

Gas- and Water-phase PAHs Emitted from a Single Hydrogen-Oxygen PEM Fuel Cell

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Abstract

This study focuses on the comparison between gas- and water-phase polycyclic aromatic hydrocarbons (PAHs) emitted from a single hydrogen-oxygen proton exchange membrane (PEM) fuel cell (FC) under different flowrates and temperatures. The results show that among 21 PAHs, the most and least dominant species were Nap and BeP, respectively. At 65°C, the concentrations of individual gas- and water-phase PAHs decreased with increasing flowrate, and the PAH concentrations were lower at anode than at cathode. The concentrations of gas-phase Total-PAHs or Total-BaP_{eq} were slightly lower at 65°C than at 90°C, but an opposite trend was observed for water-phase ones. Temperature influenced more on water-phase than on gas-phase PAH concentration profiles, and the gas- and water-phase PAHs had different concentration profiles. The performance of membrane-electrode assembly (MEA) decreased with increasing flowrate or temperature. The emission factor (EF) sum (anode + cathode) for gas- or water-phase Total-PAHs increased with increasing flowrate. This tendency is also true for gas-phase Total-PAHs EFs but not for water-phase ones when raising temperature from 65°C to 90°C. At 65°C and 52/35 sccm, the EF sums of water-phase Total-PAHs and Total-BaP_{eq} were 2.18±0.04 and 0.09±0.00 µg g-MEA⁻¹, respectively, smaller than those of gas-phase ones (3.02±0.09 and 0.12±0.00 µg g-MEA⁻¹, respectively). More environmental concern should be paid to the emitted gas-phase PAHs than water-phase ones because the anode and cathode effluents should be recycled and reused during PEMFC operation.

Keywords: Polycyclic Aromatic Hydrocarbon; Proton Exchange Membrane Fuel Cell; Emission Factor; Gas-phase PAH; Water-phase PAH.

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39 INTRODUCTION

40 Hydrogen-oxygen proton exchange membrane fuel cells (PEMFCs), regarded as a
41 zero-emission green power generators, are promising alternative power sources for
42 stationary and mobile applications (Wang *et al.*, 2006). The key advantages of
43 PEMFCs are high energy conversion efficiencies, quick start-up, and zero by-product
44 emissions (Tenson and Baby, 2017). An H₂-O₂ PEMFC has the main components of
45 membrane-electrode assembly (MEA), gas-diffusion layer (GDL), and bipolar/end
46 plate (Huang *et al.*, 2006; Lai *et al.*, 2012; Tenson and Baby, 2017). It is common that
47 a PEM is symmetrically sandwiched by anode and cathode to prepare the MEA;
48 moreover, the electrodes usually contain precious metal catalysts (e.g., Pt) (Huang *et*
49 *al.*, 2006; Lai *et al.*, 2012), Pt-based alloys (Zhang *et al.*, 2017), or non-platinum
50 group metals (Reshetenko, *et al.*, 2016) supported by carbon materials (e.g., carbon
51 black, fiber, and nanotube) which are also used for the fabrication of bipolar/end
52 plates and GDLs (e.g., carbon paper or cloth) (Tran *et al.*, 2015; Zeis, 2015; Lee and
53 Lee, 2016).

54 Extensive study has concentrated on accelerating the sluggish kinetics of cathode
55 catalyst, minimizing overall Pt content, and increasing membrane proton conductivity
56 for PEMFCs (Scofield *et al.*, 2015; Tenson and Baby, 2017). Furthermore, hydrogen
57 powered fuel cell electric vehicles are classed as ultra-low emission vehicles when
58 supposing water vapor as the only emission substance in exhaust; therefore, the use of
59 hydrogen powered fuel cell electric vehicles can reduce the emissions of pollutants
60 (which degrade local air quality) and carbon dioxide. However, little attention has
61 been paid to the emission of polycyclic aromatic hydrocarbons (PAHs) from PEMFCs

62 (Huang *et al.*, 2015; Huang *et al.*, 2016),

63 PAHs have long been concerned for their adverse impacts on health and the
64 environment (Tiwari *et al.*, 2015; Kavouras *et al.*, 2015; Pongpiachan, 2016; Li *et al.*,
65 2017). PAHs can be generated from pyrogenic, petrogenic, biogenic, or diagenetic
66 (indoor/outdoor) sources (Stogiannidis and Laane, 2015; Kavouras *et al.*, 2015;
67 Mwangi *et al.*, 2015; Zhang *et al.*, 2016; Li *et al.*, 2017; Fan *et al.*, 2017; Lai *et al.*,
68 2017). In our previous works, we reported the emissions of 21 PAHs from a H₂-O₂
69 PEMFC (installed by a lab-prepared MEA with the same Pt loading of 0.5 mg cm⁻¹ on
70 anode and cathode) (Huang *et al.*, 2015) and effect of operating conditions on PAHs
71 emission from the PEMFC (Huang *et al.*, 2016). In this study, we explored the
72 emissions, concentration profiles, and emission factors of gas- vs. water-phase PAHs
73 from the PEMFC under different flowrates and temperatures.

74

75 **MATERIALS AND METHODS**

76 ***Membrane-Electrode Assembly (MEA) and Fuel Cell Emission Test***

77 The membrane-electrode assemblies (MEAs) tested in this study were all E-TEK
78 commercial MEAs purchased from Micro Power Fuel Cell Co., Ltd. These MEAs had
79 the same catalyst loading in both anode and cathode (0.4 mg-Pt cm⁻²) which were
80 separated with a proton exchange membrane (Nafion-117). An MEA commonly
81 consists of a proton exchange membrane (solid electrolyte), two dispersed catalyst
82 layers, and two gas diffusion layers (GDL). The structure of MEA was addressed in
83 our earlier papers (Huang *et al.*, 2006; Lai *et al.*, 2012; Huang *et al.*, 2015; Huang *et*
84 *al.*, 2016).

85 The single H₂-O₂ PEMFC used in PAH emission tests had an MEA symmetrically
86 sandwiched by Teflon gaskets, carbon (graphite) blocks (with gas flow channel), and
87 copper current collectors. The single cell was installed on a standard fuel cell test

88 station that worked at 65°C or 90 °C (80°C and 70°C for the anode and cathode
89 humidifiers, respectively) and 0.6 V for 48 h for PAH collection under different anode
90 hydrogen and cathode oxygen flow rates which were regulated by two mass flow
91 controllers. The details of sampling system connected to the single cell and PAH
92 collection were provided in our previous works (Huang *et al.*, 2015; Huang *et al.*,
93 2016). For comparison, the PAH emissions from the single PEMFC were also tested
94 at various anode/cathode flowrates of 30/18, 100/100, and 200/200 sccm. It is almost
95 impossible to form PAHs under the operating conditions of this study because it is an
96 electrochemical process and the only “fuel” was hydrogen, so the PAHs collected
97 from the anode/cathode emission gases and effluents were mainly from the desorption
98 of PAHs which originally absorbed on the carbon materials of MEA in the fuel cell
99 during operation (Huang *et al.*, 2015; Huang *et al.*, 2016).

100

101 ***PAH Sampling, Analysis, and Quality Control***

102 In this study, the PAH sampling system, connected to the PEMFC, simultaneously
103 collected PAHs from the anode and cathode emission ports. For each sampling, a
104 glass water collector was used to collect the cell effluent containing water-phase
105 PAHs; the water collector was serially connected with a glass cartridge (packed with
106 XAD-16 resin granules and supported by a polyurethane foam (PUF) plug) for the
107 collection of gas-phase PAHs (Fig. 1). Triplicate PAH samplings (n = 3) were
108 conducted in each experiment. After PAH collection, cell polarization tests were
109 performed at different anode/cathode flowrates or cell temperatures.

110 We measured 21 PAHs: Naphthalene (Nap), Acenaphthylene (AcPy),
111 Acenaphthene (Acp), Fluorene (Flu), Phenanthrene (PA), Anthracene (Ant),
112 Fluoranthene (FL), Pyrene (Pyr), Benzo(a)anthracene (BaA), Chrysene (CHR),
113 Cyclopenta(c,d)pyrene (CYC), Benzo(b)fluoranthene (BbF), Benzo(k)fluoranthene

114 (BkF), Benzo(e)pyrene (BeP), Benzo(a)pyrene (BaP), Perylene (PER),
115 Indeno(1,2,3,-cd)pyrene (IND), Dibenzo(a,h)anthracene (DBA), Benzo(b)chrycene
116 (BbC), Benzo(ghi)perylene (BghiP) and Coronene (COR). The 21 PAH species were
117 classified into three categories: low molecular weight (LMW-PAHs, 2-/3-ring PAHs:
118 NaP, AcPy, AcP, Flu, Ant, and PA), middle molecular weight (MMW-PAHs, 4-ring
119 PAHs: FL, Pyr, BaA, and CHR), and high molecular weight (HMW-PAHs,
120 5-/6-/7-ring PAHs: CYC, BbF, BkF, BeP, BaP, PER, DBA, BbC, IND, BghiP, and
121 COR). The Total-PAHs concentration was obtained from the sum of the
122 concentrations of 21 PAH species.

123 The details of PAHs extraction by Soxhlet extraction method, the
124 identification/quantification of PAHs (with blank correction) by gas
125 chromatograph/mass spectrometry (GC/MS), and the quality assurance (QA)/quality
126 control (QC) procedures for PAHs GC/MS analysis were stated elsewhere (Huang *et*
127 *al.*, 2015). The correlation coefficients of the calibration curves were 0.995–0.999.
128 Ten consecutive injections of a PAH 610-M standard yielded an average relative
129 standard deviation (RSD) of the GC/MS integration area of 3.0% with a range of 0.8–
130 5.1%. The total recovery efficiencies of PAHs from seven consecutive injections
131 ranged from 74 to 110%, while the average recoveries of the five internal standards
132 were 85–93% across seven consecutive injections.

