

The Real-world Emissions from Urban Freight Trucks in Beijing

Haohao Wang^{a,b}, Yunshan Ge^{a,b}, Jianwei Tan^{a,b*}, Legang Wu^c, Pengcheng Wu^d, Lijun
Hao^{a,b}, Zihang Peng^{a,b}, Chuanzhen Zhang^{a,b}, Xin Wang^{a,b}, Yaxin Han^{a,b}, Mengzhu
Zhang^{a,b}

^a National Laboratory of Auto Performance & Emission Test, School of Mechanical Engineering,
Beijing Institute of Technology, Beijing 100081, China PR

^b Collaborative Innovation Centre of Electric Vehicles in Beijing, Beijing 100081, China PR

^c Kunming Sino-Platinum Metals Catalyst Co. Ltd, Kunming 650106, China PR

^d Graduate School of Engineering, Northeastern University, Boston MA, 02115, USA

*Corresponding Author
Email: tanjianwei@bit.edu.cn
TelePhone:00861068912035

11 **ABSTRACT:** The emissions from 7 urban freights were tested using portable
12 emission measurement system, and analyzed based on VSP method. The results show
13 that both gaseous pollutants and PM emissions increase with the rising VSP. With
14 regard to emissions of CO and HC, they decrease with the elevated speed and
15 acceleration. PM and NO_x emitted from China IV vehicles are significantly less than
16 those from China III vehicles, while emissions of CO and HC exhibit different
17 tendency and further work should be performed. In addition, speed and acceleration
18 show slight influence on NO_x emissions. The VSP plays an important role in the
19 emissions from urban freight trucks. Therefore, a comprehensive evaluation of the
20 emission characteristics of the urban freight trucks in megacities is needed for taking
21 effective measures.

22 **Key words:** PEMS, VSP, urban freight vehicles, emission

23 **1 Introduction**

24 With the remarkable development of social economic and urbanization in the
25 past years in China, the vehicle population (excluding rural vehicles) increased from
26 5.5 million in 1990 to 194 million in 2016 (NBSC, 2016; Wu *et al.*, 2017). In
27 particular, the average annual growth rate has arrived at 16% in recent years (NBSC,
28 2015). Meanwhile, as one of the most important contributors to urban air pollution,
29 vehicle emission has attracted considerable attention. As for Beijing, a largest
30 metropolis placing strict standard on vehicle emissions in China, its vehicle ownership
31 is approximately 250 per 1000 people, which is much higher than the national average
32 (NBSC, 2015). Thus, the air quality in Beijing confronts severe challenge.

33 Hitherto, several approaches have been put forward to reduce the impact on
34 urban air quality from vehicle emissions, which consist of improving transportation
35 planning and traffic management, and implementation of more stringent emission
36 standards (Yao *et al.*, 2015a; Wu *et al.*, 2017). With the implementation of these
37 measures, although the air quality has been improved, several investigations have
38 reported that there is no significant reduction in the emission factors of NO_x in the
39 on-road measurements by using portable emission measurement system (PEMS) (Huo
40 *et al.*, 2012; Shen *et al.*, 2015; Yao *et al.*, 2015c; Zhang *et al.*, 2016a). Up to now,
41 there have been numerous publications focused on emission measurement under the
42 real-world driving conditions by using PEMS (Khan *et al.*, 2012; Fu *et al.*, 2013;
43 Yang *et al.*, 2015; Yao *et al.*, 2015b; Liu *et al.*, 2017) and the available results provide
44 the data support for vehicle pollution control.

45 In addition, several works have reported the diesel truck emissions with different
46 approaches (Huo *et al.*, 2011; Tsai *et al.*, 2011; Huo *et al.*, 2012; Yao *et al.*, 2015c;
47 Yang *et al.*, 2017). Yang *et al.* (Yang *et al.*, 2015) adopted the road emission
48 intensity-based (REIB) approach to estimate the emissions of diesel freight trucks and
49 established the province-based emission inventory. Zheng *et al.* (Zheng *et al.*, 2015)
50 obtained the black carbon emissions from twenty-five heavy-duty diesel vehicles by
51 using PEMS. Yao *et al.* (Yao *et al.*, 2015a) investigated the on-road emissions

52 characteristics of nine heavy-diesel trucks in Xiamen. Furthermore, in order to
53 investigate urban diesel vehicle emissions, many works were focused on bus
54 emissions. Yu et al. (Yu *et al.*, 2016) pointed out that the diesel bus emissions and fuel
55 consumption based on the real-world test were related to vehicle's speed and
56 acceleration. López-Martínez et al. (López-Martínez *et al.* 2017) estimated the
57 emissions of urban bus fleets in Madrid using an integrated methodology. Our
58 previous work studied the bus emissions and found that nanometer size particle
59 played a key role in PN (Liu *et al.* 2011).

