

Source Profiles for PM_{10-2.5} Resuspended Dust and Vehicle Exhaust Emissions in Central India

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Abstract

Eight composite PM_{10-2.5} source profiles were developed for resuspended dust and vehicle exhaust emissions with 32 chemical species, including 21 elements (Al, As, Ca, Cd, Co, Cr, Cu, Fe, Hg, K, Mg, Mn, Mo, Na, Ni, Pb, S, Sb, Se, V, and Zn), 9 water-soluble ions (Na⁺, K⁺, Mg²⁺, Ca²⁺, NH₄⁺, Cl⁻, F⁻, NO₃⁻, and SO₄²⁻), and carbonaceous fractions (OC and EC). Dust samples were dominated by crustal elements (Al, Ca, Fe, and Mg) while exhaust emissions showed high abundances of carbonaceous aerosol (OC and EC). Crustal species (Al, Fe, Mg, and Na) were more enriched over native soils in PM_{10-2.5} as compared to PM_{2.5}. The higher coefficients of divergence (COD) indicate that profiles differ from each other. Ca accounted for nearly 30% of PM_{10-2.5} mass in construction dust while Fe accounted for nearly 20% of PM_{10-2.5} mass in paved road dust. Three- and four-wheeler diesel exhaust profiles consisted of 5-7% EC, with 6-10 times higher Pb, Se, and S abundances than those in two-wheeler gasoline exhaust profile. The heavy-duty diesel exhaust profile consist of nearly 20% EC with abundant (>0.5%) trace elements (e.g, Pb, Se, and Zn).

Keywords: PM_{10-2.5}; Source Profile; Enrichment Factor; Source Markers; Resuspended Dust; Vehicle Exhaust

64 1. Introduction

65 Air pollution is of great concern in India, especially the high levels of particulate matter (PM)
66 emitted from uncontrolled industrial processes, solid waste and biomass burning, vehicular
67 exhaust, and resuspended road dust (Pant *et al.*, 2015; Pant and Harrison, 2013). Real-world
68 source characterizations are needed to obtain chemical source profiles for input to receptor
69 models, such as the Chemical Mass Balance (CMB), to identify and quantify source
70 contributions. The U.S. EPA SPECIATE (U.S. EPA, 2013), European SPECIEUROPE
71 (Pernigotti *et al.*, 2016), and China Source Profile Shared Service (CSPSS) (Liu *et al.*, 2017)
72 databases have assembled many of these profiles.

73 Gargava and Rajagopalan (2016) found that road dust and vehicular exhaust emissions
74 account for ~30-70% and ~15-20% of the measured PM₁₀ mass, respectively, in India.
75 Various studies have been conducted (Chow *et al.*, 2003; Han *et al.*, 2014; Ho *et al.*, 2003;
76 Kong *et al.*, 2011, 2014; Liu *et al.*, 2016; Matawle *et al.*, 2015; Pant *et al.*, 2015; Patil *et al.*,
77 2013; Wang *et al.*, 2015) to derive dust and motor vehicle exhaust profiles (Chow *et al.*,
78 2004; Han *et al.*, 2014; Liu *et al.*, 2017; Matawle *et al.*, 2015). This study reports additional
79 PM_{10-2.5} chemical source profiles for resuspended dust and vehicle exhaust emissions specific
80 to India.

81

82 2. Methodology

83 *Source Sampling and Chemical Analysis*

84 Source sampling was conducted in Raipur, the capital of Chhattisgarh, India (21°14'22.7"N,
85 81°38.1"E), with a population of ~1.6 million (Census, 2011), as documented by Matawle *et*
86 *al.* (2014, 2015) for PM_{2.5}. This paper describes the PM_{10-2.5} chemical profiles for the eight
87 resuspended dust and vehicle exhaust emissions tests. Source samples are summarized in
88 Table 1. Geological samples typical of Central India include paved road and construction dust
89 in Raipur City, as well as unpaved surface dust and non-agricultural soils outside of Raipur
90 City. Sweeping and grab sampling methods were employed to obtain 0.5-1 kg of each dust
91 which were air dried (~25°C), sieved (Tyler 400 mesh to 38 µm in geological diameter), and
92 resuspended in a laboratory chamber through PM_{2.5} and PM₁₀ inlets at 5 L/min following
93 Chow *et al.* (1994) as applied in past studies (Chow *et al.*, 2004; Watson and Chow, 2001;
94 Watson *et al.*, 2001).

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96 Motor vehicle exhaust samples were acquired from four major vehicle categories that are
97 common in India including: two-wheeler gasoline, three- and four-wheeler diesel, and heavy-
98 duty diesel vehicles. Vehicles manufactured between 2000 and 2001 were selected for in-
99 plume sampling through collocated PM_{2.5} and PM₁₀ inlets on Minivol samplers (Airmetrics)
100 at a flow rate of 5 L/min. Vehicles were operated under steady state conditions for 30-60
101 minutes to ensure adequate deposit on quartz-fiber filters (Whatman catalog No. 1851-047)
102 for subsequent chemical analysis. Five sets of samples were collected from each source, for a
103 total of 40 samples.

104
105 Quartz-fiber filters were weighed before and after sampling with a $\pm 10 \mu\text{g}$ sensitivity digital
106 balance (Denver, Model, TB-2150) (Watson *et al.*, 2017). These samples were analysed for
107 21 elements (Al, As, Ca, Cd, Co, Cr, Cu, Fe, Hg, K, Mg, Mn, Mo, Na, Ni, Pb, S, Sb, Se, V,
108 Zn) by atomic absorption spectrophotometry; 8 cations and anions (Na^+ , K^+ , Mg^{2+} , Ca^{2+} ,
109 NH_4^+ , Cl^- , F^- , NO_3^- , and SO_4^{2-}) by ion chromatography (Chow and Watson, 2017);
110 ammonium (NH_4^+) by spectrophotometry; and organic and elemental carbon (OC and EC) by
111 thermal/optical transmittance.

112 Detailed chemical analysis and quality assurance/quality control procedures are documented
113 in Matawle *et al.* (2014, 2015). Laboratory filter blanks and field trip blanks were submitted
114 to the same chemical analysis to assess background levels. One standard sample was analysed
115 after each 10 samples to assure 80%-120% recovery. Triplicate analyses were performed for
116 each sample to achieve $\pm 10\%$ reproducibility. The limits of detections (LODs) for each
117 species were reported in Matawle *et al.* (2014).

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120 **3. Results and Discussion**

121 **3.1 PM_{10-2.5} Chemical Source Profile**

122 The four resuspended dust and four vehicle exhaust profiles are summarized in Tables 2 and
123 3, respectively. The sum of species accounted for 40-47% and 52-69% of PM_{10-2.5} mass for
124 dust and vehicle exhaust profiles, respectively. Crustal elements (Al, Ca, Fe, K, Mg, and Na)
125 were the most abundant species in dust, contributing 31-45% of the PM_{10-2.5} mass, whereas
126 total carbon (TC = OC + EC) constituted 49-57% exhaust. The OC/TC ratios ranged from
127 0.65-0.98, comparable to 0.57-0.98 reported in India for PM₁₀ (CPCB, 2008b) and PM_{2.5}
128 (Matawle *et al.*, 2015). The low sum of species for dust is mainly due to the lack of silicon
129 (Si) in the profile. Si is often the most abundant element in crustal dust (Chow *et al.*, 2003).

130 The quartz-fiber filter prohibits Si analysis and the use of Si/Al ratio as a source marker
131 (Contini *et al.*, 2016). Future studies should be conducted with parallel Teflon-membrane and
132 quartz-fiber filters to accommodate complete chemical speciation (Chow *et al.*, 1994; Watson
133 *et al.*, 2001).

134

135 **Source profile for Resuspended Dust**

136 Figure 1 shows four abundant crustal species: Ca, Fe, Mg, and Al. The most abundant species,
137 Ca, varied two-fold among the four profiles, from $27.9 \pm 7.3\%$ in construction dust (CD) to
138 $14.3 \pm 22\%$ in non-agricultural soils (SD). Ca is commonly found in construction dust (Kong
139 *et al.*, 2011; Pant and Harrison, 2012; Shen *et al.*, 2016; Yatkin and Bayram, 2008) owing to
140 its presence in concrete. Ca was not water soluble, with Ca^{2+}/Ca values in the range of 0.14-
141 0.18, with a lower ratio for construction dust (0.012). Fe was most abundant ($17.5 \pm 0.8\%$) in
142 unpaved road dust (UPRD), compared to a lower abundance in construction dust (CD, $7.1 \pm$
143 0.7%). Al levels were low (0.8-0.9%) in paved and unpaved road dust, but they were highest
144 at 2-3% in soil and construction dust. Mg levels were similar, in the range of 2-4% of $\text{PM}_{10-2.5}$
145 mass. These abundances are comparable to those from past studies for $\text{PM}_{2.5}$, $\text{PM}_{10-2.5}$, and
146 PM_{10} (Amato *et al.*, 2009; Chow and Watson, 1994; Matawle *et al.*, 2015; Patil *et al.*, 2013;
147 Samiksha *et al.*, 2017; Wang *et al.*, 2015; Watson *et al.*, 2001). As expected, most of the soil-
148 related K was not water soluble. K was 12 times higher than soluble K^+ with a K^+/K ratio of
149 0.08; higher than 0.1–0.5 reported in past PM_{10} (CPCB, 2008a; Kong *et al.*, 2014) and $\text{PM}_{2.5}$
150 (Matawle *et al.*, 2015; Watson *et al.*, 2001) studies. This is in contrast to biomass burning
151 profiles where the K^+/K ratio is in the range of $\sim 0.87\text{--}0.90$ (Chow *et al.*, 2004; Watson *et al.*,
152 2001). TC accounted for 2-7% of $\text{PM}_{10-2.5}$ mass. The OC abundance in PRD was $5.6 \pm 3.5\%$
153 compared to that in UPRD at $2.1 \pm 1.3\%$. Heavily travelled roads are subject to more vehicle
154 exhaust deposition. Pb, V, and S abundances were 3-5 times higher in PRD as compared to
155 other dusts, similar to abundances in other Indian cities for PM_{10} (CPCB, 2008a; Samara,
156 2005). OC/TC ratios ranged from 0.56 in UPRD to 0.98 in CD, consistent with 0.64–0.99
157 reported in past PM_{10} studies (Chow *et al.*, 2004; Gupta *et al.*, 2007; Ho *et al.*, 2003).