133 The coefficients of variation for repeat injections of the standard solution
134 (containing PAH Mixture-610 M (16 PAHs) and five Merck PAH standards) were all
135 less than 5% for all of the analyzed PAHs, whereas those obtained by replicate
136 analysis were 2–10%. Analyses of serial dilutions of PAH standards revealed the
137 limits of detection (LODs) of GC/MS to be between 0.010 and 0.474 ng for the 21
138 PAH compounds. The limit of quantification (LOQ) was defined as the LOD divided
139 by the sampling volume. The LOQ values of the 21 PAH compounds for cell effluent

140 and emission gas were 0.102–2.470 ng L⁻¹ and 0.076–0.175 ng Nm⁻³, respectively.

141

142 **RESULTS AND DISCUSSION**

143 *Effect of Flowrate on Gas-phase PAHs Emission*

144 For the tests of PEMFC PAH emissions at different flowrates, the anode/cathode
145 flowrates of 52/35 sccm represents 1.5 and 2 times of the stoichiometric requirements
146 for the anode and cathode feeding gases, respectively (Huang *et al.*, 2006; Lai *et al.*,
147 2012; Huang *et al.*, 2015; Huang *et al.*, 2016). Fig. 2 shows the anode and cathode
148 concentration profiles of gas-phase 21 PAHs collected from the exhaust of fuel cell
149 operated at 65°C and different flowrates. At anode/cathode flowrates of 52/35 sccm,
150 the concentrations of PAHs in anode and cathode emission gas ranged from not
151 detectable (ND) (BeP) to 2.961±0.005 (Nap) and ND (BeP) to 4.115±0.005 (Nap) µg
152 Nm⁻³, respectively (Fig. 2(a)), while most of the other species had concentrations of
153 0.200–0.800 µg Nm⁻³. The same PAH species showed a higher concentration in the
154 cathode emission gas than in the anode one. A similar trend was also found at the
155 lower anode/cathode flowrates of 30/18 sccm (Fig. 2(b)). This phenomenon is mainly
156 attributed to the fact that the sampling volume was greater at anode than at cathode
157 but both electrode had similar or slightly different PAH mass emission, regardless of
158 the difference in operation between anode and cathode.

159 The operating potential of anode was <1.23 V (theoretical) vs. standard hydrogen
160 electrode (SHE), while that of cathode was ~ 0 V vs. SHE. Moreover, the main
161 electrochemical reaction at anode is hydrogen oxidation which can be expressed as
162 the following reaction:



164 On the other hand, the principle electrochemical reaction at cathode is oxygen
165 reduction, described by reaction 2:



167 Nevertheless, the emission concentration of the same PAH species was lower at 30/18
168 sccm than at 52/35 sccm, because the former emitted much less PAH mass than the
169 latter, although the sampling volume was lower at 30/18 sccm than at 52/35 sccm (see
170 more discussion in the last section).

171 Dissimilarly, the concentration of individual PAH species decreased, when the gas
172 flowrate increased from 52/35 to 100/100 sccm, except NaP (Fig. 2(c)); hence, the
173 emission of Nap was more influenced by flowrate than those of the other PAH species.
174 This result is majorly attributable to the higher sampling volume at the greater
175 flowrate, although more PAH mass was present in the anode and cathode emission
176 gases at the greater flowrate (see more discussion in the last section). In our earlier
177 study, we also observed this phenomenon when using the same sampling volume at
178 different flowrates (52/35 and 100/100 sccm) or different sampling volumes (or time
179 periods: 12, 24, and 36 hr) at the same flowrate (52/35 sccm) (Huang *et al.*, 2016).
180 When the gas flowrate further increased from 100/100 to 200/200 sccm, more PAHs
181 desorption occurred and thus was collected in the anode or cathode emission gas (see
182 more discussion in the last section) (Fig. 2(d)). Therefore, flowrate significantly
183 influenced the desorption of PAHs from carbon black support during operation,
184 despite the possible strong π - π interaction between PAHs and the graphitic structure of
185 carbon materials (Kah *et al.*, 2011) and the fact that PAHs are relatively easier to
186 adsorb on than to desorb from the surface of carbon blacks (Moninot, 2010).

187 Of all the 21 PAHs, NaP was the most predominant species at different flowrates
188 while BeP was only detected at 200/200 sccm. Some PAH species such as Flu, CHR,
189 BkF, BbC, and COR were comparably dominant as PA at 30/18, 52/35 and 100/100
190 sccm, but these compounds were significantly less dominant than PA at 200/200 sccm.
191 Consequently, flowrate also affected the desorption of individual PAH species and

192 thus its dominance in the anode and cathode emission gases. This result is also
193 probably associated with the electronegative functional groups of compounds,
194 because aromatic compounds having strong electronegative functional groups
195 generally form stronger bonds with the graphene surface of carbon materials (Ma and
196 Uddin, 2013).

197

198 ***Effect of Flowrate on Water-phase PAHs Emission***

199 For comparison, the water-phase PAHs emitted from anode and cathode gases
200 were also collected under the same operating conditions used for gas-phase PAHs
201 sampling. Like the trend of anode versus cathode PAH concentration comparison for
202 gas-phase PAHs, the concentrations of individual PAHs were significantly greater in
203 cathode effluent than in anode effluent (except BkF and BghiP) at 65°C and different
204 flowrates (Figs. 3(a)–(d)). Again, among the 21 PAHs, NaP had the highest
205 concentration, followed by Flu, COR, PA, BbC, CHR, and FL, while BeP was not
206 detected under all tested flowrates. However, the concentration difference between the
207 highest (NaP) and the second highest was less for water-phase PAHs than for
208 gas-phase PAHs. This result should be associated with the fact that the 21 PAHs have
209 low water solubility. Roughly, the water solubility of PAHs decreases as the aromatic
210 ring number increases; for example, the solubility data of Nap, Pyr, and BaP at 25°C
211 are 31.0, 0.132, and 0.0038 mg L⁻¹, respectively (ER Wiki, 2017). The NaP
212 concentrations in anode effluents at 30/18, 52/35, 100/100, and 200/200 sccm were
213 1.759±0.057, 0.766±0.122, 0.419±0.053, and 0.220±0.028 µg L⁻¹, respectively, while
214 the corresponding data in cathode effluents were 8.530±0.178, 3.871±0.253,
215 2.325±0.061, and 0.954±0.060 µg L⁻¹, respectively. BeP was not detectable (ND) in
216 the anode and cathode effluents at all the tested flowrates.

217 It was also observed that the concentrations of individual water-phase PAHs were

218 smaller at anode/cathode flowrate = 52/35 sccm than at anode/cathode flowrate =
219 30/18 sccm, mainly because the collected water volume was greater at 52/35 sccm
220 (Fig. 3(a)) than at 30/18 sccm (Fig. 3(b)), but their emitted PAH masses were
221 comparable (see more discussion in the last section). For the same reason, the
222 concentrations of individual PAHs were greater at 52/35 sccm than at 100/100 sccm
223 (Fig. 3(c)). This tendency is also true when increasing the flowrate to 200/200 sccm
224 (Fig. 3(d)).

225

226 *Effect of Temperature on Gas- and Water-phase PAHs*

227 Fig. 4(a) shows the concentration profiles of 21 individual PAHs in anode and
228 cathode emission gases at 90°C and 52/35 sccm. Again, the concentration of BeP was
229 ND in anode and cathode emission gases. This result is also true in the anode and
230 cathode effluents (Fig. 4(b)). In anode emission gas, the concentrations of individual
231 PAHs ranged from 0.102±0.111 (DBA) to 3.055±0.358 (NaP) $\mu\text{g Nm}^{-3}$, excluding BeP;
232 the corresponding values were from 0.152±0.001 (DBA) to 4.585±0.718 (NaP) μg
233 Nm^{-3} in cathode emission gas. However, they ranged from 0.168±0.003 (DBA) to
234 0.557±0.106 (NaP) $\mu\text{g L}^{-1}$ in anode effluent and from 0.192±0.001 (BghiP) to
235 2.856±0.205 (NaP) $\mu\text{g L}^{-1}$ in cathode effluent. The concentrations of individual PAHs
236 were also higher in the anode side than in the cathode side, except that of water-phase
237 BkF at 90°C. At 52/35 sccm, the concentrations of gas-phase PAHs were similar at
238 both 65°C (Fig. 2(a)) and 90°C (Fig. 4(a)), while the water-phase PAH concentrations
239 were greater at 65°C (Fig. 3(a)) than at 90°C (Fig. 4(b)), consistent with the fact that
240 the water solubility of a PAH species was lower when the temperature is higher.

241

242 *Comparison of Gas- and Water-phase PAH Emission Profiles*

243 In this study, the concentrations of gas- and water-phase individual PAHs in

244 anode and cathode emission gases were normalized to those of Total-PAHs for the
245 comparison of their emission profiles (fingerprints) which denoted the fraction of a
246 PAH species (PAH_i , $i = 1-21$) using the following calculation:

247 Fraction (%) of $\text{PAH}_i = 100 \times (\text{concentration of } \text{PAH}_i / \text{concentration of Total-PAHs})$ (3).