60 Based on the aforementioned review, it is shown that the investigations on the
61 urban freight truck emissions are still scarce, especially in megacity. Although Zhang
62 et al. (Zhang *et al.*, 2016b) reported that the characteristics of gaseous and particulate
63 pollutants from three logistics transportation diesel vehicles in Chengdu China, the
64 emissions of Beijing are different from Chengdu due to the differences in road
65 conditions, traffic management, the ownership of vehicles and the freight
66 transportation. In addition, freight trucks for supermarket and logistic companies are
67 not limited by tail number requirements in Beijing. Such kinds of trucks can travel
68 from Monday to Sunday inside the fifth ring. The emissions of urban freight trucks
69 have made huge contributions to the total vehicle emissions. Thus, on-road emissions
70 of seven urban freight trucks were measured with PEMS in Beijing in this work. The
71 influence of vehicle specific power (VSP) on gaseous pollutants and PM was
72 addressed. The relevant conclusions will not only be helpful for understanding the real
73 world emission of urban freight trucks in Beijing, but also provide reference for the
74 other provinces in China and enrich the emission data.

75 **2 Materials and methods**

76 **2.1 Tested trucks and Routes**

77 The on-road emission tests were conducted in 2016 in Beijing. Seven vehicles in
78 different types were selected in the tests, including four diesel trucks of the China III
79 emission standard and three light-duty diesel trucks of the China IV emission standard.
80 The specifications of the tested vehicles are listed in Table 1. Each vehicle meets the

81 latest local emission standards. The sulfur contents of fuel obtained from the local
82 retail petrol stations in the tests are in accordance with Chinese standard
83 GB17040-2008.

84 The test routes were nearly the daily routes of these vehicles. The test area was in
85 the northwest of Beijing city, which covered from the third ring road to the fifth ring
86 road. Transportation environment simulated the actual situations of tested vehicles
87 they faced every day. According to the local traffic bans, trucks should not be driven
88 at 7:00-9:00 am and 16:00-20:00 pm.

89 **2.2 Measurement systems**

90 As shown in Fig. 1, in order to collect the data about instantaneous exhaust
91 emissions and driving pattern on the road, PEMS was employed to measure the
92 emissions from freight trucks. PEMS consisted of two sections: SEMTECH-DS
93 (Sensors Inc., U.S.) and Pegasor particle sensors MI2 (PPS-M, Pegasor Inc., Finland).
94 The SEMTECH-DS was used to measure gaseous pollutants. In this device, CO and
95 CO₂ concentrations were measured by non-dispersive infrared (NDIR); total
96 hydrocarbon (THC) was measured by flame ionization detector (FID); NO_x was
97 determined by non-dispersive infrared (NDIR). Connected with the Semtech-DS, the
98 weather station was used to detect ambient temperature, humidity and pressure. The
99 Global Positioning System (GPS) was used to obtain the speed, acceleration and road
100 grade information, which was indispensable for subsequent VSP calculation. Vehicle
101 exhaust flow rate and temperature were determined by exhaust flow meter (EFM),
102 which was installed in the upstream of the exhaust port.

103 The PPS-M was used to record the instantaneous particulate matter emission data.
104 This instrument was based on particle charging and electrical detection of charged
105 particles. The PPS-M was capable of measuring airborne particle size distribution
106 ranging from 23 nm to 2.5 μm. The laptop was adopted to collect second by second
107 data and control all the instruments.

108 It is worth noting that the equipment needed to be warmed up for at least one
109 hour before the measurement. After the warming-up, it was necessary to examine the

110 air leakage of the exhaust pipes and other flow devices. To ensure the accuracy of data,
111 the Semtech-DS must be zeroed and calibrated prior to each separate test (Fu *et al.*,
112 2012; Peng *et al.*, 2016). Besides, fuel consumption was calculated with carbon
113 balance method on the basis of CO₂, CO and THC emission data.

