

# 1 Aqueous secondary organic aerosol formation in ambient cloud water photo-oxidations

2  
3 M.I. Schurman<sup>1,2\*</sup>, A. Boris<sup>1</sup>, Y. Desyaterik<sup>1</sup>, J. L. Collett, Jr.<sup>1</sup>

4  
5 <sup>1</sup>*Department of Atmospheric Science, Colorado State University, Fort Collins, CO, 80521,*  
6 *USA*

7 <sup>2</sup>*Colorado Mountain College, Breckenridge, CO, 80424, USA*  
8

## 9 Abstract

10  
11 The current understanding of aqueous secondary organic aerosol (aqSOA) formation is  
12 based largely on laboratory investigations of very simple surrogate cloud water solutions that  
13 aid mechanistic understanding of aqueous oxidation but may not accurately reflect the  
14 influence of the complex ambient matrix present in authentic cloud waters on organic  
15 chemistry. In this study, unaltered ambient cloud water and ‘biogenically influenced’ ambient  
16 cloud water (with added pinonic acid) were photo-oxidized, atomized, and dried to simulate  
17 the formation of aqSOA in clouds, then analyzed using an Aerodyne Aerosol Mass  
18 Spectrometer. Two major chemical regimes were identified: in the first, particle organic mass  
19 is gained, then lost; sustained increases in highly oxidized fragments indicate overall organic  
20 acid formation, while increases in nominally volatile fragments suggest that evaporation may  
21 contribute to the observed mass decrease. In the second regime, the oxidation level of cloud  
22 water organic matter decreases as mass decreases, suggesting that oxidized functional groups  
23 are fragmented and lost to evaporation. Overall, the rate of aqSOA production in unaltered  
24 cloud water decreases as oxygenation increases, until organic mass loss beginning at  
25 consistent values of  $f_{44} > 0.23 \pm 0.05$  and  $O:C > 0.61 \pm 0.05$ . We hypothesize that there may  
26 be a parameterizable ‘maximum oxidation level’ for cloud water above which functional  
27 group fragmentation is dominant. These experiments are among the first to quantify organic  
28 mass production in ambient cloud water and employ the most atmospherically relevant  
29 oxidant concentrations to date.  
30

31 **Keywords:** Secondary organic aerosol; Cloud chemistry; Photooxidation; Aqueous.  
32  
33  
34  
35  
36  
37  
38  
39  
40  
41  
42  
43  
44  
45  
46  
47

---

\* Corresponding author. Tel: 1-720-485-9191  
E-mail address: mishaschurman.ms@gmail.com

## 48 INTRODUCTION

49

50 Particles in the atmosphere have wide-ranging health, environmental, and climate  
51 effects, many of which are attributed disproportionately to fine-mode ( $<2.5\mu\text{m}$  diameter)  
52 secondary organic aerosols (SOA). While gas-phase chemistry and subsequent partitioning  
53 was traditionally thought to be the main source of SOA, this mechanism fails to predict the  
54 amount, content (Chang and Thompson, 2010), distribution (Heald et al., 2005; Tsigaridis &  
55 Kanakidou, 2003), and size of ambient particles (Blando and Turpin, 2000), where ‘ambient  
56 particles’ are defined as particles found in ambient air; recent work suggests that up to ~50%  
57 of total SOA mass is formed through aqueous reactions (Ervens et al. 2011). However, most  
58 of our knowledge of SOA formation is based on laboratory investigations of very simple  
59 solutions that aid mechanistic understanding of aqueous oxidation but may not accurately  
60 reflect the influence of the complex ambient matrix (defined as the mixture of compounds  
61 comprising a cloud water sample) on organic chemistry (Boris et al., 2014). This paper seeks  
62 to investigate the relevance of previous aqueous SOA (aqSOA) studies by photo-oxidizing  
63 ambient cloud water (defined as cloud water collected in-situ from natural clouds) with and  
64 without addition of common aqSOA precursors; while the complexity of the cloud water  
65 matrix prevents precise mechanistic determination, general reaction regimes may be inferred  
66 in terms of changes in organic mass, composition, and level of oxidation. Such an approach  
67 can be very useful, especially given our incomplete understanding of the organic composition  
68 of ambient aerosols and cloud water.

69 Important aqSOA precursors include anthropogenic and biogenic volatile organic  
70 compounds (A- and B-VOCs, respectively) oxidized to form products such as glyoxal,  
71 methylglyoxal, pinonaldehyde, and pinic/pinonic acids (Fu et al., 2008; Atkinson and Arey,  
72 2003). Pinonic acid is used herein to represent oxidation products of  $\alpha$ -pinene which may be  
73 absorbed from cloud interstitial spaces (Jang, 1999; Kanakidou et al., 2005), as a proxy for

74 total interstitial BVOC absorption. We hypothesize that this addition of organic material, and  
75 its relatively low oxidation level relative to the original cloudwater, may augment and/or  
76 sustain aqSOA formation as found in the previous works (see discussion in Results Section).