248 For the anode emission under various flowrates or temperatures, the
249 concentration profiles of gas-phase individual PAHs were similar (Fig. 5(a)). A similar
250 trend was also observed for the gas-phase individual PAHs from cathode emission
251 (Fig. 5(b)). Moreover, the concentration profiles of individual PAHs from anode and
252 cathode emission gases were similar. Therefore, the flowrate or temperature in the
253 tested ranges did not significantly influence the concentration profile of gas-phase
254 individual PAHs. On the other hand, the concentration profiles of water-phase
255 individual PAHs from anode effluent looked various in the sector from BkF to DBA at
256 the lowest and highest feeding gas flowrates (30 and 200 sccm, respectively),
257 although the profiles in the other sector were similar (Fig. 5(c)). This tendency was
258 also observed for the PAH concentration profiles of cathode effluents in the sector
259 from BaA to DBA; in addition to feeding gas flowrate, temperature (90°C) also
260 affected the pattern of profile (Fig. 5(d)). However, the PAH concentration profiles of
261 cathode effluents were not similar to those of anode effluents in the sector from BaA
262 to DBA. These findings might be related to the effect of feeding gas flowrate or
263 temperature on the PAH water solubility, particularly for the PAHs with relatively low
264 water solubility (BaA, CHR, BbF, BkF, PER, IND, and DBA) (ER Wiki, 2017).
265 Moreover, the concentration profiles of water-phase PAHs were largely dissimilar to
266 those of gas-phase PAHs. The partitioning of individual PAH species between gas and
267 water phases should be responsible for this difference (Lee *et al.*, 2004).

268

269 ***Effect of Flowrate and Temperature on Gas-phase PAHs-associated BaP_{eq}***

270 **Emission**

271 At 52/35 sccm, the concentrations of molecular-weight (MW) classified PAHs
272 in anode and cathode emission gases varied with the order LMM- > HMM- >
273 MMM-PAHs (4.421 ± 0.038 , 2.942 ± 0.003 , and $1.536 \pm 0.038 \mu\text{g Nm}^{-3}$, respectively for
274 anode (Fig. 6(a)) and 6.504 ± 0.612 , 4.442 ± 0.061 , and $2.236 \pm 0.120 \mu\text{g Nm}^{-3}$,
275 respectively for cathode (Fig. 6(b)). This phenomenon also existed at anode/cathode
276 flowrate = 30/18, 100/100, and 200/200 sccm, except the HMM- > LMM- >
277 MMM-PAHs at 18 sccm. This result is attributable to the more concentration variation
278 for LMW-PAHs than for MMW- and HMW-PAHs (particularly Nap) during the
279 operation at different anode/cathode flowrates. The Total-PAHs concentration
280 decreased with the increase of flowrate in the range 52/35–200/200 sccm, chiefly due
281 to the increase of sampling volume with increasing flowrate. This result is not
282 consistent with our previous work using lab-prepared E-TEK MEAs (Huang *et al.*,
283 2016). However, the Total-PAHs concentration also decreased when the flowrate
284 reduced from 52/35 (8.892 ± 0.079 and $13.18 \pm 0.792 \mu\text{g Nm}^{-3}$ for anode and cathode,
285 respectively) to 30/18 (6.834 ± 0.452 and $10.37 \pm 0.395 \mu\text{g Nm}^{-3}$ for anode and cathode,
286 respectively) sccm, mainly because of the less PAH mass emitted at the lower
287 flowrate (see more discussion in the last section).

288 Fig. 6(c) presents that the concentrations of classified PAHs in anode and
289 cathode emission gases also followed the order LMM- > HMM- > MMM-PAHs at
290 90°C and anode/cathode = 52/35 sccm. However, the average LMM-PAHs
291 concentration was greater at the anode side than at the cathode side, opposite to the
292 observation of average HMM-PAHs concentrations at these two side. The
293 concentrations of LMW-, MMW-, HMW-, and Total-PAHs were similar at both
294 temperatures, so the emission of PAHs was more influenced by flowrate than by
295 temperature. At 90°C , the concentrations of anode LMW-, MMW-, HMW-, and

296 Total-PAHs were 4.984 ± 0.381 , 1.416 ± 0.012 , 2.980 ± 0.035 , and $9.380 \pm 0.429 \mu\text{g Nm}^{-3}$,
297 respectively, while the cathode ones were 7.409 ± 0.764 , 2.150 ± 0.102 , 4.426 ± 0.077 ,
298 and $13.99 \pm 0.944 \mu\text{g Nm}^{-3}$, respectively.

299 In this study, we calculated the benzo(a)pyrene (BaP)-equivalent carcinogenic
300 potency (BaP_{eq}) of a PAH species based on the product of its toxic equivalence factor
301 (TEF) and concentration (Nisbet and LaGoy, 1992). (Noted that BaP and DBA had the
302 highest TEF values (1) of all the 21-PAH.) The PAH-derived Total- BaP_{eq}
303 concentration decreased with increasing flowrate (30/18 ($0.547 \pm 0.003/0.675 \pm 0.003$
304 $\mu\text{g Nm}^{-3}$), 52/35 ($0.365 \pm 0.001/0.546 \pm 0.004 \mu\text{g Nm}^{-3}$), 100/100
305 ($0.305 \pm 0.003/0.378 \pm 0.002 \mu\text{g Nm}^{-3}$), and 200/200 ($0.269 \pm 0.020/0.282 \pm 0.016 \mu\text{g}$
306 Nm^{-3}) sccm) in anode and cathode emission gases (Figs. 6(a) and (b), respectively).
307 Nevertheless, The Total- BaP_{eq} concentration in anode or cathode gas only slightly
308 changed when raising temperature from 65°C to 90°C (0.365 ± 0.001 and 0.542 ± 0.003
309 $\mu\text{g Nm}^{-3}$ for anode and cathode, respectively (Fig. 6(c))).

310

311 *Effect of Flowrate and Temperature on Water-phase PAHs-associated BaP_{eq}* 312 *Emission*

313 The concentrations of MW-classified PAHs at 52/35 sccm in anode effluent
314 were in the order HMM- > LMM- > MMM-PAHs (3.646 ± 0.065 , 2.956 ± 0.214 , and
315 $1.655 \pm 0.010 \mu\text{g L}^{-1}$, (Fig. 7(a)), but those in cathode effluent followed the order
316 LMM- > HMM- > MMM-PAHs (14.54 ± 0.258 , 12.81 ± 0.109 , and $8.195 \pm 0.037 \mu\text{g L}^{-1}$,
317 respectively (Fig. 7(b)). This phenomenon was also found at anode/cathode flowrate =
318 30/18, 100/100, and 200/200 sccm. Moreover, the MW-classified PAH concentrations
319 were significantly smaller in the anode effluent than in the cathode effluent regardless
320 of the difference in flowrate, mostly resulted from the observation that the water
321 volume collected from anode was lower than (approximately one-fifth) that collected

322 from cathode at each anode/cathode flowrate. For the same reason, both anode and
323 cathode effluents exhibited decreasing Total-PAHs concentration with increasing
324 flowrate, dissimilar to the observation for anode and cathode emission gases. As a
325 result, the Total-BaP_{eq} concentration also decreased with increasing flowrate (30/18
326 (0.593±0.001/1.086±0.002 µg L⁻¹), 52/35 (0.440±0.003/0.893±0.005 µg L⁻¹), 100/100
327 (0.393±0.004/0.775±0.008 µg L⁻¹), and 200/200 (0.386±0.004/0.819±0.004 µg L⁻¹)
328 sccm) in anode and cathode effluents (except that in cathode effluent at 200/200 sccm
329 (Figs. 7(a) and (b), respectively).

330 At 90°C and anode/cathode = 52/35 sccm, similar to the trend for anode and
331 cathode emission gases at 65°C, the MW-classified PAH concentrations in anode
332 effluent varied with the order HMM- > LMM- > MMM-PAHs, while those in cathode
333 effluent followed the order LMM- > HMM- > MMM-PAHs (Fig. 7(c)). Additionally,
334 the classified PAH concentrations were also significantly smaller in the anode effluent
335 than in the cathode effluent. However, the concentrations of LMW-, MMW-, HMW-,
336 and Total-PAHs were higher at 65°C than at 90°C (anode: 2.655±0.100, 1.644±0.003,
337 3.641±0.053, and 7.940±0.149 µg L⁻¹, respectively; cathode: 13.48±0.086,
338 7.318±0.032, 11.30±0.071, and 32.09±0.185 µg L⁻¹, respectively). The Total-BaP_{eq}
339 concentration in anode effluent was smaller at 65°C (0.440±0.003 µg L⁻¹) than at 90°C
340 (0.524±0.004 µg L⁻¹; however, this tendency was reversed for cathode effluent
341 (0.893±0.005 and 0.818±0.003 µg L⁻¹ at 65°C and 90°C, respectively) (Fig. 7(c)).

343 ***Polarization Curves and Emission Factors***

344 According to the polarization (current-potential) curves shown in Fig. 8, the
345 performance of MEA installed in the PEMFC decreased when flowrate increased at
346 65°C. Furthermore, the MEA performance was even worse at 90°C, which had quick
347 potential drop at current density smaller than 100 mA cm⁻², probably because the

348 Nafion-117 membrane was drier to have lower conductivity at the higher temperature.
349 However, the MEA performance was better at 30/18 sccm than at 52/35 sccm,
350 although the former had lower stoichiometric requirements of feeding gases than the
351 latter (Huang *et al.*, 2015; Huang *et al.*, 2016).