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160 Enrichment Factors (EF) were calculated relative to Ca in local soil as a reference element
161 because: (1) The study region is located in a rock basin with high Ca abundances; (2) Ca
162 correlates with other elements in the dust matrix (Quraishi, 1997); and (3) Past studies have

163 used Ca as an EF reference element (Sharma and Pervez, 2003). The EF (Behera and Sharma,
164 2010; Cao *et al.*, 2008; Chakraborty and Gupta, 2009) is:

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$$166 \quad EF = \frac{(X_i/Ca)_{\text{sample}}}{(X_i/Ca)_{\text{crust}}} \quad (1)$$

167

168 where $(X_i/Ca)_{\text{sample}}$ and $(X_i/Ca)_{\text{crust}}$ are ratios of the abundance of element X_i and Ca in
169 PM samples and in crustal materials, respectively. Figure 2 shows elemental EFs and Figure
170 3 compares EFs for $PM_{10-2.5}$ and $PM_{2.5}$ profiles (Matawle *et al.*, 2015). EFs for both size
171 fractions are detailed in Supplemental Table S1. As shown in Figure 3, Cd had high EFs for
172 all sources ranging from 9-17 for $PM_{2.5}$ and 2-8 for $PM_{10-2.5}$, comparable to other studies
173 (Han *et al.*, 2014; Kong *et al.*, 2014). For PRD and UPRD profiles, As, Cu, and Zn were
174 enriched ($EF > 5$ for $PM_{2.5}$ and $EF > 3$ for $PM_{10-2.5}$), consistent with influences from traffic
175 emissions such as tire and brake wear. Most of the crustal species (Al and Mg) were enriched
176 in $PM_{10-2.5}$ as compared to $PM_{2.5}$, while most of the anthropogenic related elements were
177 more abundant in $PM_{2.5}$.

178

179 **Source Profiles for Vehicle Exhaust Emissions**

180 The four exhaust profiles are shown in Figure 4. TC was the most abundant species
181 accounting for 49-57% of the measured mass. The OC abundance was highest for the two-
182 wheeler gasoline exhaust (2WVG, $48\% \pm 6.6\%$) whereas the EC abundance was highest for
183 the heavy-duty diesel (HDVD, $19.7\% \pm 1.5\%$). These levels were 3-26 times higher than two-
184 to four-wheeler exhaust profiles, but comparable to PM_{10} profiles reported by Han *et al.*
185 (2014). The OC/TC ratios in the range of 0.65 to 0.98 were similar to the 0.55-0.95 for the
186 PM_{10} profiles of Han *et al.* (2014) as well as 0.57-0.98 in Matawle *et al.* (2015) and 0.66-0.80
187 in Watson *et al.* (2001) for $PM_{2.5}$. The largest difference were in the OC/EC ratios, ranging
188 from 1.9 for HDVD to 63.8 for 2WVG, mainly due to the low EC levels ($0.8 \pm 0.2\%$) for
189 two-wheeler gasoline exhaust profile. Other elemental abundances were low except for Na,
190 ranging from $0.63 \pm 1.5\%$ in 3WVD to $6.7 \pm 5.2\%$ in HDVD. The heavy-duty diesel vehicle
191 profile contained the highest Pb ($0.77 \pm 0.06\%$), Se ($0.76 \pm 0.09\%$), Cl⁻ ($1.1 \pm 0.13\%$), and
192 SO_4^{2-} ($1.0 \pm 0.2\%$), abundances. Three- and four-wheeler diesel exhaust profiles consisted of
193 5-7% of EC, with 6-10 times higher Pb, Se, and S than two-wheeler gasoline vehicles.

194

195 3.2 Mass Reconstruction

196 PM_{10-2.5} mass reconstruction evaluates closure between gravimetric mass and the major
197 chemical constituents (Watson *et al.*, 2012). Figure 5 shows reconstructed PM_{10-2.5} in seven
198 categories (Chow *et al.*, 2015; Pei *et al.*, 2016): (1) geological materials derived from a
199 modified IMPROVE equation (Malm *et al.*, 1994), without the inclusion of Si and Ti, where
200 minerals = 2.2Al + 1.63Ca + 2.42Fe; (2) other elements (all elements measured excluding
201 Na, Mg, Al, S, K, Ca, and Fe); (3) sulphate (SO₄²⁻); (4) other ions (all ions measured
202 excluding SO₄²⁻ and Ca²⁺); (5) organic matter (OM=OCx1.8) to account for unmeasured
203 oxygen and hydrogen (Pitchford *et al.*, 2007); (6) EC; and (7) unidentified species calculated
204 by subtracting the sum of categories 1-6 above from 100, which include species that are not
205 measured (such as Si and Ti) or not adequately accounted for (such as oxide forms of other
206 crustal materials or variations in the OM/OC multiplier).

207 Approximately 91-93% of measured mass was achieved for exhaust profiles, with lower
208 values (65-76%) for the dust profiles, mainly due to the lack of Si and Ti measurements. As
209 expected, OM was the major fraction (72-95%) of exhaust emissions, whereas geological
210 minerals (80-94%) dominated the dust profiles.

211

212 3.3 Coefficients of Divergence

213 To evaluate the similarities and differences among the profiles, coefficients of divergence
214 (COD) were calculated, as described by Matawle *et al.* (2014, 2015). When COD values <
215 0.2, as suggested by Contini *et al.* (2012), the two sources are similar, and when the COD >
216 0.2 the two sources are considered different (Wilson *et al.*, 2005; Wongphatarakul *et al.*,
217 1998). Table 4 shows high COD values ranging from 0.48 between 3WVD and 4WVD to
218 0.84 between PRD and HDVD, indicating that the profiles are not collinear.

219

220 3.4 Implications for source apportionment

221 Source Markers

222 Potential source markers are identified by the following equation (Kong *et al.*, 2011; Yang *et*
223 *al.*, 2002):

$$224 \text{Ratio}_{j,i} = \frac{(X_i / \sum X)_j}{(X_i / \sum X)_{\text{min}}} \quad (2)$$

225 where X_i is the i^{th} species concentration; $(X_i/\sum X)_j$ is the abundance of i^{th} species divided by
226 the sum of the measured 32 species concentration ($\sum X$) for source j ; $(X_i/\sum X)_{\text{min}}$ is the
227 minimum abundance of the i^{th} individual species divided by $\sum X$ (Chen *et al.*, 2003; Yang *et al.*,
228 *et al.*, 2002). Individual species concentrations are further normalized by dividing the i^{th} species
229 concentration by the sum of the i^{th} concentrations (Kong *et al.*, 2011). Species with the six
230 highest ratios are potential source markers. Similar approaches used by other studies are
231 summarized in Table 5. Past studies (Guttikunda, 2009; Kong *et al.*, 2011; Matawle *et al.*,
232 2015; Mitra *et al.*, 2002; Viana *et al.*, 2008; Watson *et al.*, 2008) showed that Al, Si, K, Ca,
233 Mg, and Fe were commonly used as markers for dust sources, whereas OC, EC, S, or SO_4^{2-} ,
234 and Pb were markers for exhaust. As shown in Table 5, Pb and Se may be markers for paved
235 road dust in $\text{PM}_{10-2.5}$ and for unpaved road dust in $\text{PM}_{2.5}$ (Matawle *et al.*, 2015).

236

237 **Diagnostic ratios**

238 Diagnostic ratios are used to distinguish among sources (Arditsoglou and Samara, 2005;
239 Kong *et al.*, 2011; Matawle *et al.*, 2015). The V/Ni ratio was used to assess emissions from
240 marine vessels and residual oil combustion and Cu/Sb and Cu/Zn ratios were used for traffic
241 emissions (Pey *et al.*, 2010). Arditsoglou and Samara (2005) used Zn/Pb ratios in the range of
242 0.3–0.4 as to infer exhaust emissions, and 1.2 for oil combustion. Mitra *et al.* (2002)
243 suggested a Mn/V ratio $\ll 1$ for oil burning and $\gg 1$ for coal burning emissions.