77 The hydroxyl radical (OH) is a dominant aqueous oxidant (Herrmann et al., 2010). OH  
78 can be absorbed from the gas phase, but is also formed significantly (up to ~33%; Ervens et  
79 al. 2003; Arakaki & Faust 1998) by aqueous photolysis of hydrogen peroxide ( $\text{H}_2\text{O}_2$ ; Zellner  
80 et al. 1990; Graedel & Goldberg 1983), iron(III) hydroxide ( $\text{Fe}(\text{OH})^{2+}$ ; Arakaki and Faust,  
81 1998; Weschler et al., 1986), nitrous acid (HONO; Arakaki and Faust, 1998; Fischer and  
82 Warneck, 1996), and nitric acid ( $\text{HNO}_3$ ; Graedel and Weschler, 1981);  $\text{H}_2\text{O}_2$  and  $\text{HO}_x$  ( $\text{HO}+$   
83 hydroperoxyl radical,  $\text{HO}_2$ ) may also undergo regenerative chemistry involving organic  
84 radicals (Anastasio et al., 1994; Valverde-Canossa et al., 2005). Direct organic photolytic  
85 oligomerization and lysing reactions are also possible and may form OH in turn through  $\text{HO}_x$   
86 cycling, though organic radical-initiated reactions are generally not competitive with OH  
87 oxidation (Guzman et al., 2006, Tan et al. 2010; Tan et al. 2012). Fenton reactions may  
88 produce OH in the absence of ultraviolet light (UV; Arakaki and Faust, 1998; Deguillaume et  
89 al., 2005), which motivated our no-UV control experiments. Reaction with OH and  
90 photolysis are important aqueous pinonic acid sinks (Lignell et al., 2013).

91 The current understanding of aqSOA mechanisms, products, and yields is based largely  
92 on laboratory investigations of simple solutions with unrealistically high oxidant and reactant  
93 concentrations. Solute concentration strongly influences aqSOA yield and product type, with  
94 higher concentrations forming higher molecular weight compounds such as oligomers  
95 through organic self-reaction, while lower fog- and cloud-relevant concentrations often form  
96 organic acids (Lim et al. 2010; Tan et al. 2009). Solute composition is also important, as  
97 organic radical formation, hygroscopicity, acidity, and component distribution in the droplet  
98 (e.g. organic shells) may all affect aqSOA yields (Ervens and Volkamer, 2010; Galloway et

99 al., 2011; Volkamer et al., 2009).

100 Simple solutions of common SOA precursors, while more mechanistically constrained,  
101 may not accurately represent the complex chemistry in real atmospheric waters. This work  
102 explores how photooxidation influences ambient cloud water organic mass, average oxidation  
103 state, and reaction regimes (such as functionalization versus fragmentation), and investigates  
104 relationships between oxidative progression and solution characteristics such as total organic  
105 carbon (TOC) and oxidation level (oxygen-to-carbon elemental ratios (O:C),  $f_{44}$ , etc.). The  
106 experiments presented here compare bulk photo-oxidations of real cloud water and cloud  
107 water with added pinonic acid (simulating ‘biogenic precursor’ addition), all featuring among  
108 the most atmospherically relevant reactant and oxidant concentrations to date.

109

## 110 **METHODS**

111

112 The Aerodyne High Resolution Time-of-Flight Aerosol Mass Spectrometer (AMS)  
113 measures the size and composition of submicron non-refractory particles with high time- and  
114 mass-resolution and has been described elsewhere (Decarlo et al., 2006); vaporization and  
115 electron impact ionization therein cause molecular fragmentation, complicating identification  
116 of specific reaction products as discussed later. AMS data quality is constrained by rigorous  
117 calibration; a LI-COR LI-820 monitor before the inlet constrained  $\text{CO}_{2(g)}$  contributions to the  
118  $\text{CO}_2^+$  fragment. The ToF-AMS DAQ v 1.7.2B software was used for data acquisition; data  
119 analysis utilized the SQUIRREL (v1.52K) and PIKA (v1.11K, Sueper et al., 2011) tools in  
120 Igor Pro 6.22A (WaveMetrics Inc., Lake Oswego, OR). Measurements near the AMS  
121 detection limit were not altered or eliminated. Particulars can be found in the supplement.

122 Briefly, bulk solutions are oxidized in a beaker-like vessel using UV light (Green  
123 Killing Machine, Model AAUV09W-UVC, 254 nm peak wavelength) and hydrogen peroxide,  
124 atomized (TSI 3076), dried (mimicking particle formation from cloud droplet evaporation,  
125 Lee et al. 2011; El Haddad et al. 2009; De Haan et al. 2009; Loeffler et al. 2006), and

126 analyzed via AMS with 1-minute time resolution after Lee et al. (2011b), with the addition of  
127 (a) a continuously circulating water sheath around the reaction vessel for temperature control  
128 at ~22 C (AquaBox fish-tank pump, Model AA2005), (b) a water trap after the atomizer  
129 (intermittent visible water at the atomizer outlet, see supplement), and (c) a  $^{210}\text{Po}$  neutralizer  
130 between the driers and the AMS (Figure 1; Boris et al., 2014). Photo-oxidations lasted  
131 between ~30-140 minutes; control experiments ran for ~20 minutes (see supplement). Photo-  
132 reactor and stir bar were cleaned with deionized water and air-dried between each experiment.

133 Five ambient cloud samples from Mt. Tai, China were used (named T1-4 and T7); T7  
134 was oxidized with (T7PA) and without (T7) added pinonic acid (30  $\mu\text{M}$ ). See supplement  
135 section S3 and Shen (2011) for sampling, storage, and composition details. Generally,  
136 pinonic acid (PA) concentrations in cloud water are unknown; due to similarity between  
137 yields of pinonic acid (~0.1-1%), pinic acid (~1-5%), and pinonaldehyde (~0-1%) from  $\alpha$ -  
138 and  $\beta$ -pinene aqueous oxidation (Yu et al., 1999), we conjecture that ambient pinonic acid  
139 concentrations might be similar to those of pinic acid and pinonaldehyde (0.11 and 0.8  $\mu\text{M}$ ,  
140 respectively; van Pinxteren et al., 2005). Species concentrations are normalized by total  
141 sulfate to correct noise from fluctuating atomizer output (supplement Sections S2 and S4 for  
142 detailed methodological assessment).