114 2.3 Calculation of VSP

115 As for the data analysis, emission factors were easily influenced by factors, such
116 as engine load. In this work, CO₂-based emissions and VSP were selected to analyze.
117 The VSP is highly correlated with vehicle emission. It is defined as the engine power
118 output per vehicle unit mass, which is a practical indicator of real-world driving
119 emissions (Frey *et al.*, 2010) and is expressed as a function of vehicle speed, road
120 grade and acceleration. It could be expressed as Eq. 1:

$$121 \quad VSP = v(a + g \sin \theta + gC_R) + \frac{1}{2} \rho_a \frac{C_D A}{m} v^3 \quad (1)$$

122 where VSP is vehicle specific power (kW/Kg); v is the vehicle speed (m/s); g is the
123 gravitational acceleration, 9.807 m/s²; θ is the road grade, 0; a is the vehicle
124 acceleration, (m/s²); C_R is rolling resistance term coefficient, 0.012; ρ_a is the air
125 density, 1.19275 kg/m³; C_D is air resistance term coefficient, 0.00302; m is the mass of
126 the test vehicles (kg). In order to analyze the pollutant emission rate data of seven
127 urban freight vehicles, the VSP data were divided into five different bins, namely,
128 $VSP \leq 0$, $0 < VSP \leq 0.5$, $0.5 < VSP \leq 1$, $1 < VSP \leq 2$, $VSP > 2$.

129 3 Results and discussion

130 3.1 Characteristics of CO₂ emission

131 Fig. 2 presents the CO₂ emission rates of the tested vehicles in different VSP. It
132 clearly shows that all the tested vehicles under the China IV standard emission rates
133 increase with the elevated VSP and reach the top emission level when they are at the
134 highest VSP situation. However, there is a valley at $0.5 < VSP \leq 1$ in terms of the test
135 vehicles under the China IV standard and then increases again to the peak. Taking the
136 China III-1 as an example, the CO₂ emission rates at different VSP bins are 3.42 g/s,
137 3.90 g/s, 3.12 g/s, 4.15 g/s, 5.22 g/s, respectively. The CO₂ emission rate of $VSP > 2$

138 increases 52.6% in comparison with that of $VSP \leq 0$. All of the tested vehicles
139 approach the maximum when $VSP > 2$, which means their fuel consumptions also
140 increase according to the carbon balance theory, and their fuel economy get worse.

141 **3.2 Characteristics of gaseous pollution**

142 Through analyzing the data obtained based on VSP bins, the gaseous pollution
143 emission rate vs. VSP, the gaseous pollution emission rate based on CO_2 vs. vehicle
144 speed and the gaseous pollution emission rate based on CO_2 vs. VSP are illustrated as
145 Fig.3, Fig.4, Fig.5, respectively. For both China III vehicles and China IV vehicles, it
146 clearly shows that the CO, HC and NO_x emission rates generally increase with the
147 rising VSP. With regard to CO, HC and NO_x emissions based on CO_2 , a similar
148 tendency is also observed for China III vehicles and China IV vehicles. Zhang et al.
149 (Zhang *et al.*, 2016b) and Huang et al. (Huang *et al.*, 2013) demonstrated that the
150 gaseous emission rates are mainly affected by engine power rather than vehicle speed.

151 For NO_x emissions based on CO_2 , as shown in Fig.5, NO_x emission in China IV
152 vehicles are less compared with China III vehicles. This phenomenon is consistent
153 with previous works (Huo *et al.*, 2012; Wu *et al.*, 2012; Yao *et al.*, 2015c) and the
154 main reason is that the installation of EGR control devices in diesel trucks (Yao *et al.*,
155 2015c). In general, due to the presence of advanced electronic fuel injection system,
156 the oxidation catalyst in exhaust gas after-treatment system (Yao *et al.*, 2015c) in
157 China IV vehicles, the emission rates of CO and HC should be lower compared to
158 China III vehicles. However, there is some differences in some bins, as shown in Fig.3
159 and Fig.4 for CO and HC emissions, and additional tests should be performed in
160 future work.

161 There is a decline tendency with the increasing speed and acceleration for CO
162 emissions based on CO_2 in Fig.3b. This phenomenon occurs because vehicles need
163 much more air and fuel to generate enough power when speed and acceleration rise.
164 As a result, the CO emission rate becomes higher. However, the engine has become
165 worse when speed and acceleration are low; thus, the combustion efficiency is low
166 and the vehicle generates more CO. When the vehicle go into the condition of high
167 speed and acceleration, mobility of air in cylinder and atomization of fuel all get better.