244

245 Nine elemental ratios (Mn/V, Cu/Sb, As/V, V/Ni, Zn/Pb, Zn/Cd, Cu/Zn, Cu/Cd, and Cu/Pb)
246 are compared with previous studies in Table 6. The Mn/V ratios for dust profiles ranged from
247 2.98–24.8, mainly due to elevated Mn (2.1–3.7%) and low V (0.003–0.007%) abundances.
248 Mn/V ratios (0.56 to 1.86) in exhaust profiles were higher than in past studies (0.05–0.74) due
249 to lower V abundances (0.001–0.006%) in $\text{PM}_{10-2.5}$. The Cu/Sb ratios varied from low (1.08–
250 4.2) for exhaust profiles, to high (1.68–48.3) for dust profiles, similar to past PM_{10} studies
251 (CPCB, 2008a; 2008b). The V/Ni ratios (0.28–1.08) for exhaust were comparable to 0.11–
252 0.85 found in the corresponding $\text{PM}_{2.5}$ fractions (Matawle *et al.*, 2015), but five times higher
253 than in other studies (Kong *et al.*, 2011; Lee *et al.*, 2000; Moreno *et al.*, 2006; Samara *et al.*,
254 2003). High Zn/Pb and Zn/Cd ratios in dust profiles suggest a Zn enrichment due to
255 deposition of vehicle exhaust and tire/brake wear.

256 **4. Conclusion**

257 $\text{PM}_{10-2.5}$ source profiles from paved road and construction dust in Raipur, unpaved road dust
258 and non-agricultural soil outside of Raipur, along with vehicle exhaust from gasoline two-

259 wheelers, diesel three- and four-wheeler and heavy-duty diesel vehicles were acquired. In
260 addition to gravimetric mass, these samples were analysed for 21 elemental species, 9 water
261 soluble ions, and carbon (OC and EC). Crustal elements (Al, Ca, Fe and, Mg) dominated the
262 resuspended dust while carbonaceous species (OC and EC) were more abundant in vehicle
263 exhaust emissions. Ca was most abundant in construction dust ($27.9 \pm 7.3\%$ of $PM_{10-2.5}$ mass)
264 while the most abundant Fe ($17.5 \pm 0.8\%$) was found in unpaved road dust. Heavy-duty
265 diesel vehicles (HDVD) reported the highest EC abundance ($19.7 \pm 1.5\%$) with very low EC
266 ($0.75 \pm 0.21\%$) found in gasoline two wheelers (2WVG). Elevated levels of Pb ($0.77 \pm$
267 0.06%), Se ($0.76 \pm 0.09\%$), and Zn ($0.91 \pm 0.31\%$) were also apparent in HDVD. The
268 coefficients of divergence (COD) ranged 0.48 to 0.84 suggesting profiles were significantly
269 different from each other. Lower than usual mass reconstruction for resuspended dust (65-76
270 %) reconfirm the importance to include Si and Ti in future studies. Source markers were
271 identified as Al, Ca, and Fe for resuspended dust and OC, EC, and Pb for vehicle exhaust
272 emissions. These region-specific profiles are more representative of pollution source
273 characteristics and can be used for future source apportionment studies.

274

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283

284 **References**

- 285 Amato, F., Pandolfi, M., Escrig, A., Querol, X., Alastuey, A., Pey, J., Perez, N. and Hopke,
286 P.K. (2009). Quantifying road dust resuspension in urban environment by multilinear
287 engine: a comparison with PMF2. *Atmos. Environ.* 43: 2770–2780.
- 288 Arditoglou, A. and Samara, C. (2005). Levels of total suspended particulate matter and
289 major trace elements in Kosovo: a source identification and apportionment study.
290 *Chemosphere* 59: 669–678.
- 291 Behera, S.N. and Sharma, M. (2010). Reconstructing primary and secondary components of
292 $PM_{2.5}$ composition for an urban atmosphere. *Aerosol Sci. Technol.* 44: 983–992.

- 293 Cao, J.J., Chow, J.C., Watson, J.G., Wu, F., Han, Y.M., Jin, Z.D., Shen, Z.X. and An, Z.S.
294 (2008). Size-differentiated source profiles for fugitive dust in the Chinese Loess Plateau.
295 *Atmos. Environ.* 42: 2261–2275.
- 296 Census, (2011). Office of the Registrar General and Census Commissioner, India. Available
297 from: <http://www.censusindia.gov.in/2011>
- 298 Chakraborty, A. and Gupta, T. (2009). Chemical characterization of submicron aerosol in
299 Kanpur region: a source apportionment study. *Int. J. Env. Ac. Eng.* 1: 19–27.
- 300 Chen, S.J., Hsieh, L.T. and Chiu, S.C. (2003). Emission of polycyclic aromatic hydrocarbons
301 from animal carcass incinerators. *Sci. Total Environ.* 313: 61-76.
- 302 Chow, J.C., Lowenthal, D.H., Chen, L.-W.A., Wang, X. and Watson, J.G. (2015). Mass
303 reconstruction methods for PM_{2.5}: a review. *Air Qual. Atmos. Heal.* 8: 243–263.
- 304 Chow, J.C. and Watson, J.G. (2017). Enhanced ion chromatographic speciation of water-
305 soluble PM_{2.5} to improve aerosol source apportionment. *Aerosol Science and*
306 *Engineering* 1(1): 7-24.
- 307 Chow, J.C. and Watson, J.G. (1994). Contemporary source profiles for geological material
308 and motor vehicle emissions. Rep. no. DRI 2625.
- 309 Chow, J.C., Watson, J.G., Ashbaugh, L.L. and Magliano, K.L. (2003). Similarities and
310 differences in PM₁₀ chemical source profiles for geological dust from the San Joaquin
311 Valley, California. *Atmos. Environ.* 37: 1317–1340.
- 312 Chow, J.C., Watson, J.G., Houck, J.E., Pritchett, L.C., Rogers, C.F., Frazier, C.A., Egami,
313 R.T. and Ball, B.M. (1994). A Laboratory Resuspension Chamber to Measure Fugitive
314 Dust Size Distributions and Chemical Compositions. *Atmos. Environ.* 28: 3463–3481.
- 315 Chow, J.C., Watson, J.G., Kuhns, H., Etyemezian, V., Lowenthal, D.H., Crow, D., Kohl, S.D.,
316 Engelbrecht, J.P. and Green, M.C. (2004). Source profiles for industrial, mobile, and
317 area sources in the Big Bend Regional Aerosol Visibility and Observational study.
318 *Chemosphere* 54: 185–208.
- 319 Contini, D., Belosi, F., Gambaro, A., Cesari, D., Stortini, A.M. and Bove, M.C. (2012).
320 Comparison of PM₁₀ concentrations and metal content in three different sites of the
321 Venice Lagoon: an analysis of possible aerosol sources. *J. Environ. Sci.* 24: 1954–1965.
- 322 Contini, D., Cesari, D., Conte, M. and Donato, A. (2016). Application of PMF and CMB
323 receptor models for the evaluation of the contribution of a large coal-fired power plant to
324 PM₁₀ concentrations. *Sci. Total Environ.* 560: 131-140.
- 325 CPCB (2008a). Stationary Sources Emission Profiles, Central Pollution Control Board, Delhi.
326 Available from http://www.cpcb.nic.in/Stationary_Sources_Emission_Profiles.xls.
- 327 CPCB (2008b). Vehicular Sources Emission Profiles. Central Pollution Control Board, New
328 Delhi, India. http://cpcb.nic.in/Vehicular_Sources_Emission_Profiles.xls, Last Access:
329 09.10.2014.
- 330 Gargava, P. and Rajagopalan, V. (2016). Source apportionment studies in six Indian cities—
331 drawing broad inferences for urban PM₁₀ reductions. *Air Qual. Atmos. Heal.* 5: 471–481.

- 332 Gupta, A.K., Karar, K. and Srivastava, A. (2007). Chemical mass balance source
333 apportionment of PM₁₀ and TSP in residential and industrial sites of an urban region of
334 Kolkata, India. *J. Hazard. Mater.* 142: 279–287.
- 335 Guttikunda, S.K. (2009). Urban Particulate Pollution Source Apportionment: Part 1—
336 Definition, Methodology, and Resources. *Simple Interactive Models for Better Air*
337 *Quality: SIM-air Working Paper Series*, 16.
- 338 Han, J., Han, B., Li, P., Kong, S., Bai, Z., Han, D., Dou, X. and Zhao, X. (2014). Chemical
339 characterizations of PM₁₀ profiles for major emission Sources in Xining, Northwestern
340 China. *Aerosol Air Qual. Res.* 14: 1017–1027.
- 341 Han, S., Youn, J.S. and Jung, Y.W. (2011). Characterization of PM₁₀ and PM_{2.5} Source
342 Profiles for Resuspended Road Dust Collected Using Mobile Sampling Methodology.
343 *Atmos. Environ.* 45: 3343–3351
- 344 Ho, K.F., Lee, S.C., Chow, J.C. and Watson, J.G. (2003). Characterization of PM₁₀ and PM_{2.5}
345 source profiles for fugitive dust in Hong Kong. *Atmos. Environ.* 37: 1023–1032.
- 346 Kong, S., Ji, Y., Lu, B., Chen, L., Han, B., Li, Z. and Bai, Z. (2011). Characterization of
347 PM₁₀ source profiles for fugitive dust in Fushun—a city famous for coal. *Atmos. Environ.*
348 45: 5351–5365.
- 349 Kong, S., Ji, Y., Lu, B., Zhao, X., Han, B. and Bai, Z. (2014). Similarities and differences in
350 PM_{2.5}, PM₁₀ and TSP chemical profiles of fugitive dust sources in a coastal oilfield city
351 in China. *Aerosol Air Qual. Res.* 14: 2017–2028.
- 352 Lee, S.W., Pomalis, R. and Kan, B. (2000). A new methodology for source characterization
353 of oil combustion particulate matter. *Fuel Process. Technol.* 65: 189–202.
- 354 Liu, Y., Zhang, W., Bai, Z., Yang, W., Zhao, X., Han, B. and Wang, X. (2017). China Source
355 Profile Shared Service (CSPSS): The Chinese PM_{2.5} Database for Source Profiles.
356 *Aerosol Air Qual. Res.* 17: 1401–1414.
- 357 Liu, Y., Zhang, W., Bai, Z., Yang, W., Zhao, X., Han, B. and Wang, X. (2016).
358 Characteristics of PM₁₀ chemical source profiles for geological dust from the south-west
359 region of China. *Atmosphere (Basel)*. 7: 146.
- 360 Malm, W.C., Sisler, J.F., Huffman, D., Eldred, R.A. and Cahill, T.A. (1994). Spatial and
361 seasonal trends in particle concentration and optical extinction in the United States. *J.*
362 *Geophys. Res. Atmos.* 99: 1347–1370.
- 363 Matawle, J.L., Pervez, S., Dewangan, S., Shrivastava, A., Tiwari, S., Pant, P., Deb, M.K. and
364 Pervez, Y. (2015). Characterization of PM_{2.5} source profiles for traffic and dust sources
365 in Raipur, India. *Aerosol Air Qual. Res.* 15: 2537–2548.
- 366 Matawle, J., Pervez, S., Dewangan, S., Tiwari, S., Bisht, D.S. and Pervez, Y.F. (2014). PM_{2.5}
367 chemical source profiles of emissions resulting from industrial and domestic burning
368 activities in India. *Aerosol Air Qual. Res.* 14: 2051–2066.
- 369 Mitra, A.P., Morawska, L., Sharma, C. and Zhang, J. (2002). Chapter two: methodologies for
370 characterisation of combustion sources and for quantification of their emissions.
371 *Chemosphere* 49: 903–922.