143 Hydroxyl radicals were produced via photolysis of hydrogen peroxide (300  $\mu\text{M}$ ). The  
144 resulting  $[\text{OH}]_{\text{aq}}$  was quantified in the cloud samples (and deionized water for comparison  
145 purposes) via absorption spectroscopy of 202 nm light by salicylic acid, which was below  
146 detection limit in the cloud samples, preceded by reverse phase liquid chromatography  
147 separation to prevent measurement interference from other absorbing species (N=3;  
148 accounting for salicylic acid direct photolysis and using the second-order rate constant of  $1.6$   
149  $\times 10^{10} \text{ M}^{-1}\text{s}^{-1}$ ; see Boris et al., 2014); these measurements yielded  $1.9 \times 10^{-14} \text{ M OH}$  within  
150 the cloud water and  $5.4 \pm 0.5 \times 10^{-14} \text{ M OH}$  within pure water (Boris et al., 2014).

151 Before hydrogen peroxide addition, these ambient cloud samples contained  $[\text{HOOH}]_{\text{aq}}$   
152 = 1-100  $\mu\text{M}$  (Shen 2011); assuming that the rate constant for OH production by hydrogen  
153 peroxide photolysis used herein holds in the complex ambient matrix, and that OH-organic  
154 reaction rate increases are linear within the given  $[\text{OH}]$  concentration range, this initial  
155 hydrogen peroxide content produces concentrations of  $[\text{OH}]_{\text{aq}} \approx 10^{-15}$ – $10^{-14}$  M which, adding  
156 the supplementary hydrogen peroxide, yields  $[\text{OH}]_{\text{aq}} \approx 10^{-14}$  M. Ambient cloud  $[\text{OH}]$  is  
157 difficult to measure, but has been estimated at  $\sim 10^{-14}$  to  $10^{-13}$  M (Arakaki & Faust 1998;  
158 Ervens et al. 2003; Ervens & Volkamer 2010; Jacob 1986; Deguillaume et al. 2005), 1-2  
159 orders of magnitude lower than used in similar laboratory experiments (Lim et al. 2010; Lee  
160 et al. 2011a). These solutions are among the most atmospherically relevant used in the  
161 laboratory to date; methodological assessment of evaporative cycling effects and OH  
162 exposure is discussed further in supplement Section S4.

## 163 164 **RESULTS**

### 165 166 ***Control Experiments***

167 Control experiments evaluated contributions of dark (noUV/noH<sub>2</sub>O<sub>2</sub>, ‘baseline’),  
168 photolytic (+UV/noH<sub>2</sub>O<sub>2</sub>), and hydrogen peroxide-initiated (noUV/+H<sub>2</sub>O<sub>2</sub>) reactions to  
169 aqSOA. For ambient cloud samples from Mt. Tai, China (T1-T7), neither organic mass (Total  
170 Organics/Total Sulfate, normalized to t = 0) nor oxidation indicator *f*<sub>44</sub> (mass at *m/z* 44/Total  
171 Organics, normalized to t = 0) change significantly under any of the control conditions,  
172 indicating probable dominance of OH-initiated reactions in the full photo-oxidations (Figure  
173 S1).

### 174 ***Ambient Cloud Water Photo-oxidation***

175 Organic mass (Org/SO<sub>4</sub>), O:C, and *f*<sub>44</sub> (*m/z* 44/Org) for ambient cloud water photo-  
176 oxidation experiments are shown in Figure 2; sample volume limited the duration of T1.  
177 Figure 3 shows timelines of CO<sub>2</sub><sup>+</sup>/SO<sub>4</sub> (Takegawa et al., 2007), C<sub>2</sub>H<sub>3</sub>O<sup>+</sup>/SO<sub>4</sub> (at *m/z* 43,

178 indicating carbonyl content; McLafferty and Turecek, 1993), and  $C_2H_3O^+/C_4H_7^+$  (ratio of  
179 most to least oxidized fragment at  $m/z$  43), as well as  $CH_2O_2^+/SO_4$ , which is, nominally,  
180 formic acid (a major oxidation product of glyoxal compounds; see supplement S4 for caveats  
181 to formic acid quantification via AMS). Although determination of specific oxidative  
182 mechanisms is precluded by both fragmentation in the AMS and the complexity of the  
183 ambient cloud water composition, possible functionalization pathways are examined using  
184 delta analysis and movement in van Krevelen space.

185 Experiments may be grouped according to their oxidative behavior into two groups. In  
186 Functionalization-Dominant Regime 1 (T1 and T3) samples first gain, then lose organic mass,  
187 and gain, then sustain  $CO_2^+$  fragments &  $f_{44}$  representing carboxylic acids (and other  
188 oxygenated fragments, as seen in difference spectra Figure S4) indicating retention of  
189 oxygenated products. In contrast, in Fragmentation-Dominant Regime 2 (T4 and T7), organic  
190 mass is unchanging for the first 30-60 minutes and decreasing significantly thereafter; there is  
191 no increase in organic mass as seen in Regime 1. Additionally, there is early formation, then  
192 pronounced loss of oxygenated fragments; together, these suggest mass loss due to  
193 functional-group fragmentation. Evidence supporting these regime groupings is detailed in  
194 the following sections.