352 Table 1 summarizes the emission factors (EFs) of gas-phase and water-phase
353 Total-PAHs and Total-BaP_{eq} for the anode and cathode and their sums under different
354 operating flowrates or temperatures. The EF of anode gas-phase Total-PAHs increased
355 with increasing flowrate (0.72 ± 0.05 , 1.51 ± 0.01 , 2.73 ± 0.37 , and 5.11 ± 0.18 μg
356 g-MEA^{-1} at 30/18, 52/35, 100/100, and 200/200 sccm, respectively, at 65°C) or
357 temperature (1.51 ± 0.01 and 1.59 ± 0.07 $\mu\text{g g-MEA}^{-1}$ at 65°C and 90°C respectively, at
358 52/35 sccm). A similar phenomenon was also observed for the EFs of cathode
359 gas-phase Total-PAHs ((0.66 ± 0.03) – (4.72 ± 0.30) $\mu\text{g g-MEA}^{-1}$) and thus the sums of
360 anode and cathode gas-phase Total-PAHs EFs (1.73 ± 0.05 , 3.02 ± 0.09 , 5.01 ± 0.47 , and
361 9.83 ± 0.35 $\mu\text{g g-MEA}^{-1}$ at 30/18, 52/35, 100/100, and 200/200 sccm, respectively, at
362 65°C and 3.19 ± 0.13 $\mu\text{g g-MEA}^{-1}$ at 90°C and 52/35 sccm). This result is attributed to
363 the more mass of gas-phase PAHs desorbed from carbon material support at higher
364 flowrate or temperature, and this PAH desorption behavior was more enhanced by
365 increasing flowrate than by increasing temperature. Nevertheless, the EFs of anode
366 and cathode gas-phase Total-PAHs were similar at 52/35 sccm regardless of difference
367 in operating temperature, whereas the EFs of anode gas-phase Total-PAHs were
368 higher than those of cathode ones. This result is possibly associated with the presence
369 of the product (water) generated from the reaction 2 at cathode). The sums of anode
370 and cathode gas-phase Total-PAHs EFs for a commercial E-TEK MEA in this study
371 (3.02 ± 0.09 $\mu\text{g g-MEA}^{-1}$) was lower than that of our previous study using a
372 lab-prepared MEA (21.4 ± 1.28 $\mu\text{g g-MEA}^{-1}$) (Huang *et al.*, 2016), primarily due to the
373 differences in MEA weight and structure. However, it is not appropriate to directly

374 compare the PAH emission factors obtained in this study with those of fuel
375 consumption or driving distance (mileage) based data reported in literature.

376 Increasing flowrate from 30/18 to 200/200 sccm at 65°C also increased the EFs
377 of anode, cathode, and sum water-phase Total-PAHs ((1.15±0.02)–(1.34±0.02),
378 (0.93±0.01)–(1.09±0.01), and (2.08±0.02)–(2.43±0.03) µg g-MEA⁻¹, respectively)
379 (Table 1). However, the EFs of anode, cathode, and water-phase Total-PAHs were
380 slightly greater at 65°C (1.17±0.04, 1.01±0.01, and 2.18±0.04 µg g-MEA⁻¹,
381 respectively) than at 90°C (1.12±0.02, 0.91±0.01, and 2.03±0.02 µg g-MEA⁻¹,
382 respectively). The difference between anode and cathode of water-phase Total-PAHs
383 EFs was less than that of gas-phase Total-PAHs ones. The anode, cathode, and sum
384 EFs were smaller for water-phase Total-PAHs than for gas-phase ones, except those at
385 30/18 sccm. The anode, cathode, and sum EFs of water-phase Total-BaP_{eq} ranged
386 from (0.06±0.00)–(0.09±0.00), (0.02±0.00)–(0.03±0.03), and (0.08±0.00)–(0.12±0.00)
387 µg g-MEA⁻¹, respectively, while those of gas-phase ones were (0.05±0.00)–
388 (0.06±0.00), (0.03±0.00)–(0.06±0.03), and (0.08±0.00)–(0.12±0.00) µg g-MEA⁻¹,
389 respectively. Accordingly, it is suggested that the fuel cell should be operated at
390 anode/cathode flowrates of 52/35 (or 30/18) and 65°C to reduce the emissions of gas-
391 and water-phase PAHs and achieve appropriate MEA's performance.

392 It is known that at equilibrium, the air-water partition coefficient (K_{AW}) of a
393 PAH species is equal to H/RT , where H , R , and T are Henry's constant, gas constant,
394 and absolute temperature, respectively (Shiu and Mackay, 1997; Bamford *et al.*, 1999).
395 PAHs usually show a trend of diminishing H or K_{AW} with increasing molecular mass
396 at constant temperature (Shiu and Mackay, 1997); however, the K_{AW} increases when
397 temperature rises (Bamford *et al.*, 1999). In this study, the fuel cell system might
398 possibly reach steady-state in the operation at fixed flowrate and temperature, but it
399 was not the case of equilibrium. Therefore, it is not appropriate to use our gas- and

400 water-phase PAH data for K_{AW} calculation.

401

402 CONCLUSIONS

403 In this study, at anode/cathode flowrates of 52/35 sccm, the concentrations of
404 PAHs in anode and cathode emission gas ranged from not detectable (ND) (BeP) to
405 2.961 ± 0.005 (Nap) and ND (BeP) to 4.115 ± 0.005 (Nap) $\mu\text{g Nm}^{-3}$. Roughly, the
406 concentrations of individual gas- and water-phase PAHs decreased when feeding gas
407 flowrate increased, and the anode PAH concentrations were lower than the cathode
408 ones at 65°C . At 52/35 sccm, the concentrations of gas-phase PAHs were similar at
409 both 65°C and 90°C , while the water-phase PAH concentrations were greater at 65°C
410 than at 90°C . Similar concentration profiles were observed for anode, cathode, or
411 anode vs. cathode gas-phase PAHs under various flowrates or temperatures; however,
412 those for water-phase PAHs were partially different, and the gas- and water-phase
413 PAHs had different concentration profiles.

414 The concentrations of molecular-weight (MW) classified PAHs in anode and
415 cathode emission gases varied with the order LMM- > HMM- > MMM-PAHs under
416 different flowrates (except 18 sccm) or temperatures, and the gas-phase Total-PAHs or
417 Total-BaP_{eq} concentration decreased with the increase of flowrate (except 30/18 sccm).
418 At 52/35 sccm, the concentrations of gas-phase LMW-, MMW-, and HMW-PAHs at
419 65°C were similar to those at 90°C ; therefore, the gas-phase Total-PAHs or
420 Total-BaP_{eq} concentrations were also similar at both temperatures (Total-PAHs:
421 8.892 ± 0.079 and 13.18 ± 0.792 $\mu\text{g Nm}^{-3}$ for anode and cathode, respectively;
422 Total-BaP_{eq}: 0.365 ± 0.001 and 0.546 ± 0.004 $\mu\text{g Nm}^{-3}$ for anode and cathode,
423 respectively, at 65°C). The concentrations of water-phase Total-PAHs were higher at
424 65°C (7.940 ± 0.149 $\mu\text{g L}^{-1}$ and 32.09 ± 0.185 $\mu\text{g L}^{-1}$ for anode and cathode effluents,
425 respectively) than at 90°C . The Total-BaP_{eq} concentration in anode and cathode

426 effluents were 0.440 ± 0.003 and $0.893 \pm 0.005 \mu\text{g L}^{-1}$, respectively, at 65°C .

427 The performance of MEA decreased with increasing flowrate or temperature. The
428 EF sum (anode + cathode) for gas- or water-phase Total-PAHs increased with
429 increasing flowrate. This tendency is also true for gas-phase Total-PAHs EFs but not
430 for water-phase Total-PAHs EFs when raising temperature from 65°C to 90°C . At
431 65°C and 52/35 sccm, the EF sums of water-phase Total-PAHs and Total-BaP_{eq} were
432 2.18 ± 0.04 and $0.09 \pm 0.00 \mu\text{g g-MEA}^{-1}$, respectively, smaller than those of gas-phase
433 ones (3.02 ± 0.09 and $0.12 \pm 0.00 \mu\text{g g-MEA}^{-1}$, respectively).

434

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440 **REFERENCES**

- 441 Bamford, H.A., Poster, D.L. and Baker, J.E. (1999). Temperature Dependence OF
442 Henry's Law Constants of Thirteen Polycyclic Aromatic Hydrocarbons Between
443 4°C and 31°C . *Environ. Toxicol. Chem.* 18(9): 1905–1912.
- 444 ER Wiki accessible 10/07/2017, Polycyclic Aromatic Hydrocarbons (PAHs),
445 [http://www.environmentalrestoration.wiki/index.php?title=Polycyclic_Aromatic_](http://www.environmentalrestoration.wiki/index.php?title=Polycyclic_Aromatic_Hydrocarbons_(PAHs))
446 [Hydrocarbons_\(PAHs\)](http://www.environmentalrestoration.wiki/index.php?title=Polycyclic_Aromatic_Hydrocarbons_(PAHs)).
- 447 Fan, Z.L., Chen, X.C., Lui, K.H., Ho, S.S.H., Cao, J.J., Lee, S.C., Huang, H. and Ho,
448 K.F. (2017). Relationships between Outdoor and Personal Exposure of
449 Carbonaceous Species and Polycyclic Aromatic Hydrocarbons (PAHs) in Fine
450 Particulate Matter (PM_{2.5}) at Hong Kong. *Aerosol Air Qual. Res.* 17: 666–679.
- 451 Huang, K.L., Lai, Y.C. and Tsai, C.H. (2006). Effects of Sputtering Parameters on the

452 Performance of Electrodes Fabricated for Proton Exchange Membrane Fuel Cells.
453 *J. Power Sources* 156: 224–231.

454 Huang, K.L., Tsai, T.H., Tsai, J.H., Chen, S.J., and Lee, W.J. (2015). Emission of
455 PAHs from a Single Hydrogen-Oxygen PEM Fuel Cell: In relation to Fuel Cell
456 Carbon Materials. *Aerosol Air Qual. Res.* 15: 2654–2667.

457 Huang, K.L., Wu, M.S., Tsai, J.H., Lin, D.Y., Chen, S.J. and Lee, W.J. (2016). Effect
458 of Operating Conditions on PAHs Emission from a Single H₂-O₂ PEM Fuel Cell.
459 *Aerosol Air Qual. Res.* 16: 2186–2197.