168 The economy of fuel also gets better. In addition, a higher value of China III-3 vehicle
169 can be clearly seen in the figure and the main reason is that the displacement of this
170 truck is 4.75L, which is far more than others.

171 A similar tendency to CO emission of HC emission characteristics can be found
172 in Fig.4b. The causes of this phenomenon mainly comprise the following several
173 aspects. Firstly, it is well known that the fuel incomplete combustion leads to the
174 production of HC, and fuel is not effectively burnt, which results in higher HC
175 emission. Secondly, the high rotation speed enhances the streaming mixture and eddy
176 diffusion at a high speed and acceleration in the cylinder. Thirdly, part of HC would
177 be transformed to other CO at higher temperature and the cylinder wall temperature is
178 rising, which will shorten the quenching distance to alleviate HC production (Wu *et*
179 *al.*, 2012; Wang *et al.*, 2016). Therefore, HC emission based on CO₂ decreases with
180 the increasing speed and acceleration, while HC emission increase with the rising
181 VSP.

182 As illustrated in Fig.5b, the value of NO_x emissions based on CO₂ all maintain in
183 a relatively steadfast. The changes of speed and acceleration have a very slight
184 influence on the NO_x emission based on CO₂. Similar results were reported by Yao *et*
185 *al.* (Yao *et al.*, 2015c), who concluded that NO_x emission level based on CO₂ mainly
186 depended on the bursting state in the cylinder rather than the actual on-road driving
187 conditions.

188 **3.3 Characteristics of PM emission**

189 Regarded as one of the main pollutants from diesel emission, PM has attracted
190 widespread concern (He *et al.*, 2017). Characteristics of PM emission rates are shown
191 in Fig.6. It is known that the PM emission limit is lower with the more strict standards
192 and the advanced after-treatment technologies are employed from China III to China
193 IV. Thus, it is found that the PM emission rate of all China IV trucks is lower
194 compared to all China III trucks in Fig. 6a. In addition, as the same reasons like
195 gaseous pollutions, it is concluded that PM emission rate and PM based on CO₂
196 increase with the growing VSP value.

197 **4 Conclusions**

198 To better understand the emission characteristics of urban freight trucks in
199 Beijing, the gaseous pollutants and PM by using PEMS based on VSP were measured.
200 The relevant conclusions drawn from this work were listed as follows:

201 (1) Both gaseous pollutants and PM emissions increase with the rising VSP;

202 (2) Compared with China III vehicles, PM and NO_x emission in China IV
203 vehicles are less. Speed and acceleration have a very slight influence on the NO_x
204 emission based on CO₂;

205 (3) There is a decline tendency with the increasing speed and acceleration for
206 both CO and HC emissions based on CO₂;

207 (4) The urban freight vehicles drive under high VSP and negative VSP condition
208 reasonably is necessary in Beijing should be avoided. Cities not only need to develop
209 more specialized roads for those vehicles with special functions, but also increase the
210 penalty for the illegal. In addition, the urban freight vehicles should avoid overweight
211 and suddenly speed up or down;

212 **Acknowledgments**

213 This work is conducted with funding from the project supported by the National
214 High-tech Research and Development Program (Grant No.2016YFC0208005) and the
215 National Natural Science Foundation of China (Grant No. 51476012).