### 195 ***Functionalization-Dominant Regime 1: Oxygenated products maintained***

196 For the T1 and T3 experiments with ambient Mt. Tai cloud water,  $Org/SO_4$  increases  
197 for the first 15-30 minutes (to ~110% and ~140% of the  $t = 0$  value, respectively), then  
198 decreases for the remainder of the experiment. Oxidation proxy  $f_{44}$  and carboxylic acid  
199 indicator  $CO_2^+/SO_4$  increase for the first 15-30 minutes (T1 and T3); the T3 experiment ran  
200 longer and these proxies were observed to stay constant thereafter through 90 minutes.  
201 Reduction in  $C_2H_3O^+/SO_4$  beginning at ~10 min may result from oxidation of carbonyls to  
202 carboxylic acids; contemporaneous rapid oxidation of  $C_xH_y$  species is indicated by an

203 *increase* in  $C_2H_3O^+/C_4H_7^+$  (from ~10-25 minutes in T1 and ~6-16 minutes in T3), indicating  
204 that  $C_4H_7^+$  must be decreasing faster than  $C_2H_3O^+$ . After ~10-15 minutes, both  $C_2H_3O^+/SO_4$   
205 and  $C_2H_3O^+/C_4H_7^+$  decrease, indicating a net loss of hydrocarbons and carbonyls. Most  
206 chemical change discernable via AMS occurs in the first half-hour or so of oxidation.

207 A picture emerges of rapid oxidation of hydrocarbons to carbonyls and carbonyls to  
208 carboxylic acids, consistent with other similar studies (Lee et al., 2011a). This is supported  
209 by difference spectra for the major organic fragment ‘families’ (CH, CHO,  $CHO_{n>1}$ ; Figure  
210 S4), in which CH fragments decrease and CHO and  $CHO_{n>1}$  generally increase, especially at  
211  $m/z$  44 ( $CO_2^+$ ); slight decreases in some CHO fragments commensurate with increases in  
212  $CHO_{n>1}$  fragments at the same  $m/z$  reflect oxidation of carbonyls to carboxylic acids. These  
213 patterns can also be seen in delta mass spectral analysis (see supplement S6 and discussion  
214 below).

215 The ‘formic acid’ fragment increases in both T1 and T3, suggesting that volatile formic  
216 acid and/or molecules fragmenting to  $CH_2O_2^+$  are early-generation products. Formic acid  
217 production occurs in aqueous oxidations of glyoxal and is responsible for the eventual loss of  
218 total organic mass in those experiments due to its volatility and resultant loss during drying  
219 (Lee et al. 2011b); the same mechanism could also be at play in T1 and T3, accompanied by  
220 decomposition of oxidized molecules to hydrocarbons and other non-formic-acid volatile  
221 fragments that are lost during evaporation as evidenced above.

222 Fragmentation may be further explored through the slope of movement in van Krevelen  
223 space (O:C vs H:C, see Figure 4) as an isolated particle/droplet population reacts (e.g. Heald  
224 et al. 2010). T1 and T3, though with different compositions and therefore different starting  
225 positions in the diagram, move similarly toward enhanced O:C with slopes near -0.3, which is  
226 between  $m=-0.5$ , indicative of either carboxylic addition with C-C bond breakage or addition  
227 of both acid and alcohol/peroxide functional groups without fragmentation (total 1 H loss per



228 2 O addition, Figure 4), and  $m=0$ , indicative of alcohol or peroxide addition. Delta analysis  
229 can reveal which of these mechanisms are important (see supplement S6), as functional group  
230 composition dictates fragment  $m/z$  values such that delta ( $\Delta$ ) =  $m/z - 14n + 1$ , where  $n$  is the  
231 number of methylene ( $\text{CH}_2$ ) bridges on the functional group (Table S3, McLafferty and  
232 Turecek, 1993). For T1 and T3, reductions in  $\Delta_{-2}$ ,  $\Delta_0$ , and  $\Delta_6$ , which may be partially  
233 attributed to alcohols (Table S3), suggests at least that alcohol addition does not exceed  
234 consumption of unsaturated hydrocarbons and carbonyls (at  $\Delta_0$ ,  $\Delta_{-2}$ ), and implies that  
235 carboxylic addition with C-C fragmentation is a more important reaction pathway for this  
236 dataset; this fragmentation may also contribute to the organic mass loss observed if the non-  
237 functionalized fragments are volatile. These slopes are similar to gas-phase chamber  
238 oxidations of both generated and ambient OA (Chhabra et al., 2011; Lambe et al., 2011; Ng  
239 et al., 2011).

240 In contrast to experiments T1 and T3, there is relatively little change in composition  
241 during oxidation of T2. Org/SO<sub>4</sub> is unchanging for the first ~20 minutes, with subsequent  
242 decrease (Figure 2).  $f_{44}$  increases slightly, but given that  $\text{CO}_2^+/\text{SO}_4$  and  $\text{C}_2\text{H}_3\text{O}^+/\text{SO}_4$  are  
243 constant (statistically, with a slight  $\text{CO}_2^+/\text{SO}_4$  decrease toward the end), this is more a  
244 function of total organic mass loss than proportional gains at  $m/z$  44;  $\text{C}_2\text{H}_3\text{O}^+/\text{C}_4\text{H}_7^+$  increases  
245 in the first ~20 minutes, indicating decrease in hydrocarbon  $\text{C}_4\text{H}_7^+$ , which is echoed in  
246 general (though not universal)  $\text{C}_x\text{H}_y$  reduction in the difference mass spectrum (Figure S4).  
247 However, clear progression toward higher O:C in the van Krevelen diagram (Figure 4)  
248 implies that oxygenated compounds are gained; family difference spectra reveal increases in  
249 fragments  $> m/z$  50 from the oxidized  $\text{CHO}_{n>1}$  family, though fragment abundance changes  
250 are small compared to the other experiments. The progression to highly oxidized ( $\text{CHO}_{n>1}$ )  
251 fragments  $> m/z$  50 and decreased  $\text{CO}_2^+$  suggests higher product stability against  
252 fragmentation within the AMS. By all measures, T2 started at a higher level of oxidation than

253 the other experiments, which may reduce its ability to participate in further oxidation  
254 reactions, if, as discussed later, there exists a ‘maximum’ level of average solution oxidation  
255 above which subsequent organic mass formation is hindered.