460 Kah, M., Zhang, X., Jonker, M.T. and Hofmann, T. (2011) Measuring and Modeling
461 Adsorption of PAHs to Carbon Nanotubes over a Six Order of Magnitude Wide
462 Concentration Range. *Environ. Sci. Technol.* 45: 6011–6017.

463 Kavouras, I.G., DuBois, D.W., Nikolich, G. and Etyemezian, V. (2015). Monitoring,
464 Source Identification and Health Risks of Air Toxics in Albuquerque, New Mexico,
465 U.S.A. *Aerosol Air Qual. Res.* 15: 1160–1167.

466 Lai, Y.C., Huang, K.L., Tsai, C.H., Lee, W.J. and Chen, Y.L. (2012). Sputtered Pt
467 loadings of membrane electrode assemblies in proton exchange membrane fuel
468 cells. *Int. J. Energy Res.* 32(8): 918–927.

469 Lai, Y.C., Tsai, C.H., Chen, Y.L. and Chang-Chien, G.P. (2017). Distribution and
470 Sources of Atmospheric Polycyclic Aromatic Hydrocarbons at an Industrial
471 Region in Kaohsiung, Taiwan. *Aerosol Air Qual. Res.* 17: 776–787.

472 Lee, J.J., Huang, K.L., Yu, Y.Y. and Yang, S.C. (2004). Laboratory Retention of
473 Vapor-phase PAHs Using XAD Adsorbents. *Atmos. Environ.* 38(36): 6185–6193.

474 Lee, D. and Lee, D.G. (2016). Carbon Composite Bipolar Plate for High-Temperature
475 Proton Exchange Membrane Fuel Cells (HT-PEMFCs). *J. Power Sources* 327:
476 119–126.

477 Li, Y.Y., Yang, L.X., Chen, X.F., Gao, Y., Jiang, P., Zhang, J.M., Yu, H. and Wang,

478 W.X. (2017). PM_{2.5}-Bound PAHs in Indoor and Outdoor of Hotels in Urban and
479 Suburban of Jinan, China: Concentrations, Sources, and Health Risk Impacts.
480 *Aerosol Air Qual. Res.* 17: 2463–2473.

481 Ma, X and Uddin S. (2013). Desorption of 1,3,5-Trichlorobenzene from Multi-Walled
482 Carbon Nanotubes: Impact of Solution Chemistry and Surface Chemistry.
483 *Nanomaterials* 3: 289–302.

484 Moninot, G. (2010). Where are the PAHs in the Carbon Black? Available at
485 www.iom3.org/fileproxy/331784 (accessed Aug. 2015)

486 Mwangi, J.K., Lee, W.J., Tsai, J.H. and Wu, T.S. (2015). Emission Reductions of
487 Nitrogen Oxides, Particulate Matter and Polycyclic Aromatic Hydrocarbons by
488 Using Microalgae Biodiesel, Butanol and Water in Diesel Engine. *Aerosol Air*
489 *Qual. Res.* 15: 1160–1167.

490 Nisbet, C. and LaGoy, P. (1992). Toxic Equivalency Factors (TEFs) for Polycyclic
491 Aromatic Hydrocarbons (PAHs). *Regul. Toxicol. Pharmacol.* 16: 290–300.

492 Pongpiachan, S. (2016). Incremental Lifetime Cancer Risk of PM_{2.5} Bound
493 Polycyclic Aromatic Hydrocarbons (PAHs) Before and After the Wildland Fire
494 Episode. *Aerosol Air Qual. Res.* 16: 2907–2919.

495 Reshetenko, T., Serov A., Artyushkova K., Matanovic I., Stariha S. and Atanassov, P.
496 (2016). Tolerance of Non-Platinum Group Metals Cathodes Proton Exchange
497 Membrane Fuel Cells to Air Contaminants. *J. Power Sources* 324: 556–571.

498 Scofield, M.E., Liu, H. and Wong, S.S. (2015). A Concise Guide to Sustainable
499 PEMFCs: Recent Advances in Improving Both Oxygen Reduction Catalysts and
500 Proton Exchange Membranes. *Chem. Soc. Rev.* 44: 5836–5860.

501 Shiu, W.Y. and Mackay, D. (1997). Henry's Law Constants of Selected Aromatic
502 Hydrocarbons, Alcohols, and Ketones. *J. Chem. Eng. Data* 42: 27–30.

503 Stogiannidis, E. and Laane, R. (2015). Source Characterization of Polycyclic

504 Aromatic Hydrocarbons by Using Their Molecular Indices: An Overview of
505 Possibilities, Vol. 234, In Whitacre, D.M. (Ed.) Reviews of Environmental
506 Contamination and Toxicology, Springer International Publishing, Switzerland, p.
507 56–72.

508 Tenson, T.J. and Baby, R. (2017). Recent Advances in Proton Exchange Membrane
509 Fuel Cells: A Review. *IJARSET* 4(SI 6): 34–40.

510 Tiwari, M., Sahu, S.K. and Pandit, G.G. (2015). Inhalation Risk Assessment of PAH
511 Exposure Due to Combustion Aerosols Generated from Household Fuels. *Aerosol*
512 *Air Qual. Res.* 15: 582–590.

513 Tran, P.D., Morozan, A., Archambault, S., Heidkamp, J., Chenevier, P., Dau, H.,
514 Fontecave, M., Martinet, A., Joussemme B. and Artero, V. (2015). A Noble
515 Metal-free Proton-exchange Membrane Fuel Cell Based on Bio-inspired
516 Molecular Catalysts. *Chem. Sci.* 6: 2050–2053.

517 Wang, X., Zhang, H., Zhang, J., Xu, H., Zhu, X.B., Chen, J. and Yi, B.L. (2006). A
518 Bi-functional Micro-porous Layer with Composite Carbon Black for PEM Fuel
519 Cells. *J. Power Sources* 162: 474–479.

520 Zhang, J.Q., Zhan, C.L., Liu, H.X., Liu, T., Yao, R.H. Hu, T.P., Xiao, W.S., Xing, X.L.,
521 Xu, H.M. and Cao, J.J. (2016). Characterization of Polycyclic Aromatic
522 Hydrocarbons (PAHs), Iron and Black Carbon within Street Dust from a Steel
523 Industrial City, Central China. *Aerosol Air Qual. Res.* 16: 2452–2461.

524 Zhang, C.L., Shen, X.C., Pan, Y.B. and Peng, Z.M. (2017). A review of Pt-based
525 Electrocatalysts for Oxygen Reduction Reaction. *Front. Energy* 11(3): 268–285.

526 Zeis, R. (2015). Materials and Characterization Techniques for High-temperature
527 Polymer Electrolyte Membrane Fuel Cells. *Beilstein J. Nanotechnol.* 6: 68–83.

528

529 **Table 1.** Emission factors (EFs) of gas-phase and water-phase Total-PAHs and
 530 Total-BaP_{eq} for the anodes (A), cathodes (C), and sums (A + C) under different
 531 operating flowrates or temperatures.

A/C (sccm)	T (°C)	Total-PAHs EFs (μg g-MEA ⁻¹)			Total-BaP _{eq} EFs (μg g-MEA ⁻¹)		
		A	C	Sum	A	C	Sum
Gas-phase							
30/18	65	0.72±0.05	0.66±0.03	1.37±0.05	0.06±0.00	0.04±0.00	0.10±0.00
52/35	65	1.51±0.01	1.51±0.09	3.02±0.09	0.06±0.00	0.06±0.00	0.12±0.00
100/100	65	2.73±0.37	2.28±0.29	5.01±0.47	0.05±0.00	0.05±0.00	0.10±0.00
200/200	65	5.11±0.18	4.72±0.30	9.83±0.35	0.05±0.00	0.03±0.00	0.08±0.00
52/35	90	1.59±0.07	1.60±0.11	3.19±0.13	0.06±0.00	0.06±0.00	0.12±0.00
Water-phase							
30/18	65	1.15±0.02	0.93±0.01	2.08±0.02	0.09±0.00	0.03±0.00	0.12±0.00
52/35	65	1.17±0.04	1.01±0.01	2.18±0.04	0.06±0.00	0.03±0.00	0.09±0.00
100/100	65	1.21±0.03	1.03±0.02	2.24±0.04	0.06±0.00	0.02±0.00	0.08±0.00
200/200	65	1.34±0.02	1.09±0.01	2.43±0.03	0.06±0.00	0.03±0.00	0.09±0.00
52/35	90	1.12±0.02	0.91±0.01	2.03±0.02	0.07±0.00	0.02±0.00	0.09±0.00

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533

534 **Figure Captions**

535 **Fig. 1.** The scheme of single PEMFC for PAH emission tests (W: water-phase PAHs
536 collector, X: gas-phase PAHs collector (cartridge), Y: flow meter, and Z: pump).

537 **Fig. 2.** Characteristic profiles of gas-phase PAHs emitted at different PEMFC
538 anode/cathode flowrates (65°C): (a) 52/35, (b) 30/18, (c) 100/100, and (d) 200/200
539 sccm (n = 3).

540 **Fig. 3.** Characteristic profiles of water-phase PAHs emitted at different PEMFC
541 anode/cathode flowrates (65°C): (a) 52/35, (b) 30/18, (c) 100/100, and (d) 200/200
542 sccm (n = 3).

543 **Fig. 4.** Characteristic profiles of PAHs emitted at 90°C ((a) gas-phase (G) and (b)
544 water-phase (W)) (anode/ cathode = 52/35 sccm) (n = 3).

545 **Fig. 5.** Comparison of PAHs concentration fraction profiles (normalized to
546 Total-PAHs concentration at different PEMFC anode/cathode flowrates (52/35, 30/18,
547 100/100, and 200/200 sccm) and temperature (65°C and 90°C) (n = 3). (a) anode (A)
548 gas-phase (G), (b) cathode (C) gas-phase, (c) anode water-phase (W), and (d) cathode
549 water-phase.

