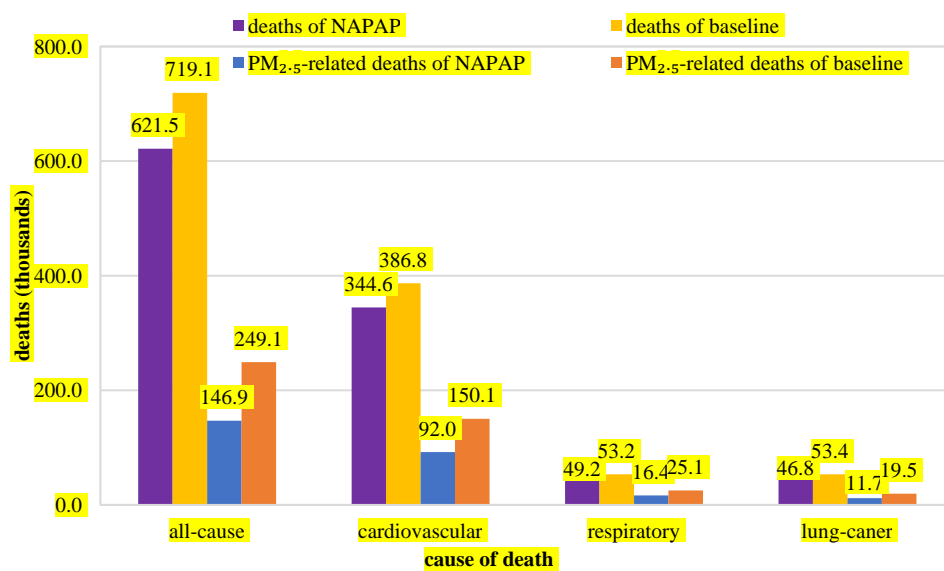
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Fig. 2. PM_{2.5}-related mortality and PM_{2.5}-related mortality ratio in the BTH area under the two

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scenarios

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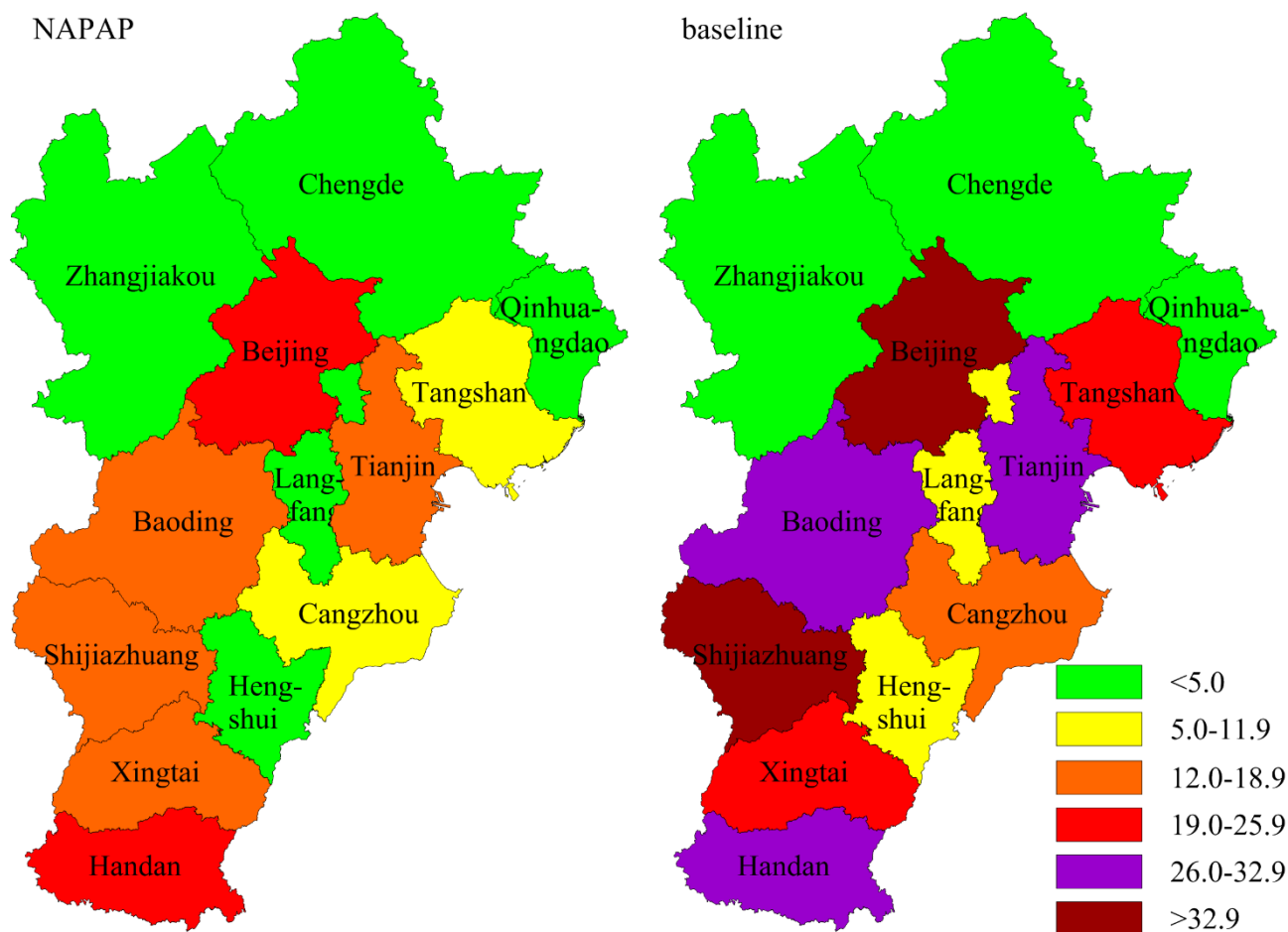
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Fig. 3. Deaths and PM_{2.5}-related deaths in the BTH area under the two scenarios

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Fig. 4. PM_{2.5}-related deaths (thousands) in the BTH area under the two scenarios

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