### 256 ***Fragmentation-Dominant Regime 2: Oxygenated fragments lost***

257 In Mt. Tai cloud water experiments T4 and T7, organic mass is unchanging for the first  
258 30-60 minutes and decreasing significantly thereafter; there is no increase in organic mass.  
259 Figure 2 shows that the normalized amount of Org/SO<sub>4</sub> does not increase as in Regime 1, but  
260 only decreases after the beginning of photooxidation at t=0. There is some ‘noise’ in this data,  
261 which is discussed thoroughly in Supplement Section S2: *Description of variable atomizer*  
262 *output*; nonetheless, the normalized Org/SO<sub>4</sub> value begins to fall below 1 immediately after  
263 t=0 and ‘noise’ does not exceed 1.09 in either T4 and T7, nor does it show the clear  
264 increasing, then decreasing trends seen in T1 and T3. This is in contrast to maximum  
265 normalized Org/SO<sub>4</sub> of 1.43 and 1.12 in T3 and T1.

266 O:C and *f*<sub>44</sub> generally increase for the first 15-30 minutes and then decrease (Figure 2).  
267 T4 and T7 show early formation, then pronounced loss (beginning at ~35-40 min) of  
268 CO<sub>2</sub><sup>+</sup>/SO<sub>4</sub>; the first half-hour looks similar to Regime 1 (T1 and T3), with fragment ratios  
269 indicating progressive carbonyl, then carboxylic acid increases (as above), but is followed by  
270 strong decreases in C<sub>2</sub>H<sub>3</sub>O<sup>+</sup> (to ~50% of t=0 value in T4) and CO<sub>2</sub><sup>+</sup> (max ~200% at t=20 min  
271 in T4; back to starting value by experiment’s end). This is in clear contrast to Regime 1, in  
272 which oxygenated products were sustained throughout the experiments. The formic acid  
273 fragment is near zero in T4 and with some mass in T7, but shows no significant trend in  
274 either time series. The fragment-family difference spectra show universal hydrocarbon (CH)  
275 loss and decreased C<sub>2</sub>H<sub>3</sub>O<sup>+</sup> (*m/z* 43), with small increase or decrease in experiment-end CO<sub>2</sub><sup>+</sup>  
276 in T4 and T7, respectively, consistent with the fragment timelines.

277 In the van Krevelen diagram, T4 and T7 initially progress toward higher O:C and lower  
278 H:C as would be expected of the carbonyl and carboxylic acid formation inferred from the  
279 fragment ratios above, but then move back toward lower O:C and higher H:C as photo-  
280 oxidation progresses; along with the loss of organic mass and oxygenated fragments, this  
281 suggests that oxidized functional groups are chemically fragmented in solution and lost  
282 during atomization/drying, leaving behind the hydrocarbon ‘backbones’ of the molecules.

### 283 *Ambient Cloud Water Photo-oxidations with Added Pinonic Acid Precursor*

284 In contrast to T7 and the other cloud samples, aqueous photo-oxidation of T7PA (T7  
285 with added pinonic acid) has fairly constant Org/SO<sub>4</sub>, f<sub>44</sub>, and CO<sub>2</sub><sup>+</sup>/SO<sub>4</sub> (Figures 2 & 3).  
286 C<sub>2</sub>H<sub>3</sub>O<sup>+</sup>/C<sub>4</sub>H<sub>7</sub><sup>+</sup> and moderate C<sub>2</sub>H<sub>3</sub>O<sup>+</sup>/SO<sub>4</sub> increases imply hydrocarbon loss and carbonyl  
287 gains that are corroborated in the difference spectra (Figures 3 & S4). In the van Krevelen  
288 diagram, T7PA resides in a region similar to T7 (Figure 4); T7PA’s lower starting O:C  
289 implies that unaltered T7 is on average even more functionalized than pinonic acid, which  
290 contains both a carboxylic acid and a carbonyl group. While T7PA develops slightly higher  
291 O:C and lower H:C with time (Figure 4), these changes lack the amplitude and organization  
292 seen in T7 and the other samples in van Krevelen space. The family difference spectra reveal  
293 addition of CHO and CHO<sub>n>1</sub> fragments (including acids and carbonyls); delta analysis also  
294 indicates carbonyl formation in increasing Δ<sub>2</sub> (alkyls, saturated carbonyls).