550 **Fig. 6.** Concentrations of gas-phase LMW-, MMW-, HMW-, and Total-PAHs and
551 Total-BaP_{eq} at different flowrates ((a) anode and (b) cathode) and temperatures (c) (n
552 = 3).

553 **Fig. 7.** Concentrations of water-phase LMW-, MMW-, HMW-, and Total-PAHs and
554 Total-BaP_{eq} at different flowrates ((a) anode and (b) cathode) and temperatures (c) (n
555 = 3).

556 **Fig. 8.** Polarization (current-potential) curves under different operating conditions for
557 the PEMFC.

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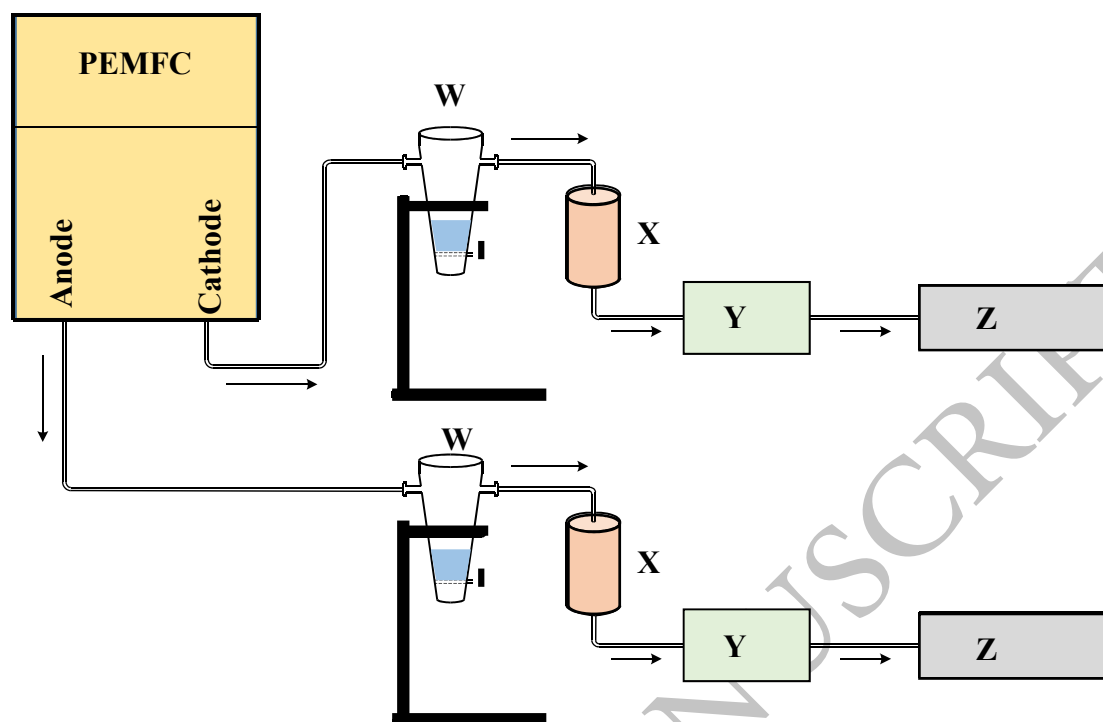
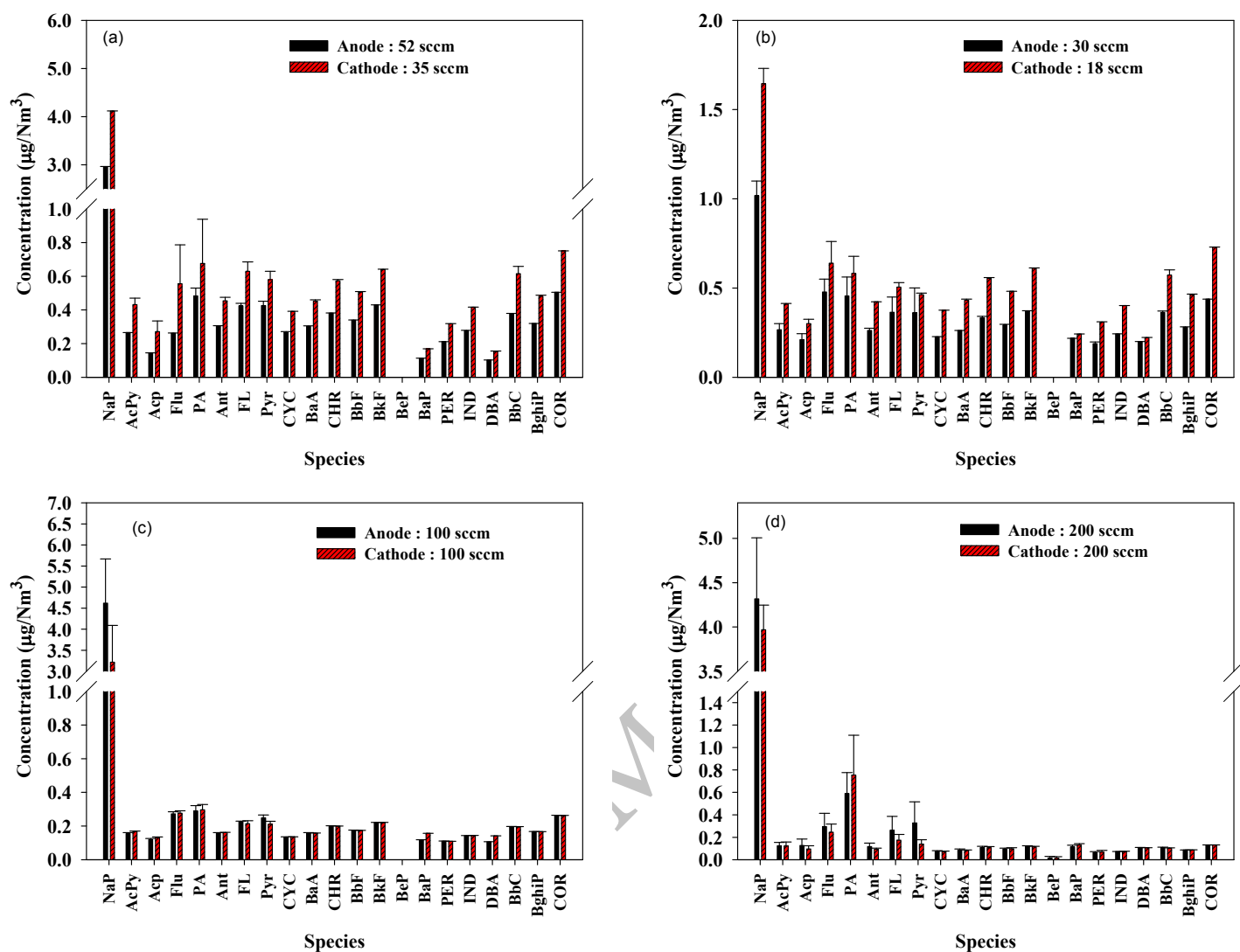


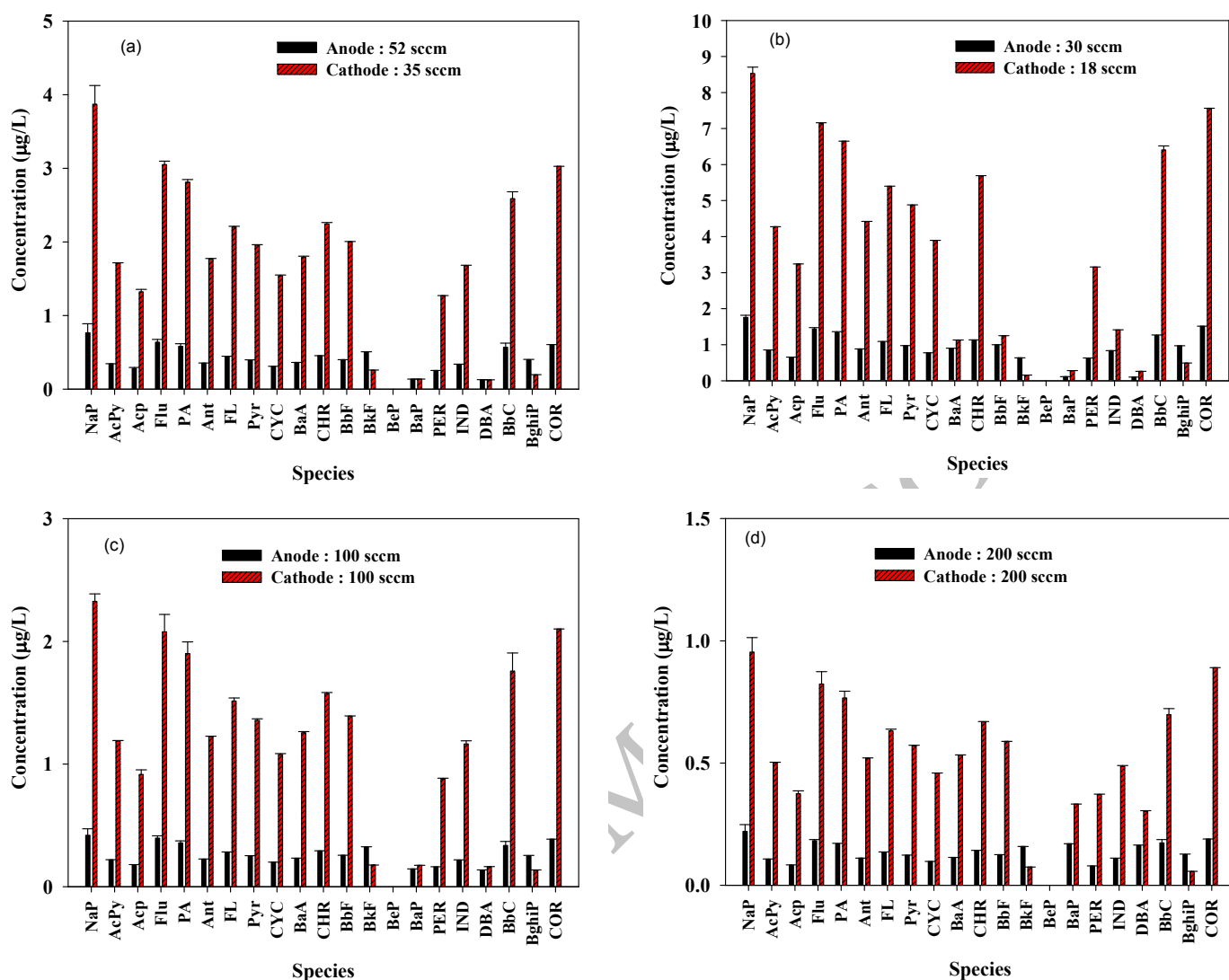
Fig. 1. The scheme of single PEMFC for PAH emission tests (W: water-phase PAHs collector, X: gas-phase PAHs collector (cartridge), Y: flow meter, and Z: pump).