- 372 Moreno, T., Querol, X., Castillo, S., Alastuey, A., Cuevas, E., Herrmann, L., Mounkaila, M.,
373 Elvira, J. and Gibbons, W. (2006). Geochemical variations in aeolian mineral particles
374 from the Sahara–Sahel dust corridor. *Chemosphere* 65: 261–270.
- 375 Pant, P., Baker, S.J., Shukla, A., Maikawa, C., Pollitt, K.J.G. and Harrison, R.M. (2015). The
376 PM₁₀ fraction of road dust in the UK and India: Characterization, source profiles and
377 oxidative potential. *Sci. Total Environ.* 530: 445–452.
- 378 Pant, P. and Harrison, R.M. (2013). Estimation of the contribution of road traffic emissions to
379 particulate matter concentrations from field measurements: A review. *Atmos. Environ.*
380 77: 78–97.
- 381 Pant, P. and Harrison, R.M. (2012). Critical review of receptor modelling for particulate
382 matter: A case study of India. *Atmos. Environ.* 49: 1–12.
- 383 Patil, R.S., Kumar, R., Menon, R., Shah, M.K. and Sethi, V. (2013). Development of
384 particulate matter speciation profiles for major sources in six cities in India. *Atmos. Res.*
385 132: 1–11.
- 386 Pei, B., Wang, X., Zhang, Y., Hu, M., Sun, Y., Deng, J., Dong, L., Fu, Q. and Yan, N. (2016).
387 Emissions and source profiles of PM_{2.5} for coal-fired boilers in the Shanghai megacity,
388 China. *Atmos. Pollut. Res.* 7: 577–584.
- 389 Pernigotti, D., Belis, C.A. and Spanò, L. (2016). SPECIEUROPE: The European data base
390 for PM source profiles. *Atmos. Pollut. Res.* 7: 307–314.
- 391 Pey, J., Querol, X. and Alastuey, A. (2010). Discriminating the regional and urban
392 contributions in the North-Western Mediterranean: PM levels and composition. *Atmos.*
393 *Environ.* 44: 1587–1596.
- 394 Pitchford, M., Malm, W., Schichtel, B., Kumar, N., Lowenthal, D. and Hand, J. (2007).
395 Revised algorithm for estimating light extinction from IMPROVE particle speciation
396 data. *J. Air Waste Manage. Assoc.* 57(11): 1326–1336.
- 397 Quraishi, Y.F. (1997). Study of physico-chemical characteristics of fugitive dusts in relation
398 to respiratory ailments. Ph. D. Thesis, Pt. Ravishankar Shukla University, Raipur, India.
- 399 Samara, C. (2005). Chemical mass balance source apportionment of TSP in a lignite-burning
400 area of Western Macedonia, Greece. *Atmos. Environ.* 39: 6430–6443.
- 401 Samara, C., Kouimtzis, T., Tsitouridou, R., Kaniyas, G. and Simeonov, V. (2003). Chemical
402 mass balance source apportionment of PM₁₀ in an industrialized urban area of Northern
403 Greece. *Atmos. Environ.* 37: 41–54.
- 404 Samiksha, S., Raman, R.S., Nirmalkar, J., Kumar, S. and Sirvaiya, R. (2017). PM₁₀ and PM_{2.5}
405 chemical source profiles with optical attenuation and health risk indicators of paved and
406 unpaved road dust in Bhopal, India. *Environ. Pollut.* 1–9.
- 407 Sharma, R. and Pervez, S. (2003). Enrichment and exposure of particulate lead in a traffic
408 environment in India. *Environ. Geochem. Health* 25: 297–306.
- 409 Shen, Z., Sun, J., Cao, J., Zhang, L., Zhang, Q., Lei, Y., Gao, J., Huang, R.-J., Liu, S. and
410 Huang, Y. (2016). Chemical profiles of urban fugitive dust PM_{2.5} samples in Northern
411 Chinese cities. *Sci. Total Environ.* 569: 619–626.

- 412 USEPA (2013). PM Composite Profiles by Source Category. Available from:
413 http://cfpub.epa.gov/si/speciate/ehpa_speciate_browse.cfm.
- 414 Viana, M., Kuhlbusch, T.A.J., Querol, X., Alastuey, A., Harrison, R.M., Hopke, P.K.,
415 Winiwarter, W., Vallius, M., Szidat, S. and Prévôt, A.S.H. (2008). Source
416 apportionment of particulate matter in Europe: a review of methods and results. *J.*
417 *Aerosol Sci.* 39: 827–849.
- 418 Wang, X., Chow, J.C., Kohl, S.D., Percy, K.E., Legge, A.H. and Watson, J.G. (2015).
419 Characterization of PM_{2.5} and PM₁₀ fugitive dust source profiles in the Athabasca Oil
420 Sands Region. *J. Air Waste Manage. Assoc.* 65: 1421–1433.
- 421 Watson, J.G., Antony Chen, L.-W., Chow, J.C., Doraiswamy, P. and Lowenthal, D.H. (2008).
422 Source apportionment: findings from the US supersites program. *J. Air Waste Manage.*
423 *Assoc.* 58: 265–288.
- 424 Watson, J.G. and Chow, J.C. (2001). Source Characterization of Major Emission Sources in
425 the Imperial and Mexicali Valleys along the US/Mexico Border. *Sci. Total Environ.* 276:
426 33–47.
- 427 Watson, J.G., Chow, J.C. and Houck, J.E. (2001). PM_{2.5} chemical source profiles for vehicle
428 exhaust, vegetative burning, geological material, and coal burning in Northwestern
429 Colorado during 1995. *Chemosphere* 43: 1141–1151.
- 430 Watson, J.G., Chow, J.C., Lowenthal, D.H., Chen, L.-W.A., Wang, X. and Biscay, P. (2012).
431 Reformulation of PM_{2.5} Mass Reconstruction Assumptions for the San Joaquin Valley
432 Final Report.
- 433 Watson, J.G., Tropp, R.J., Kohl, S.D., Wang, X. and Chow, J.C. (2017). Filter processing and
434 gravimetric analysis for suspended particulate matter samples. *Aerosol Science and*
435 *Engineering* 1(2): 93-105.
- 436 Weckwerth, G. (2001). Verification of Traffic Emitted Aerosol Components in the Ambient
437 Air of Cologne (Germany). *Atmos. Environ.* 35: 5525–5536.
- 438 Wilson, J. G., Kingham, S., Pearce, J. and Sturman, A. P. (2005). A review of intraurban
439 variations in particle air pollution: implications for epidemiological research. *Atmos.*
440 *Environ.* 39: 6444–6462.
- 441 Wongphatarakul, V., Friedlander, S.K. and Pinto, J.P. (1998). A comparative study of PM_{2.5}
442 ambient aerosol chemical databases. *Environ. Sci. Technol.* 32: 3926–3934.
- 443 Yang, H.H., Lai, S.O., Hsieh, L.T., Hsueh, H.J. and Chi, T.W. (2002). Profiles of PAH
444 Emission from Steel and Iron Industries. *Chemosphere* 48: 1061–1074.
- 445 Yatkin, S. and Bayram, A. (2008). Determination of major natural and anthropogenic source
446 profiles for particulate matter and trace elements in Izmir, Turkey. *Chemosphere* 71:
447 685–696.