295 T7PA differs from the unaltered sample (T7) in a number of ways. Addition of pinonic  
296 acid reduces the initial level of solution oxidation and pH, enhances formation of carbonyls  
297 and larger, more stable molecules (increased CHO and CHO<sub>n>1</sub> > m/z 50 fragments), and  
298 does not contain ‘formic acid’ signal; organic mass is maintained instead of decreased. This  
299 mass retention may be influenced by increasing total organic carbon (regardless of organic  
300 composition), as added PA constitutes a 57% increase in TOC for T7PA; increasing  
301 precursor concentration often results in higher mass production as hypothesized for this

302 experiment and discussed in the introduction. Chamber pinonic acid-OH oxidations have  
303 resulted in highly functionalized, low-volatility molecules (Müller et al., 2012) similar to  
304 unpublished pinonic acid aqueous photo-oxidations in our laboratory (identified via ESI-ToF-  
305 MS). Our observation of dominant carbonyl formation compares favorably to single-  
306 precursor pinonic acid photo-oxidations in our photoreactor (Schurman, 2014) and other  
307 aqueous *cis*-pinonic acid (Aljawhary et al., 2016) and  $\alpha$ -pinene oxidations (Bleier and Elrod,  
308 2013); however,  $\alpha$ -pinene products appear to vary with OH exposures, with dominant  
309 carboxylic acid formation and little change in  $f_{43}$  under high-OH conditions (George and  
310 Abbatt, 2010), and carbonyl formation in the form of acetone, formaldehyde, formic acid, etc.  
311 under lower [OH] (Nozière et al., 1999); Lignell et al. (2013) show that direct aqueous PA  
312 photolysis is a minor mechanism. Since our precursor, pinonic acid, is a fairly early-  
313 generation  $\alpha$ -pinene product and our equivalent OH exposure is approximately 1-to-1 with  
314 ambient exposure, our samples experience only a few hours of ageing compared to the days-  
315 to-weeks exposures in many chamber experiments that observe oxidation to carboxylic acids  
316 (which may be less relevant to cloud lifetimes, as well).

317 These results suggest that biogenic precursor addition may suppress or delay the  
318 oxidative mass loss of organics observed in aqueous photo-oxidations of unaltered cloud  
319 samples, and that total organic mass formation or loss may be related to the average level of  
320 solution oxidation. To investigate this, we evaluated the relationship between O:C,  $f_{44}$ , and  
321 the incremental change in Org/SO<sub>4</sub>, finding that mass decreases begin only when O:C > 0.61  
322  $\pm 0.05$ ,  $f_{44} > 0.23 \pm 0.05$  and, generally,  $f_{43} < \sim 0.08 \pm 0.02$  (T1 organic mass decreases occur  
323 with  $f_{43}$  steady at  $\sim 0.084$ , but this holds true for all other ambient-cloud experiments).  
324 Because multiple fragments of different degrees of oxidation contribute to both  $m/z$  43 and 44,  
325 the  $f_{44}$  and  $f_{43}$  'break-over points' (at which organic mass begins to decrease) may not be as  
326 useful (or universal) as that of O:C.

327 **DISCUSSION**

328  
329 Two patterns emerge in these ambient cloud water photo-oxidations: in the first  
330 (Functionalization-Dominant Regime 1; T1 & T3), organic mass is gained, then lost, and  
331 sustained increases in highly oxidized fragments indicate overall organic acid formation;  
332 decrease in hydrocarbon fragments and production of the potentially-volatile  $\text{CH}_2\text{O}_2^+$   
333 fragment ('formic acid,' which partitions mostly to the gas phase; see discussion below and  
334 Liu et al., 2012), suggests that volatilization of these species may contribute to mass decrease.  
335 In Fragmentation-Dominant Regime 2 experiments (T4 & T7), oxidized fragments decrease  
336 as mass decreases, indicating that oxidized functional groups are fragmented and lost to  
337 evaporation.

338 pH varies significantly between functionalization-dominant Regime 1 (pH~5.5-6.4) and  
339 fragmentation-dominant Regime 2 (pH~4.6-4.7; Figure 5). These pH variations may drive  
340 formic acid production (or at least particle partitioning, as explained below), with substantial  
341 'formic acid' fragments occurring only in samples with pH > ~5.5 (T1 and T3); T3 had the  
342 highest pH, and also the greatest increase in nominally formic acid fragments, while T1 had a  
343 lower pH and lower formic acid concentrations. In Regime 2, for which solution pH values  
344 were  $\leq 4.74$ , little or no formic acid was detected. This is logical, since the Henry's Law  
345 partitioning of formic acid is driven toward the particle phase by higher pH (Liu et al. 2012b).  
346 Furthermore, particulate formic acid is enhanced significantly beyond the Henry's Law  
347 equilibrium in recent studies of urban-influenced particles (including one in Beijing, China;  
348 Liu et al. 2012b; Wang et al. 2007); possible explanations include varying pH between  
349 particle bulk and surface (such that bulk measurements do not capture the pH relevant to the  
350 reaction) and variability in LWC (due to varying degrees of drying over the course of our  
351 experiments; Liu et al. 2012). Also, cations such as  $\text{Ca}^{2+}$  can undergo acid-base reactions that  
352 lead to formic acid retention in the particle phase (Fornaro and Gutz, 2003; Löflund et al.,

353 2002; Malm et al., 2009); this chemistry has been indirectly evidenced in one case,  
354 coincidentally involving long-range transport of formic acid and dust from China (Kawamura  
355 et al., 2012).

356 Initial solution O:C (ranging from 0.35-0.53) does not vary significantly between  
357 Regime 1 and Regime 2, but Regime 1 reactions generally end at higher O:C (0.69-0.95,  
358 versus ~0.64 for Type 2) due to retention of oxygenated functional groups. On average,  
359 Regime 1 samples had higher starting concentrations of TOC, hydrogen peroxide, and other  
360 solutes, which may explain the tendency toward higher functionalization and therefore mass  
361 retention as described in the introduction; however, these concentration differences were not  
362 significant using the Mann-Whitney-Wilcoxon rank-sum test, and samples with equivalent  
363 TOC concentrations were found between the types (Figure 5). Meteorological back-  
364 trajectories indicate the same source region (the heavily populated, industrial South China  
365 Plain) for air masses involved in all cloud samples (Shen, 2011).