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581 **Fig. 2.** Characteristic profiles of gas-phase PAHs emitted at different PEMFC
 582 anode/cathode flowrates (65°C): (a) 52/35, (b) 30/18, (c) 100/100, and (d) 200/200
 583 sccm (n = 3).
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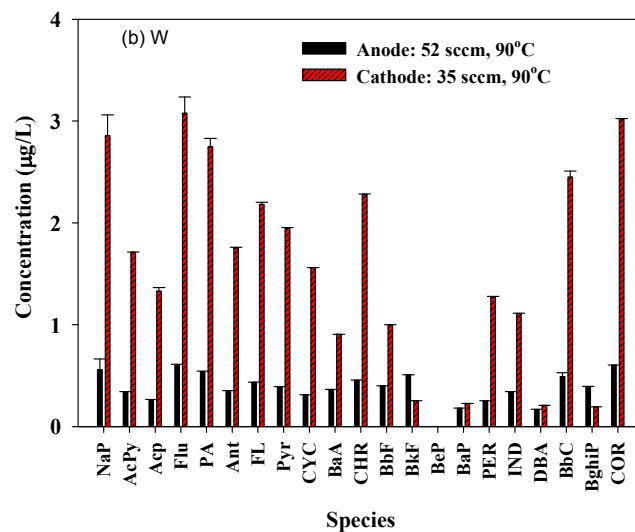
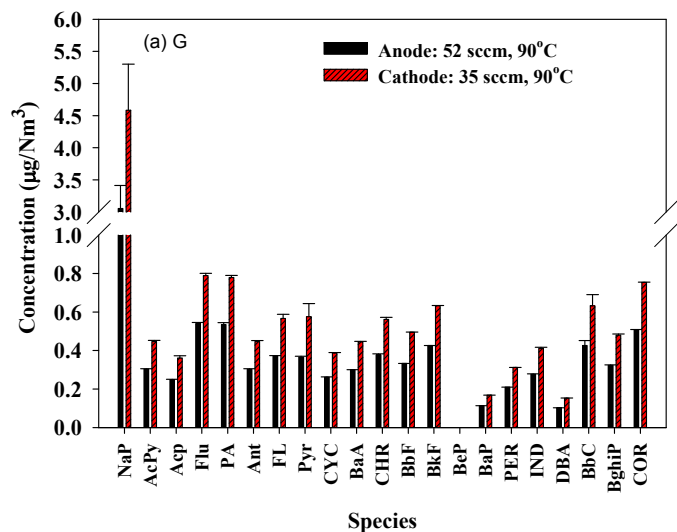
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586 **Fig. 3.** Characteristic profiles of water-phase PAHs emitted at different PEMFC
 587 anode/cathode flowrates (65°C): (a) 52/35, (b) 30/18, (c) 100/100, and (d) 200/200
 588 sccm (n = 3).

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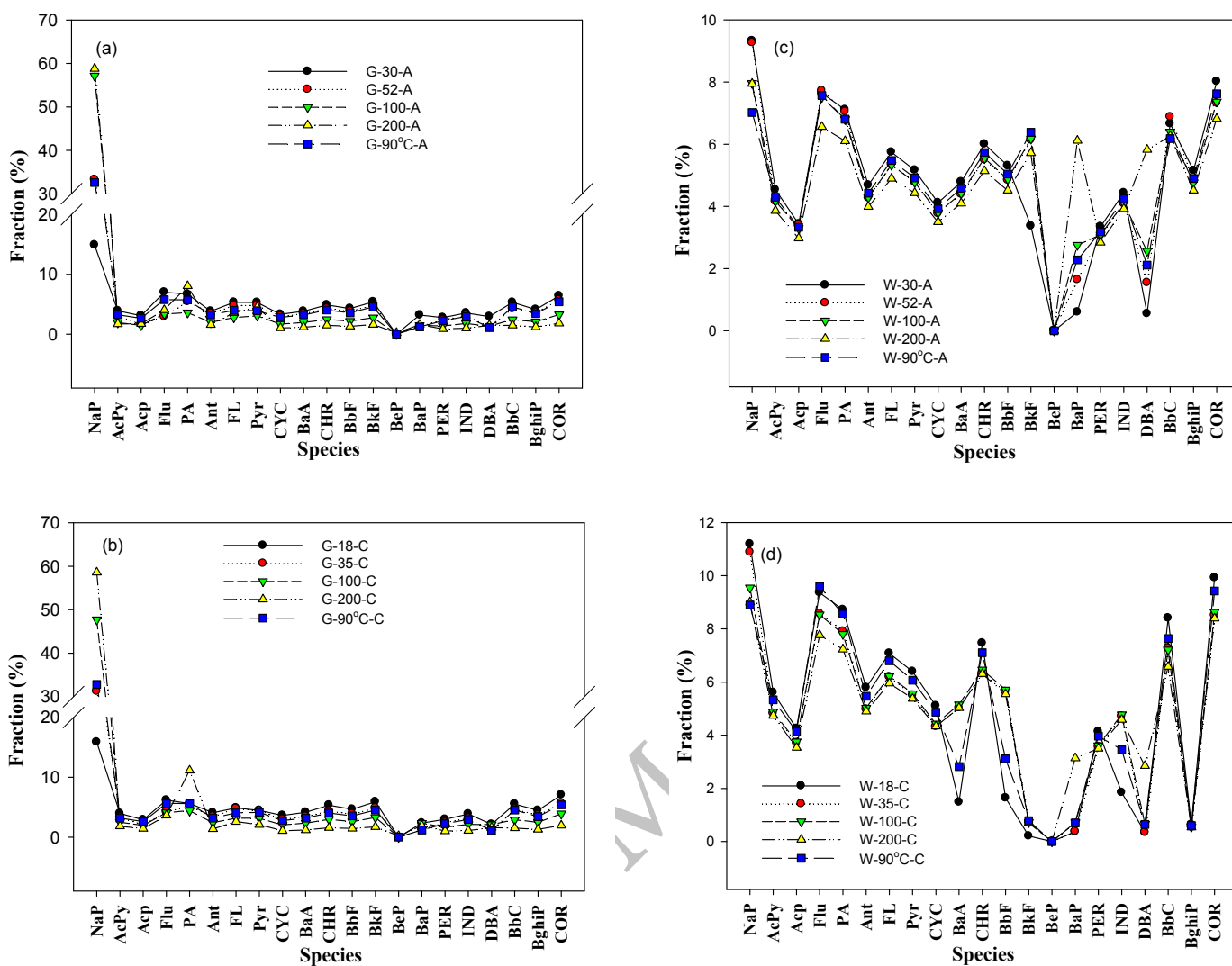


591 **Fig. 4.** Characteristic profiles of PAHs emitted at 90°C ((a) gas-phase (G) and (b)
 592 water-phase (W)) (anode/ cathode = 52/35 sccm) (n = 3).