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450 **Supplementary materials**

451 **Table S1. Enrichment Factors of different elements in $PM_{2.5}$ and $PM_{10-2.5}$ for**
452 **resuspended dust**

453

454 **Tables Caption:**

455 **Table 1. Descriptions of source type, sampling location, and source sampling method**

456 **Table 2. $PM_{10-2.5}$ composite sources profiles (weight percent by mass) for resuspended**
457 **dust inside and outside of Raipur City**

458 **Table 3. $PM_{10-2.5}$ composite sources profiles (weight percent by mass) for vehicle exhaust**
459 **emissions**

460 **Table 4. Coefficients of Divergence (COD) for resuspended dust and vehicle exhaust**
461 **emissions**

462 **Table 5. Source markers of $PM_{10-2.5}$ for resuspended dust and vehicle exhaust emissions**

463 **Table 6. Comparison of diagnostic ratios for different source profiles**

464

465 **Figures Caption:**

466 **Figure 1. $PM_{10-2.5}$ source profiles for non-agriculture soil dust (SD), construction dust**
467 **(CD), paved road dust (PRD), and unpaved road dust (UPRD)**

468 **Figure 2. Enrichment Factors (EF) for elements in $PM_{10-2.5}$ resuspended dust. Ca was**
469 **used as the reference element.**

470 **Figure 3. Enrichment Factors for $PM_{10-2.5}$ and $PM_{2.5}$ resuspended dust**

471 **Figure 4. $PM_{10-2.5}$ vehicle exhaust profiles for gasoline two-wheeler (2WVG), diesel**
472 **three- and four-wheelers (3WVD and 4WVD), and heavy-duty diesel (HDVD) vehicles.**

473 **Figure 5. $PM_{10-2.5}$ Mass reconstruction for vehicle and resuspended dust sources (See**
474 **Table 1 for profile Mnemonics).**

Table 1. Descriptions of source type, sampling location, and source sampling method

Profile Mnemonic	Source type	Description/Location^a	Source sampling method
SD	Natural Soil Dust	Non-agricultural soil outside the city of Raipur	Chamber resuspension sampling
CD	Civil Construction Dust	Dust samples from a construction site located in the study area	Chamber resuspension sampling
PRD	Paved Road Dust	Dust samples from the surface of paved road of the study area	Chamber resuspension sampling
UPRD	Unpaved Road Dust	Dust samples from the surface of unpaved road outside the city of Raipur	Chamber resuspension sampling
2WVG	Two-Wheeler Vehicles (gasoline)	Samples from exhaust pipes of petrol driven 2-wheelers	In-plume sampling
3WVD	Three-Wheeler Vehicles (diesel)	Samples from exhaust pipes of diesel driven 3-wheelers passenger auto rickshaws	In-plume sampling
4WVD	Four-Wheeler Vehicles (diesel)	Samples from exhaust pipes of diesel driven 4-wheelers personal cars	In-plume sampling
HDVD	Heavy Duty Vehicles	Samples from exhaust pipes of diesel driven heavy duty trucks	In-plume sampling

^a Five samples were collected and composited to develop each source profile

Table 2. PM_{10-2.5} composite sources profiles (weight percent by mass) for resuspended dust inside and outside of Raipur City

Species	Profile Mnemonic ^a			
	SD	CD	PRD	UPRD
Al	3.374±0.451	2.346±0.871	0.844±0.119	0.906±0.144
As	0.000±0.000	0.000±0.000	0.006±0.005	0.003±0.004
Ca	14.331±2.187	27.859±7.313	18.573±2.607	15.049±3.569
Cd	0.000±0.000	0.001±0.001	0.003±0.003	0.001±0.001
Co	0.001±0.001	0.001±0.001	0.001±0.001	0.001±0.001
Cr	0.040±0.007	0.002±0.001	0.009±0.002	0.003±0.002
Cu	0.103±0.030	0.016±0.007	0.021±0.003	0.013±0.003
Fe	9.014±0.504	7.053±0.705	11.291±0.658	17.457±0.811
Hg	0.000±0.001	0.001±0.003	0.007±0.007	0.008±0.013
K	0.613±0.036	0.151±0.040	0.431±0.067	0.435±0.060
Mg	2.279±0.446	3.715±0.434	2.125±0.151	3.016±0.214
Mn	0.026±0.013	0.152±0.014	0.069±0.012	0.088±0.014
Mo	0.000±0.000	0.002±0.004	0.003±0.001	0.002±0.001
Na	1.726±0.097	3.508±0.209	1.024±0.083	0.649±0.037
Ni	0.015±0.012	0.007±0.005	0.017±0.009	0.007±0.009
Pb	0.001±0.001	0.001±0.000	0.016±0.009	0.004±0.005
S	0.007±0.003	0.004±0.992	0.037±0.730	0.008±0.083
Sb	0.002±0.001	0.001±0.000	0.009±0.005	0.008±0.004
Se	0.000±0.000	0.000±0.000	0.009±0.006	0.013±0.004
V	0.003±0.001	0.006±0.003	0.023±0.004	0.007±0.002
Zn	0.042±0.055	0.134±0.115	0.069±0.005	0.006±0.002
F ⁻	0.005±0.005	0.001±0.001	0.488±0.036	0.026±0.013
Cl ⁻	0.961±0.118	0.027±0.008	0.089±0.064	0.099±0.050
NO ₃ ⁻	0.649±0.198	0.015±0.008	0.019±0.014	0.025±0.011
SO ₄ ²⁻	0.847±0.135	0.005±0.005	0.988±0.076	0.324±0.033
Na ⁺	0.417±0.083	0.001±0.000	0.512±0.152	0.077±0.016
NH ₄ ⁺	0.073±0.026	0.014±0.009	0.089±0.064	0.069±0.049
K ⁺	0.051±0.031	0.010±0.002	0.061±0.059	0.082±0.089
Ca ²⁺	1.943±0.199	0.362±0.095	3.435±0.413	2.416±0.394
Mg ²⁺	0.027±0.002	0.108±0.014	0.532±0.058	0.975±0.046
OC	4.257±8.175	2.214±2.156	5.568±3.528	2.111±1.255
EC	1.908±0.873	0.056±3.049	1.636±0.818	1.629±0.515
TC	6.165±9.048	2.270±5.205	7.204±4.346	3.741±1.770
OC/EC	2.23	39.75	3.40	1.30
OC/TC	0.69	0.98	0.77	0.56
SUM%	40.278±3.376	47.291±4.962	43.467±4.908	41.972±3.112

^a See profile description in Table 1

Table 3. PM_{10-2.5} composite sources profiles (weight percent by mass) for vehicle exhaust emissions

Species	Profile Mnemonic ^a			
	2WVG	3WVD	4WVD	HDVD
Al	0.004±0.009	0.007±0.005	0.216±0.175	0.103±0.335
As	0.001±0.000	0.000±0.000	0.000±0.000	0.001±0.001
Ca	0.376±0.483	0.498±0.719	0.075±0.159	0.133±0.168
Cd	0.000±0.000	0.000±0.000	0.001±0.001	0.000±0.000
Co	0.000±0.000	0.000±0.000	0.000±0.000	0.000±0.000
Cr	0.003±0.002	0.001±0.001	0.001±0.001	0.001±0.001
Cu	0.059±0.018	0.017±0.013	0.024±0.011	0.019±0.021
Fe	0.477±0.527	0.749±0.627	0.548±0.306	0.286±0.304
Hg	0.004±0.008	0.001±0.002	0.001±0.001	0.002±0.006
K	0.002±0.002	0.014±0.018	0.014±0.003	0.002±0.003
Mg	0.019±0.018	0.046±0.025	0.046±0.027	0.005±0.003
Mn	0.006±0.005	0.002±0.004	0.003±0.003	0.002±0.002
Mo	0.003±0.003	0.001±0.003	0.001±0.002	0.002±0.002
Na	1.249±1.537	0.629±1.523	1.647±1.694	6.655±5.157
Ni	0.006±0.007	0.002±0.001	0.005±0.004	0.005±0.005
Pb	0.042±0.002	0.310±0.028	0.473±0.029	0.774±0.063
S	0.050±0.293	0.547±0.082	0.579±0.047	0.528±0.139
Sb	0.014±0.001	0.014±0.002	0.013±0.003	0.018±0.002
Se	0.040±0.005	0.260±0.056	0.058±0.014	0.757±0.085
V	0.007±0.004	0.001±0.001	0.001±0.001	0.004±0.003
Zn	0.399±0.346	0.601±0.541	0.567±0.307	0.906±0.314
F ⁻	0.083±0.018	0.000±0.000	0.000±0.000	0.000±0.000
Cl ⁻	0.135±0.121	0.563±0.243	0.062±0.034	1.100±0.129
NO ₃ ⁻	0.002±0.002	0.004±0.005	0.000±0.000	0.001±0.000
SO ₄ ²⁻	0.118±0.096	0.805±0.107	0.876±0.105	1.002±0.173
Na ⁺	0.694±0.282	0.486±0.276	0.307±0.664	0.828±0.482
NH ₄ ⁺	0.068±0.019	0.178±0.062	0.069±0.029	0.209±0.011
K ⁺	0.001±0.001	0.003±0.001	0.005±0.001	0.002±0.001
Ca ²⁺	0.116±0.027	0.316±0.211	0.058±0.019	0.061±0.031
Mg ²⁺	0.005±0.006	0.031±0.036	0.029±0.008	0.003±0.004
OC	48.103±6.589	45.205±2.647	44.537±3.753	37.167±1.887
EC	0.754±0.210	4.621±1.216	7.406±0.497	19.711±1.524
TC	48.857±6.800	49.826±3.864	51.943±4.250	56.878±3.411
OC/EC	63.77	9.78	6.01	1.89
OC/TC	0.98	0.91	0.86	0.65
SUM%	52.026±7.329	55.076±5.935	57.213±7.208	69.396±8.339

^a See profile description in Table 1

Table 4. Coefficients of Divergence (COD) for resuspended dust and vehicle exhaust emissions