366 Because it is difficult to both collect and to fully and continuously speciate cloud water,  
367 many studies use simple surrogate cloud water solutions to better constrain yields and  
368 reaction pathways, impeding direct comparison of our findings to the literature. However,  
369 using methods quite similar to those herein, though with higher OH concentrations, Lee et al.  
370 (2012) noted that ambient cloud samples formed highly oxidized SOA similar to that found  
371 in ambient aerosols, with organic mass production maxima of ~110-130% of the initial value  
372 followed by decreasing mass, similar to this work (Lee et al. 2012). A model simulation of  
373 rural SOA<sub>(aq+g)</sub> with monoterpene emissions suggested ~21-25% SOA mass increase in the  
374 first ~1 hour when clouds (pH = 4.5) were added, with a slight reduction in mass after ~2  
375 hours much like our experiments (Chen et al. 2007a), while urban aerosol with anthropogenic  
376 emissions (NO<sub>x</sub>, aromatics, alkanes) increased about 10% with no subsequent decrease; as in  
377 this work, model addition of less oxidized precursors appeared to augment and/or sustain

378 aqSOA increases, supporting the hypothesis that pinonic acid addition would do the same.  
379 This hypothesis is also supported by a methacrolein cloud chamber photo-oxidation  
380 experiment involving a continuous interstitial source of less-oxidized precursors gases, which  
381 found a non-linear 47-fold increases in SOA mass sustained over a 22-hour experiment (El  
382 Haddad et al., 2009; Liu et al., 2009; Monod et al., 2007; Monod et al., 2000).

383 The methodology used herein does differ from ambient cloud processing in some ways.  
384 First, bulk-phase photooxidations disregard the continuous evaporation and re-condensation  
385 of cloud droplets, which influences evaporative losses and the cycling of interstitial VOCs,  
386 semi-volatile compounds, oxidants, and oxygen (O<sub>2</sub>) through the gas and droplet phases; this  
387 may in turn change aqSOA production (De Haan et al. 2009). With typical cloud droplet  
388 lifetimes of ~10 minutes (Roelofs and Kamphuis, 2009), a real cloud could undergo 12  
389 droplet cycles over the course of the experiments herein, implying a more continuous source  
390 of fresh interstitial VOCs than was recreated in our bulk-phase experiments. The bulk phase  
391 also limits gas exchange by eliminating droplet surface area at which exchange takes place:  
392 this lack of droplet surface area and attendant gas exchange may suppress both precursor  
393 absorption and volatile product evaporation (although drying before AMS analysis may offset  
394 the latter effect). The effect of evaporative cycling on aqSOA is a focus of current research.

395 In all, the bulk-aqueous/AMS analytical method is expected to lose a large and  
396 variable fraction of volatile oxidation products, the amount of which is highly dependent on  
397 the oxidation precursors and the volatility of their dominant products and is assumed to  
398 mimic the behavior of evaporating cloud droplets in the real atmosphere. Loss mechanisms  
399 may include a) evaporative loss during diffusion drying (Lee et al. 2011a), and b) evaporative  
400 loss during drying as the particle beam enters the high vacuum in the AMS (Zelenyuk et al.,  
401 2006). Mechanism a) might be more atmospherically relevant than b), but the current dataset  
402 cannot determine the contributions of each of these mechanisms to organic evaporative loss.

403 Also, while the continuously-mixed bulk solution methodology might prevent the phase  
404 separation observed in atmospheric aerosol water (e.g. organic shells, etc. which may change  
405 water evaporation and/or solute interaction and therefore reactivity), this effect is not thought  
406 to be important in cloud-relevant concentrations, where relative dilution leads to  
407 homogeneous mixing (Marcolli and Peter, 2005; Shiraiwa et al., 2013; Ziemann, 2010;  
408 Zuend et al., 2008). Under the assumption that the drying method used herein resemble those  
409 in the real atmosphere, the AMS does capture evaporated-cloud-relevant particulate organics.

410

## 411 CONCLUSIONS

412

413 This work explores how photooxidation influences ambient cloud water organic mass,  
414 oxidation state, and reaction regimes (such as functionalization versus fragmentation), and  
415 investigates relationships between oxidative progression and solution characteristics such as  
416 oxidation level (O:C) and initial total organic carbon (TOC). Unaltered ambient cloud water  
417 and ‘biogenically influenced’ ambient cloud water (via addition of pinonic acid) were photo-  
418 oxidized using UVC light and H<sub>2</sub>O<sub>2</sub> (to form OH), atomized, and dried to simulate the  
419 formation of aqSOA from clouds; these experiments are among the first to quantify organic  
420 mass production in ambient cloud water and employ the most atmospherically relevant  
421 oxidant concentrations to date.