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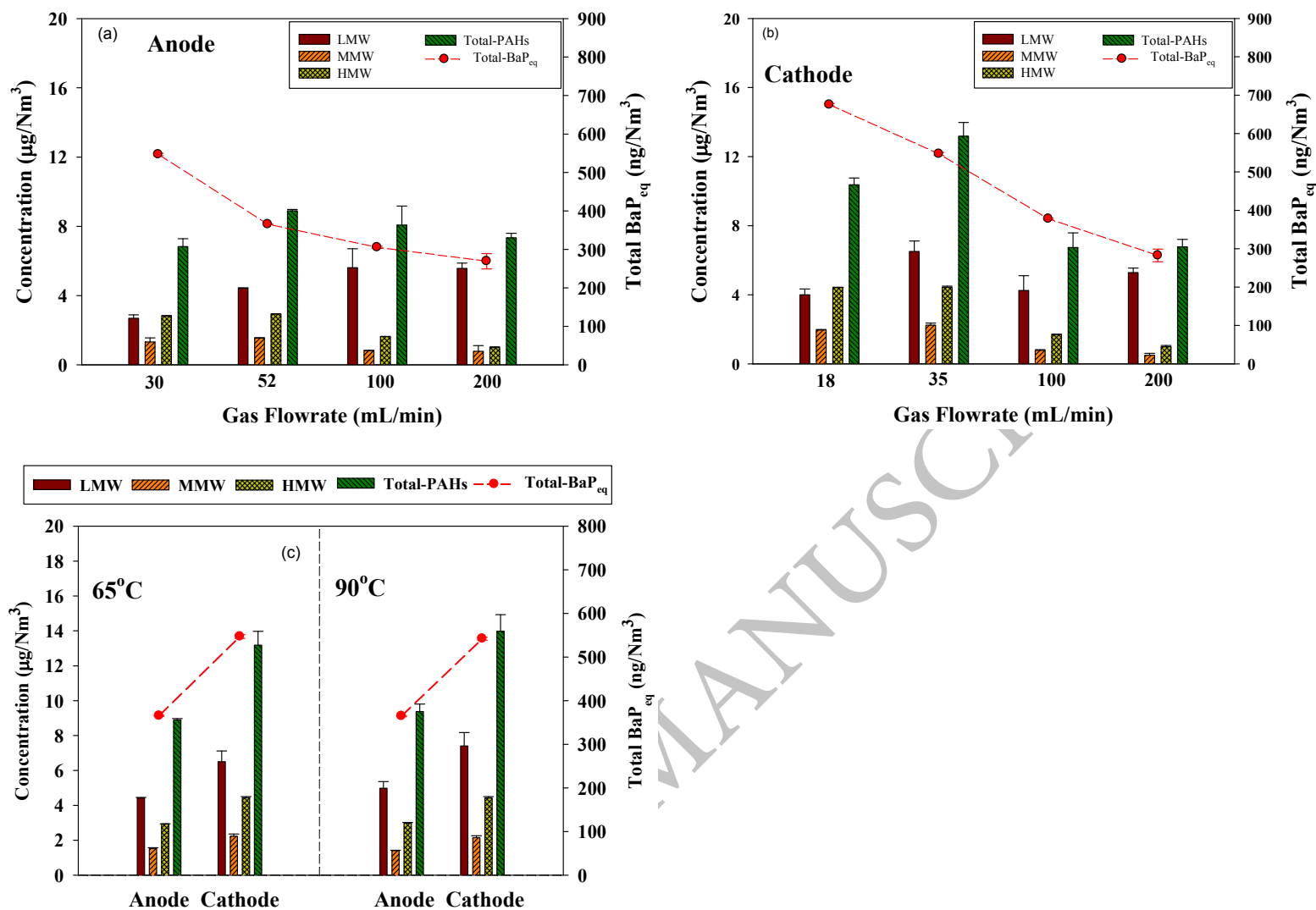
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595 **Fig. 5.** Comparison of PAHs concentration fraction profiles (normalized to
 596 Total-PAHs concentration) at different PEMFC anode/cathode flowrates (52/35, 30/18,
 597 100/100, and 200/200 sccm) and temperature (65°C and 90°C) (n = 3). (a) anode (A)
 598 gas-phase (G), (b) cathode (C) gas-phase, (c) anode water-phase (W), and (d) cathode
 599 water-phase.

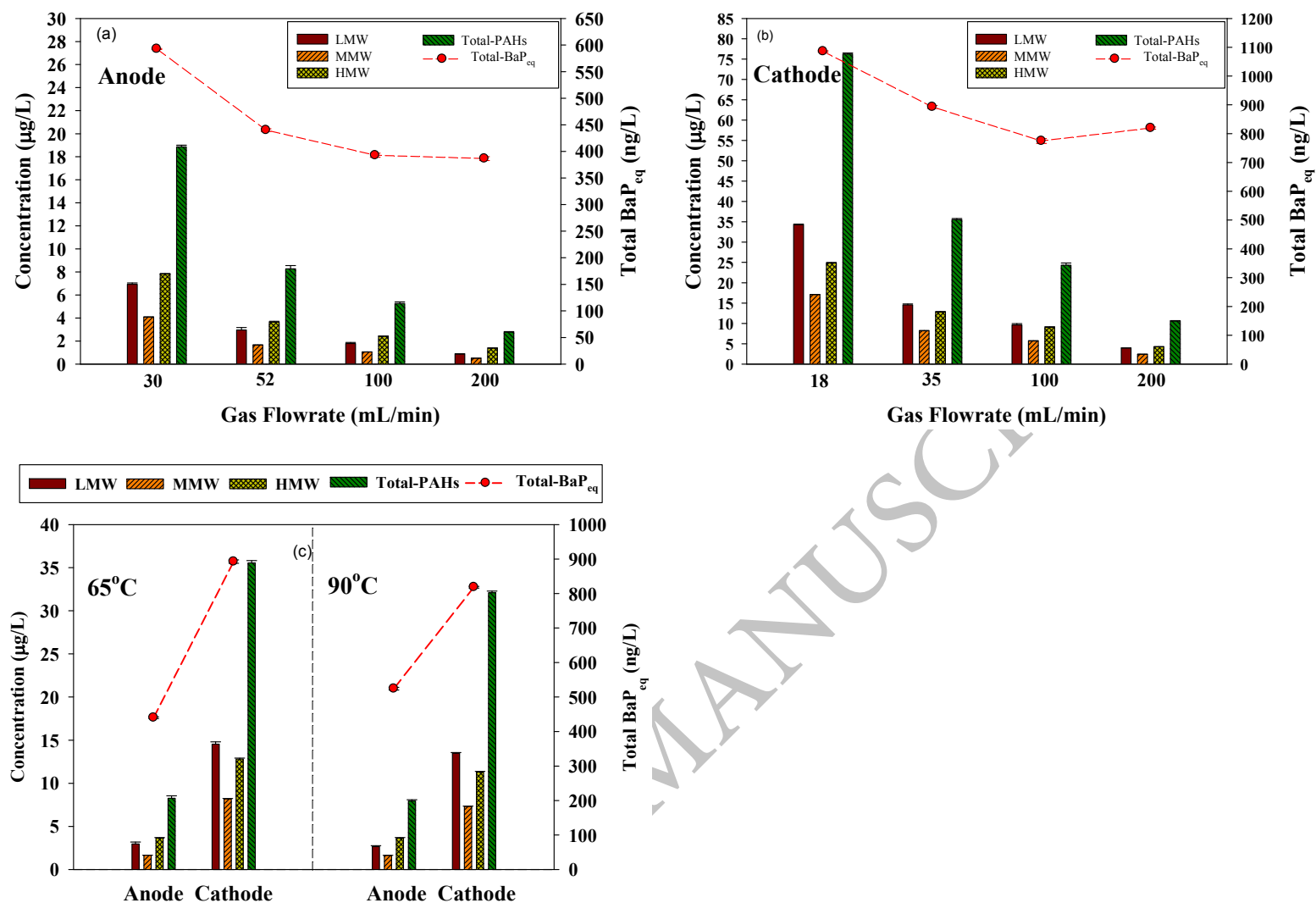
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602 **Fig. 6.** Concentrations of gas-phase LMW-, MMW-, HMW-, and Total-PAHs and
 603 Total-BaP_{eq} at different flowrates ((a) anode and (b) cathode) and temperatures (c) (n
 604 = 3).

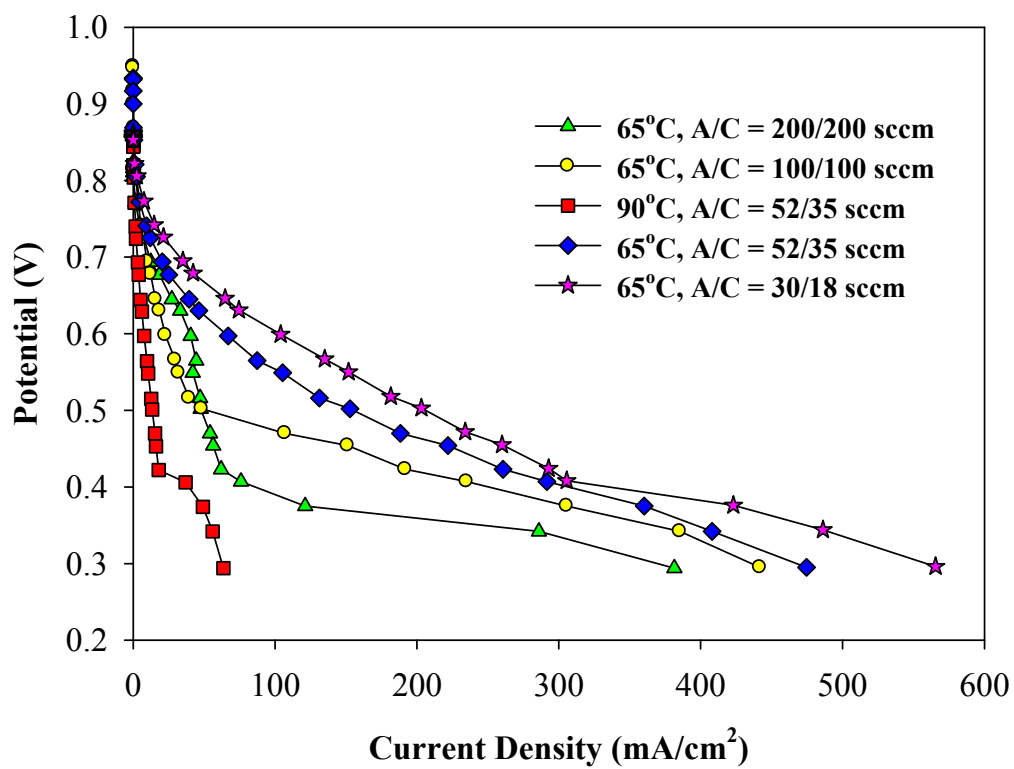
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607 **Fig. 7.** Concentrations of water-phase LMW-, MMW-, HMW-, and Total-PAHs and
 608 Total-BaP_{eq} at different flowrates ((a) anode and (b) cathode) and temperatures (c) (n
 609 = 3).

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614 **Fig. 8.** Polarization (current-potential) curves under different operating conditions for
615 the PEMFC.

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