Profile Mnemonic^a	SD	CD	PRD	UPRD	2WVG	3WVD	4WVD	HDVD
SD	0.00							
CD	0.64	0.00						
PRD	0.64	0.68	0.00					
UPRD	0.56	0.68	0.54	0.00				
2WVG	0.78	0.78	0.76	0.76	0.00			
3WVD	0.73	0.74	0.79	0.73	0.57	0.00		
4WVD	0.75	0.73	0.82	0.75	0.57	0.48	0.00	
HDVD	0.79	0.80	0.84	0.78	0.55	0.53	0.49	0.00

^aSee profile description in Table 1

Table 5. Source markers of PM_{10-2.5} for resuspended dust and vehicle exhaust emissions

Aerosol fraction	Source Signatures	References
Resuspended Dust Sources		
1. Soil Dust (SD)		
PM _{10-2.5}	Al, K, Fe, Ca, NO₃⁻, and SO₄²⁻	Present study
PM _{2.5}	Na ⁺ , SO ₄ ²⁻ , Zn, Se, K ⁺ , and Cl ⁻	Matawle <i>et al.</i> , 2015
PM ₁₀	S, NO ₃ ⁻ , NH ₄ ⁺ , Zn, Ni, and K ⁺	Kong <i>et al.</i> , 2011
PM	Al, Si, Sc, Ti, Fe, Sm, and Ca	Guttikunda, 2009
2. Construction Dust (CD)		
PM _{10-2.5}	Al, Ca, Mg, NO₃⁻, K, and Mg²⁺	Present study
PM _{2.5}	Zn, Na, Mo, Al, Mg ²⁺ , and Ca	Matawle <i>et al.</i> , 2015
PM ₁₀	Zn, Mg, V, Mg ²⁺ , As, and NO ₃ ⁻	Kong <i>et al.</i> , 2011
PM _{2.5}	Al, Si, K, Ca, and Fe	Watson <i>et al.</i> , 2008
3. Paved Road Dust (PRD)		
PM _{10-2.5}	Pb, Mg, Se, NO₃⁻, Ca, and K	Present study
PM _{2.5}	Na ⁺ , SO ₄ ²⁻ , As, F ⁻ , Mg ²⁺ , and Se	Matawle <i>et al.</i> , 2015
PM ₁₀	S, Zn, NO ₃ ⁻ , Cl ⁻ , Mg ²⁺ , and NH ₄ ⁺	Kong <i>et al.</i> , 2011
PM	Ca, Al, Sc, Si, Ti, Fe, and Sm	Guttikunda, 2009
PM _{2.5}	Al, Si, K, Ca, and Fe	Watson <i>et al.</i> , 2008
4. Unpaved Road Dust (UPRD)		
PM _{10-2.5}	Mg, Mg²⁺, NO₃⁻, K, Al, and Fe	Present study
PM _{2.5}	Na ⁺ , SO ₄ ²⁻ , F ⁻ , Mg ²⁺ , Se, and Pb	Matawle <i>et al.</i> , 2015
PM	Ca, Al, Sc, Si, Ti, Fe, and Sm	Guttikunda, 2009
PM _{2.5}	Al, Si, K, Ca, and Fe	Watson <i>et al.</i> , 2008
Vehicle Exhaust Emissions		
5. Two-Wheeler Vehicles (gasoline) (2WVG)		
PM _{10-2.5}	OC, EC, S, NO₃⁻, Cu, and V	Present study
PM _{2.5}	F ⁻ , Cr, Cd, V, Na ⁺ , and Ni	Matawle <i>et al.</i> , 2015
PM	EC, Br, Ce, La, Pt, SO ₄ ²⁻ , and NO ₃ ⁻	Guttikunda, 2009
PM ₁₀	Carbon, Fe, Ba, Zn, Cu, and Pb	Vianna <i>et al.</i> , 2008
PM _{2.5}	OC, EC, NH ₃ , S, Fe, and Zn	Watson <i>et al.</i> , 2008
PM	Br, Pb, and Ba	Mitra <i>et al.</i> , 2002
6. Three-Wheeler Vehicles (diesel) (3WVD)		
PM _{10-2.5}	Pb, S, EC, SO₄²⁻, OC, and NH₄⁺	Present study
PM _{2.5}	Ca ²⁺ , Mg ²⁺ , NH ₄ ⁺ , K, Se, and SO ₄ ²⁻	Matawle <i>et al.</i> , 2015
PM	OC, EC, S, SO ₄ ²⁻ , and NO ₃ ⁻	Guttikunda, 2009
PM ₁₀	Carbon, Fe, Ba, Zn, Cu, and Pb	Vianna <i>et al.</i> , 2008
PM _{2.5}	OC, EC, NH ₃ , S, Fe, and Zn	Watson <i>et al.</i> , 2008
PM	Br, Pb, and Ba	Mitra <i>et al.</i> , 2002
7. Four-Wheeler Vehicles (diesel) (4WVD)		
PM _{10-2.5}	S, EC, OC, SO₄²⁻, Pb, and Zn	Present study
PM _{2.5}	F ⁻ , NO ₃ ⁻ , Cd, Pb, SO ₄ ²⁻ , and EC	Matawle <i>et al.</i> , 2015
PM	OC, EC, S, SO ₄ ²⁻ , and NO ₃ ⁻	Guttikunda, 2009
PM ₁₀	Carbon, Fe, Ba, Zn, Cu, and Pb	Vianna <i>et al.</i> , 2008
PM _{2.5}	OC, EC, NH ₃ , S, Fe, and Zn	Watson <i>et al.</i> , 2008
PM	Br, Pb, and Ba	Mitra <i>et al.</i> , 2002
8. Heavy Duty Vehicles (diesel) (HDVD)		
PM _{10-2.5}	EC, S, SO₄²⁻, OC, NH₄⁺, and Se	Present study
PM _{2.5}	F ⁻ , NH ₄ ⁺ , Se, Pb, SO ₄ ²⁻ , and EC	Matawle <i>et al.</i> , 2015
PM	OC, EC, S, SO ₄ ²⁻ , and NO ₃ ⁻	Guttikunda, 2009
PM ₁₀	Carbon, Fe, Ba, Zn, Cu, and Pb	Vianna <i>et al.</i> , 2008
PM _{2.5}	OC, EC, NH ₃ , S, Fe, and Zn	Watson <i>et al.</i> , 2008
PM	Br, Pb, and Ba	Mitra <i>et al.</i> , 2002

Table 6. Comparison of diagnostic ratios for different source profiles

Source Types	Diagnostic Ratio								
	Mn/V	Cu/Sb	As/V	V/Ni	Zn/Pb	Zn/Cd	Cu/Zn	Cu/Cd	Cu/Pb
This study (PM_{10-2.5})									
SD	8.40	48.25	0.01	0.21	84.25	384.33	2.43	934.26	204.80
CD	24.84	24.85	0.01	0.81	356.69	434.28	0.12	51.64	42.42
PRD	2.98	2.28	0.26	1.39	4.39	26.54	0.29	7.83	1.30
UPRD	12.66	1.68	0.50	0.99	1.38	15.92	2.16	34.40	2.99
2WVG	0.82	4.19	0.04	1.08	9.42	3380.18	0.15	503.50	1.40
3WVD	1.26	1.16	0.12	0.80	1.94	10011.33	0.03	278.90	0.05
4WVD	1.86	1.82	0.10	0.28	1.20	2407.64	0.04	101.22	0.05
HDVD	0.56	1.08	0.11	0.80	1.17	53312.94	0.02	1121.53	0.02
Compiled from National studies									
Matawle <i>et al.</i>, 2015 (PM_{2.5})									
Soil	56.51	2.31	0.11	0.93	310.81	3418.93	0.15	524.53	47.68
CD	149.46	1.06	0.07	0.78	1537.17	1566.73	0.03	45.69	44.83
PRD	56.42	1.93	5.48	0.33	3.98	95.58	0.04	4.21	0.18
UPRD	19.75	1.01	1.19	0.82	114.22	670.45	0.02	16.16	2.75
2WVG	0.53	60.51	0.03	0.11	10.04	5127.08	0.36	1865.75	3.65
3WVD	0.15	35.16	0.01	0.73	0.67	3171.63	0.1	320.88	0.07
4WVD	0.08	2.02	0.01	0.62	0.16	326.46	0.29	95.5	0.05
HDVD	0.18	1.34	0.03	0.85	0.01	393.0	1.84	723.5	0.02
Other Studies									
PM₁₀ Size Fractions									
Soil ^h	7.17-114.57	0.62-7.96	0.51-3.82	0.12-1.17	0.74-3.66	7.42-69.18	0.08-0.38	2.85-24.39	0.26-0.95
Paved Road Dust ^h	7.32-141.17	0.73-21.76	0.25-3.91	0.04-1.38	0.97-4.94	18.94-144.83	0.05-0.60	7.56-70.36	0.25-2.55
Unpaved Road Dust ^h	8.49-80.38	0.48-26.70	0.43-5.10	0.19-0.49	1.81-4.18	10.97-67.76	0.08-1.34	2.69-32.32	0.19-2.42
Construction ^h	15.22	0.33	1.95	0.41	4.10	115.29	0.02	2.89	0.10
PM_{2.5} Size Fractions									
¹ (Comp-2S2WG-all) ⁱ	0.28	0.07	0.00	-	7.81	-	0.02	-	0.17
² (Comp-3WD-2) ⁱ	0.74	0.13	-	3.82	-	-	-	-	0.96
³ (Comp-LCVD-all) ⁱ	0.05	1.82	0.08	9.11	108.84	26.65	0.01	0.14	0.59
⁴ (Comp-HCVD-all) ⁱ	-	0.33	-	-	6.70	2.77	0.09	0.25	0.61
Compiled from International studies									
PM_{10-2.5} Size Fractions									
Road Dust ^f	-	8.54	-	-	-	-	-	-	-
Road Dust ^g	4.87	4.32	0.10	0.94	3.32	227.16	0.20	34.00	0.66
Soil ^e	6.92	1.86	0.08	6.77	2.30	44.22	0.25	7.84	0.50
PM₁₀ Size Fractions									
Soil ^b	62.40	-	-	0.77	9.10	344.20	-	-	-
Gasoline vehicles ^c	-	315.00	1.10	0.02	3.40	56.00	-	-	-
Diesel vehicles ^c	-	700.00	0.007	0.15	7.60	407.00	-	-	-
Cement Plant ^b	27.10	-	-	-	21.90	74.50	-	-	-
Cement Plant ^c	-	7.40	0.03	11.00	42.0	195.00	-	-	-
Oil Burning ^c	-	71.00	0.02	4.00	1.20	190.00	-	-	-
Road dust ^{b,i}	50.60	-	-	0.60	8.50	200.70	-	-	-
Construction dust ^b	37.80	-	-	0.57	11.30	68.10	-	-	-
PM_{2.5} Size Fraction									
Coal Combustion ^a	-	0.50	4.80	0.70	1.90	17.00	-	-	-
Soil ^a	6.30	0.30	0.10	8.30	3.00	9.41	0.30	2.80	0.90
Gasoline+Diesel ^a	-	0.40	-	-	1.70	0.90	0.30	0.30	0.60
Oil burning ^d	-	-	-	2.00	-	-	-	-	-
Traffic ^c	-	3.00-5.00	-	-	-	-	0.10-1.80	200-600	1.20-3.50