216 References

- 217 Frey, H.C., Zhang, K. and Roupail, N.M. (2010). Vehicle-specific emissions modeling based upon
218 on-road measurements. *Environ. Sci. Technol.* 44: 3594-3600.
- 219 Fu, M., Ge, Y., Tan, J., Zeng, T. and Liang, B. (2012). Characteristics of typical non-road machinery
220 emissions in China by using portable emission measurement system. *Sci. Total Environ.* 437:
221 255-261.
- 222 Fu, M., Ge, Y., Wang, X., Tan, J., Yu, L. and Liang, B. (2013). NO_x emissions from Euro IV busses
223 with SCR systems associated with urban, suburban and freeway driving patterns. *Sci. Total
224 Environ.* 452: 222-226.
- 225 He, C., Li, J., Wang, Y., Tan, J., Song, G., Jia, D. and Zhao, L. (2017). Size-segregated particulate
226 matter emission characteristics of a heavy-duty diesel engine with oxygenated fuels. *Appl.
227 Therm. Eng.* 125: 1173-1180.
- 228 Huang, C., Lou, D., Hu, Z., Feng, Q., Chen, Y., Chen, C., Tan, P. and Yao, D. (2013). A PEMS study
229 of the emissions of gaseous pollutants and ultrafine particles from gasoline- and diesel-fueled
230 vehicles. *Atmos. Environ.* 77: 703-710.
- 231 Huo, H., Yao, Z., Zhang, Y., Shen, X., Zhang, Q. and He, K. (2012). On-board measurements of
232 emissions from diesel trucks in five cities in China. *Atmos. Environ.* 54: 159-167.
- 233 Huo, H., Zhang, Q., He, K., Yao, Z., Wang, X., Zheng, B., Streets, D.G., Wang, Q. and Ding, Y.
234 (2011). Modeling vehicle emissions in different types of Chinese cities: Importance of vehicle
235 fleet and local features. *Environ. Pollut.* 159: 2954-2960.
- 236 Khan, M.Y., Johnson, K.C., Durbin, T.D., Jung, H., Cocker, D.R., Bishnu, D. and Giannelli, R. (2012).
237 Characterization of PM-PEMS for in-use measurements conducted during validation testing
238 for the PM-PEMS measurement allowance program. *Atmos. Environ.* 55: 311-318.
- 239 Liu, J., Ge, Y., Wang, X., Hao, L., Tan, J., Peng, Z., Zhang, C., Gong, H. and Huang, Y. (2017).
240 On-board measurement of particle numbers and their size distribution from a light-duty diesel
241 vehicle: Influences of VSP and altitude. *J. Environ. Sci. China* 57: 238-248.
- 242 Liu, Z., Ge, Y., Johnson, K.C., Shah, A.N., Tan, J., Wang, C. and Yu, L. (2011). Real-world operation
243 conditions and on-road emissions of Beijing diesel buses measured by using portable emission
244 measurement system and electric low-pressure impactor. *Sci. Total Environ.* 409: 1476-1480.
- 245 López-Martínez, J.M., Jiménez, F., Javier Páez-Ayuso, F., Nuria Flores-Holgado, M., Arenas, A.N.,
246 Arenas-Ramírez, B. and Aparicio-Izquierdo, F. (2017). Modelling the fuel consumption and
247 pollutant emissions of the urban bus fleet of the city of Madrid. *Transport. Res. D: Tr. E.* 52:
248 112-127.
- 249 National Bureau of Statistic of China.(NBSC) (2015). *China statistical yearbook* (in Chinese).
- 250 National Bureau of Statistic of China.(NBSC) (2016). *China statistical yearbook* (in Chinese).
- 251 Peng, Z., Ge, Y., Tan, J., Fu, M., Wang, X., Chen, M., Yao, H. and Wu, Y. (2016). Real-world
252 emission from in-use construction equipment in China. *Aerosol Air Qual. Res.* 16: 1893-1902.
- 253 Shen, X., Yao, Z., Zhang, Q., Wagner, D.V., Huo, H., Zhang, Y., Zheng, B. and He, K. (2015).
254 Development of database of real-world diesel vehicle emission factors for China. *J. Environ.
255 Sci. China* 31: 209-220.
- 256 Tsai, Y.I., Yang, H.H., Wang, L.C., Huan, J.L., Young, L.H., Cheng, M.T. and Chiang, P.C. (2011).
257 The influences of diesel particulate filter installation on air pollutant emissions for used
258 vehicles. *Aerosol Air Qual. Res.* 11: 578-583.