^aWatson *et al.*, 2001; ^bKong *et al.*, 2011; ^cSamara *et al.*, 2003; ^dLee *et al.*, 2000; ^eWeckwerth, 2001; ^fHan *et al.*, 2011; ^gChow *et al.*, 2004; ^hCPCB, 2008a; ⁱCPCB, 2008b

¹ 2-wheeler vehicle-gasoline based (Composite)

² 3-wheeler vehicle-diesel based (Composite)

³ 4-wheeler vehicle-diesel based (Composite)

⁴ Heavy-duty vehicle-diesel based (Composite)

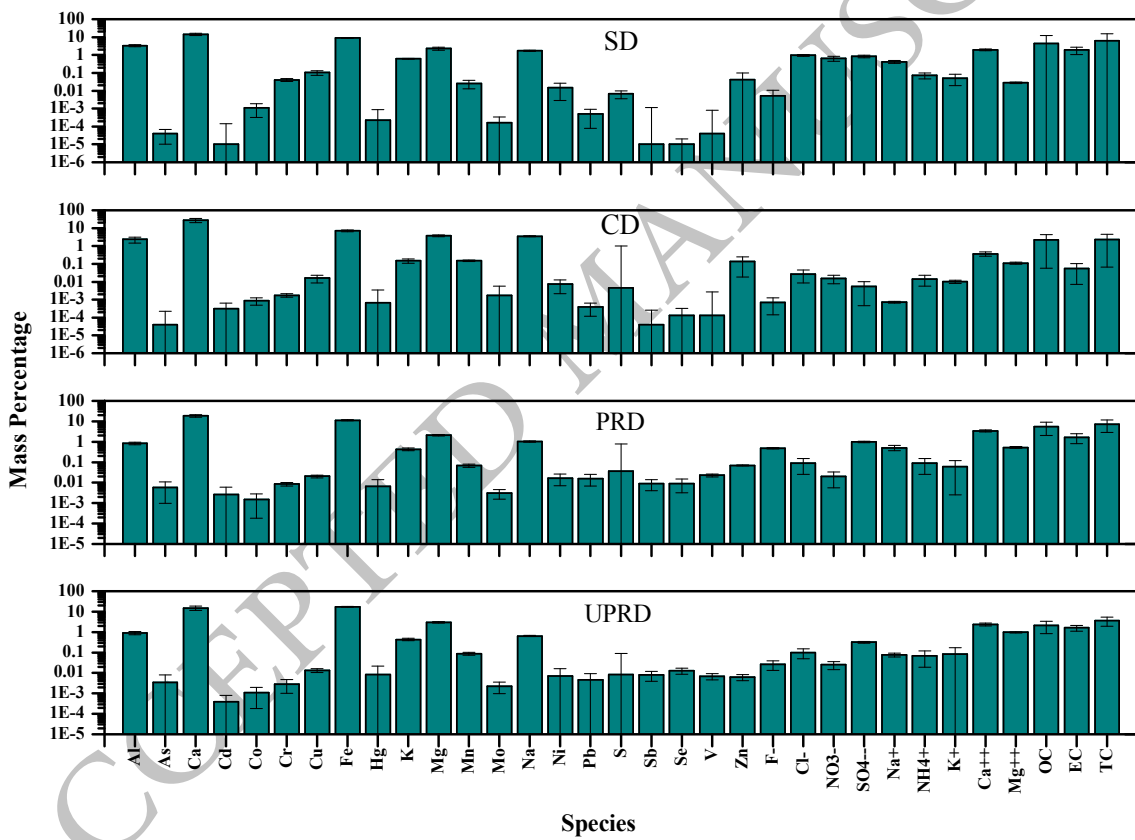


Figure 1. PM_{10-2.5} source profiles for non-agriculture soil dust (SD), construction dust (CD), paved road dust (PRD), and unpaved road dust (UPRD)

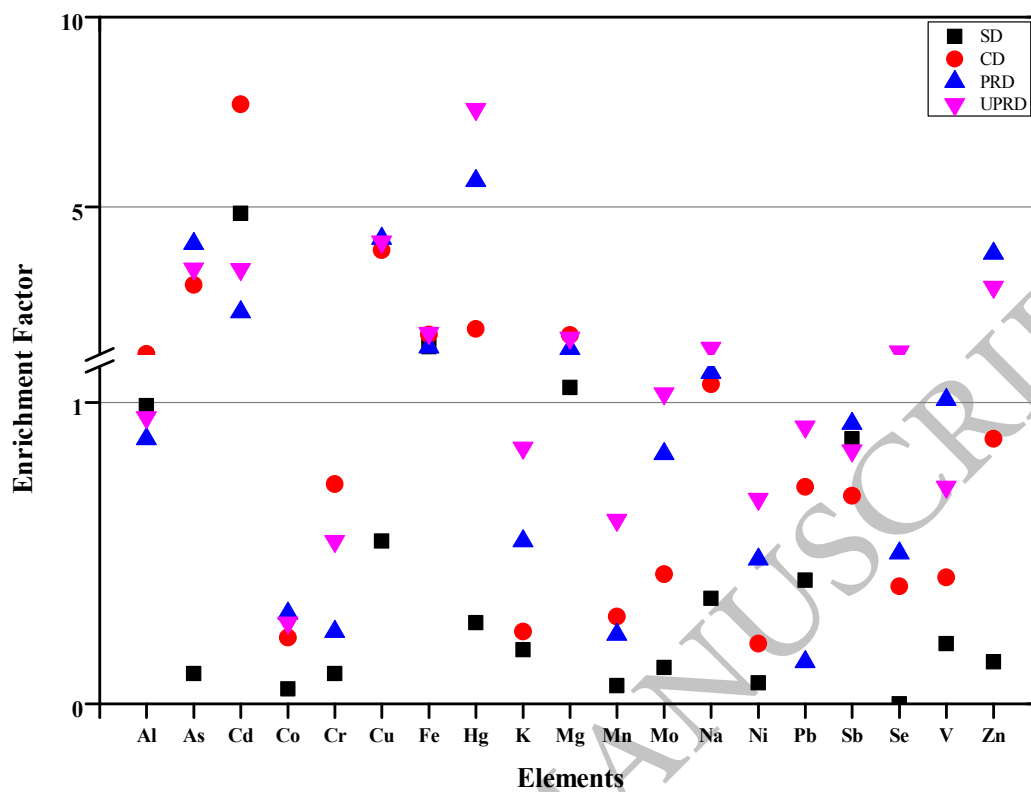


Figure 2. Enrichment Factors (EF) for elements in $PM_{10-2.5}$ resuspended dust. Ca was used as the reference element.

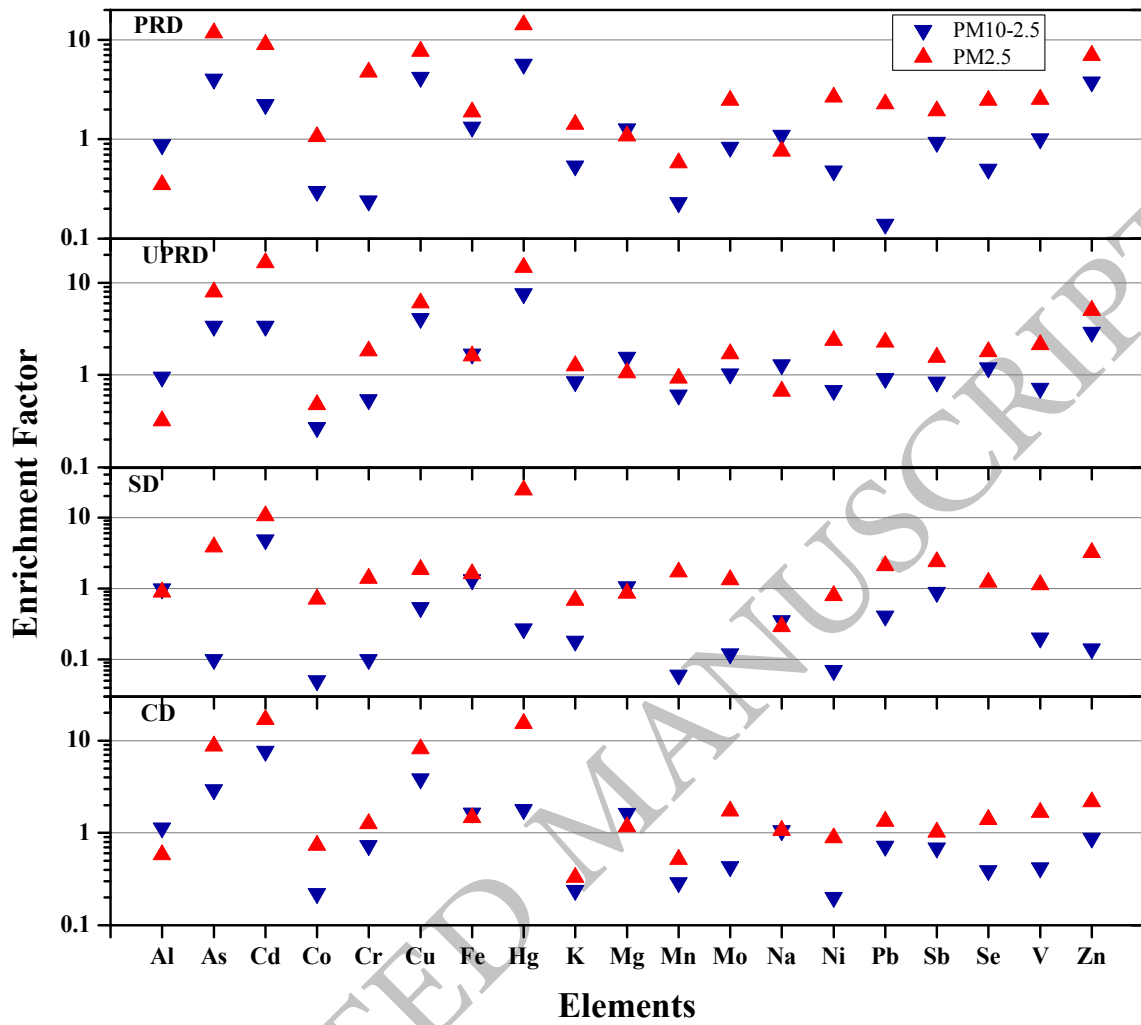


Figure 3. Enrichment Factors for PM_{10-2.5} and PM_{2.5} resuspended dust

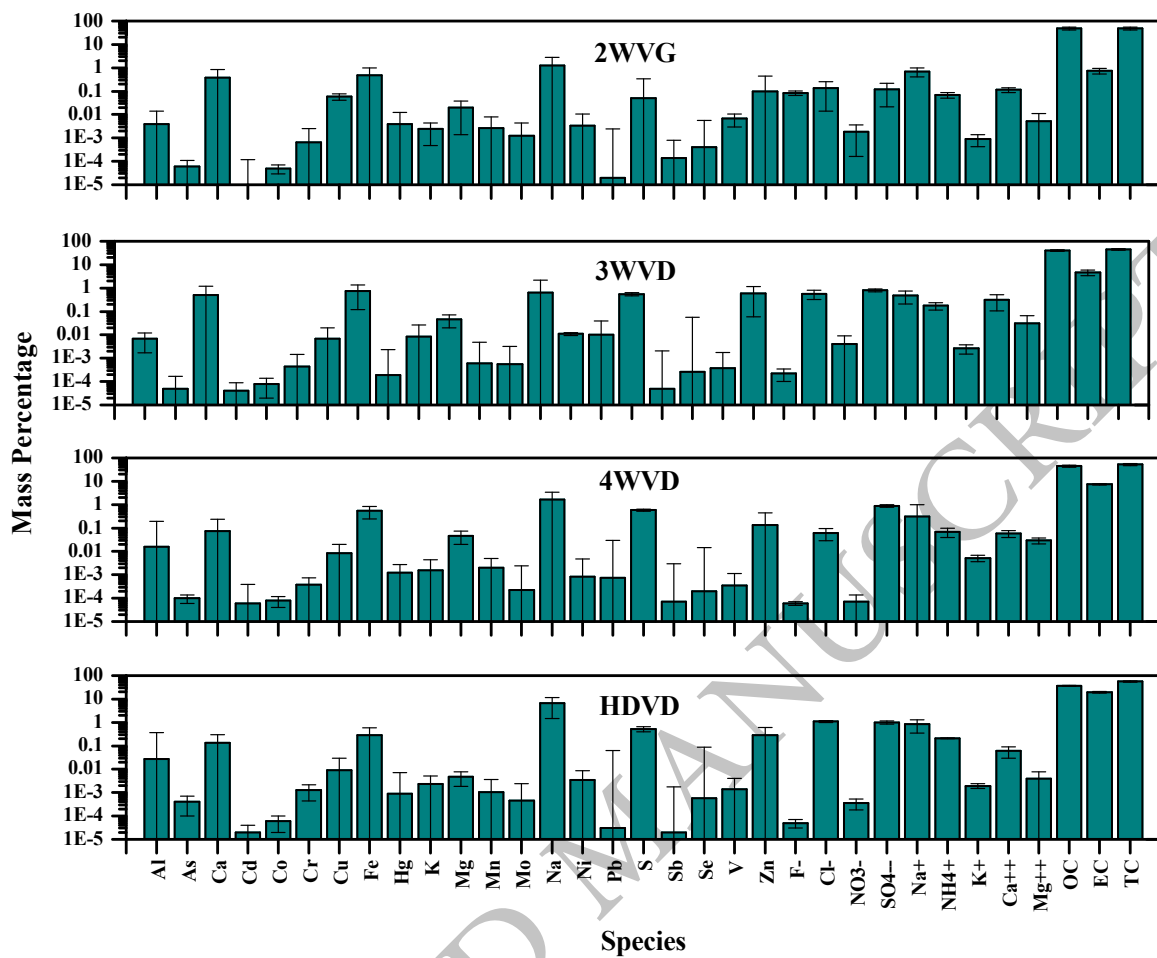


Figure 4. $PM_{10-2.5}$ vehicle exhaust profiles for gasoline two-wheeler (2WVG), diesel three- and four-wheelers (3WVD and 4WVD), and heavy-duty diesel (HDVD) vehicles.

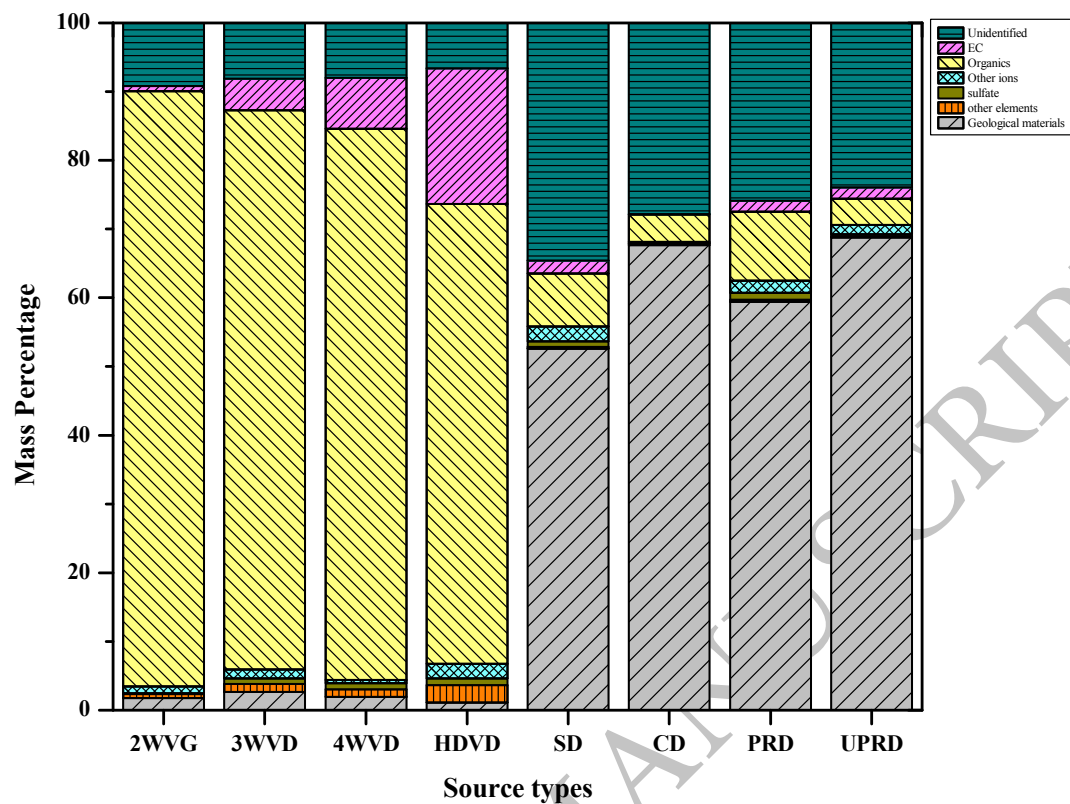


Figure 5. PM_{10-2.5} Mass reconstruction for vehicle and resuspended dust sources (See Table 1 for profile Mnemonics).