- 259 Wang, G., Cheng, S., Lang, J., Li, S. and Tian, L. (2016). On-board measurements of gaseous pollutant
260 emission characteristics under real driving conditions from light-duty diesel vehicles in
261 Chinese cities. *J. Environ. Sci. China* 46: 28-37.
- 262 Wu, Y., Zhang, S., Hao, J., Liu, H., Wu, X., Hu, J., Walsh, M.P., Wallington, T.J., Zhang, K.M. and
263 Stevanovic, S. (2017). On-road vehicle emissions and their control in China: A review and
264 outlook. *Sci. Total Environ.* 574: 332-349.
- 265 Wu, Y., Zhang, S.J., Li, M.L., Ge, Y.S., Shu, J.W., Zhou, Y., Xu, Y.Y., Hu, J.N., Liu, H., Fu, L.X., He,
266 K.B. and Hao, J.M. (2012). The challenge to NO_x emission control for heavy-duty diesel
267 vehicles in China. *Atmos. Chem. Phys.* 12: 9365-9379.
- 268 Yang, C.Q., Wang, Y.M. and Wu, L.G. (2017). Influence of ZrO₂-Al₂O₃ ratio in carrier on performance
269 of Pt/ZrO₂-Al₂O₃ catalyst. *Rare Metal Mat. Eng.* 46: 2049-2054.
- 270 Yang, X.F., Liu, H., Man, H.Y. and He, K.B. (2015). Characterization of road freight transportation
271 and its impact on the national emission inventory in China. *Atmos. Chem. Phys.* 15:
272 2105-2118.
- 273 Yao, Z., Jiang, X., Shen, X., Ye, Y., Cao, X., Zhang, Y. and He, K. (2015 a). On-road emission
274 characteristics of carbonyl compounds for heavy-duty diesel trucks. *Aerosol Air Qual. Res.* 15:
275 915-925.
- 276 Yao, Z., Shen, X., Ye, Y., Cao, X., Jiang, X., Zhang, Y. and He, K. (2015 b). On-road emission
277 characteristics of VOCs from diesel trucks in Beijing, China. *Atmos. Environ.* 103: 87-93.
- 278 Yao, Z., Wu, B., Wu, Y., Cao, X. and Jiang, X. (2015c). Comparison of NO_x emissions from China III
279 and China IV in-use diesel trucks based on on-road measurements. *Atmos. Environ.* 123, Part
280 A: 1-8.
- 281 Yu, Q., Li, T.Z. and Li, H. (2016). Improving urban bus emission and fuel consumption modeling by
282 incorporating passenger load factor for real world driving. *Appl. Energ.* 161: 101-111.
- 283 Zhang, D. F., Wang, H. H., Wang, Q. Q., Li, W., Jiang, W. P., Huo, P. L., Zhang, J., Zhu, L., Duan G.
284 Q., Du, C. C. (2016 a). Interactions of nitric oxide with various rank coals: Implications for
285 oxy-coal combustion flue gas sequestration in deep coal seams with enhanced coalbed
286 methane recovery. *Fuel* 182: 704-712.
- 287 Zhang, Q., Wu, L., Yang, Z., Zou, C., Liu, X., Zhang, K. and Mao, H. (2016 b). Characteristics of
288 gaseous and particulate pollutants exhaust from logistics transportation vehicle on real-world
289 conditions. *Transport. Res. D: Tr. E.* 43: 40-48.
- 290 Zheng, X., Wu, Y., Jiang, J., Zhang, S., Liu, H., Song, S., Li, Z., Fan, X., Fu, L. and Hao, J. (2015).
291 Characteristics of on-road diesel vehicles: Black carbon emissions in Chinese cities based on
292 portable emissions measurement. *Environ. Sci. Technol.* 49: 13492-13500.
- 293

294 Table list

295

Table 1 Specifications of the tested freight tucks

Test no.	Fuel Type	Emission Standard	Total Mass (kg)	Date of Production	Power (kW)
1	Diesel	China III	8495	2010.6.10	115
2	Diesel	China III	8490	2010.4.19	103
3	Diesel	China III	8010	2009.2.23	96
4	Diesel	China III	8495	2012.9.6	115
5	Diesel	China IV	7700	2011.10.20	100
6	Diesel	China IV	4390	2013.9.24	95
7	Diesel	China IV	4490	2015.6.19	87

296

ACCEPTED MANUSCRIPT

297 Figure list
298 **Fig. 1 Schematic diagram of test system**
299 **Fig. 2 CO₂ emission rates under different VSP**
300 **Fig. 3 Characteristics of CO emission rates: (a) CO emission vs. VSP; (b) CO emission based**
301 **on CO₂ vs. Speed and (c) CO emission based on CO₂ vs. VSP.**
302 **Fig. 4 Characteristics of HC emission rates: (a) HC emission vs. VSP; (b) HC emission based**
303 **on CO₂ vs. Speed and (c) HC emission based on CO₂ vs. VSP.**
304 **Fig. 5 Characteristics of NO_x emission rates: (a) NO_x emission vs. VSP; (b) NO_x emission**
305 **based on CO₂ vs. Speed and (c) NO_x emission based on CO₂ vs. VSP.**
306 **Fig. 6 Characteristics of PM emission rates: (a) PM emission vs. VSP; (b) PM emission based**
307 **on CO₂ vs. Speed and (c) PM emission based on CO₂ vs. VSP.**

ACCEPTED MANUSCRIPT