422 All ambient samples (except the heavily oxidized T2) gained carbonyl, then organic  
423 acid fragments; two of the five samples increased in sulfate-normalized organic mass (to a  
424 maximum of 110-140% of the t=0 value). All samples eventually lost organic mass (reaching  
425 60-80% of their initial values), which is attributed to formation of volatile products in some  
426 cases (‘Functionalization-Dominant Regime 1’ evidenced by increasing ‘formic acid’  
427 fragment and sustained increases in CO<sub>2</sub><sup>+</sup>) and molecular decomposition leading to loss of  
428 oxidized functional groups in others (‘Fragmentation-Dominant Regime 2’; CO<sub>2</sub><sup>+</sup>, C<sub>2</sub>H<sub>3</sub>O<sup>+</sup>,



429 O:C, and total organic mass decrease simultaneously); Regime 1 had significantly higher pH  
430 than Regime 2, and generally (though not significantly) higher initial total organic carbon,  
431 consistent with literature results showing formation of more functionalized products in  
432 higher-concentration solutions. The rate of aqSOA production in unaltered cloud water  
433 decreases as oxygenation increases, until organic mass loss beginning at a consistent value of  
434  $O:C > 0.61 \pm 0.05$  and  $f_{44} > 0.23 \pm 0.03$  ( $f_{43} < 0.08 \pm 0.02$ ). This implies that there may be a  
435 parameterizable ‘maximum oxidation level’ for cloud water above which organic molecule  
436 fragmentation is dominant; however, we were unable to further mathematically describe the  
437 relationship between aqSOA production/loss rate and oxidation level. The consistency of this  
438 ‘maximum oxidation level’ within the Mt. Tai cloud water should prompt further  
439 investigation of cloud samples with varying composition and concentration. If widely  
440 applicable, such a parameterization would contribute greatly to the understanding and  
441 modeling of the secondary organic aerosols that contribute to so many important health,  
442 environmental, and climate effects. Lastly, addition of TOC in the form of pinonic acid to an  
443 ambient sample (meant to simulate interstitial biogenic precursor absorption) appeared to  
444 prevent organic mass loss and foster carbonyl formation, rendering the un-altered  
445 experiments herein lower-bound estimations of aqSOA formation in photooxidized clouds.

446  
447

#### 448 **ACKNOWLEDGMENTS**

449

450 Collection of Mt. Tai cloud samples was funded by NSF Grant ATM-0711102. Thanks  
451 to Taehyoung Lee and Xinhua Shen for collecting the Mt. Tai cloud samples used in this  
452 study, and Tao Wang, Wenxing Wang, and Xinfeng Wang for their help and partnership in  
453 planning the Mt. Tai cloud sampling campaign.

#### 454 **REFERENCES**

455

456 Aljawhary, D., Zhao, R., Lee, A.K.Y., Wang, C. and Abbatt, J.P.D. 2016. Kinetics,  
457 Mechanism, and Secondary Organic Aerosol Yield of Aqueous Phase Photo-oxidation of  
458  $\alpha$ -Pinene Oxidation Products. *The Journal of Physical Chemistry A* 120(9), pp. 1395–1407.

459 Anastasio, C., Faust, B. C. and Allen, J. M.: Aqueous phase photochemical formation of  
460 hydrogen peroxide in authentic cloud waters, *J. Geophys. Res.*, 99(D4), 8231,  
461 doi:10.1029/94JD00085, 1994.

462 Arakaki, T. and Faust, B. C.: Sources, sinks, and mechanisms of hydroxyl radical ( $\bullet$ OH)  
463 photoproduction and consumption in authentic acidic continental cloud waters from  
464 Whiteface Mountain, New York: The role of the Fe(r) (r = II, III) photochemical cycle, *J.*  
465 *Geophys. Res.*, 103(D3), 3487, doi:10.1029/97JD02795, 1998.

466 Atkinson, R. and Arey, J.: Atmospheric degradation of volatile organic compounds., *Chem.*  
467 *Rev.*, 103(12), 4605–38, doi:10.1021/cr0206420, 2003.

468 Blando, J. D. and Turpin, B. J.: Secondary organic aerosol formation in cloud and fog  
469 droplets: a literature evaluation of plausibility, *Atmos. Environ.*, 34(10), 1623–1632,  
470 doi:10.1016/S1352-2310(99)00392-1, 2000.

471 Bleier, D. B. and Elrod, M. J.: Kinetics and Thermodynamics of Atmospherically Relevant  
472 Aqueous Phase Reactions of  $\alpha$ -Pinene Oxide, *J. Phys. Chem. A*, 117(20), 4223–4232,  
473 2013.

474 Boris, A. J., Desyaterik, Y. and Collett, J. L.: How do components of real cloud water affect  
475 aqueous pyruvate oxidation?, *Atmos. Res.*, 143, 95–106,  
476 doi:10.1016/j.atmosres.2014.02.004, 2014.

477 Chang, J. L. and Thompson, J. E.: Characterization of colored products formed during  
478 irradiation of aqueous solutions containing H<sub>2</sub>O<sub>2</sub> and phenolic compounds, *Atmos.*

479 Environ., 44(4), 541–551, doi:10.1016/j.atmosenv.2009.10.042, 2010.

480 Chen, J., Griffin, R. J., Grini, A. and Tulet, P.: Modeling secondary organic aerosol formation  
481 through cloud processing of organic compounds, *Atmos. Chem. Phys.*, 7(20), 5343–5355,  
482 doi:10.5194/acp-7-5343-2007, 2007.

483 Chhabra, P. S., Ng, N. L., Canagaratna, M. R., Corrigan, A. L., Russell, L. M., Worsnop, D.  
484 R., Flagan, R. C. and Seinfeld, J. H.: Elemental composition and oxidation of chamber  
485 organic aerosol, *Atmos. Chem. Phys.*, 11(17), 8827–8845, doi:10.5194/acp-11-8827-2011,  
486 2011.

487 Collett, J. L., Daube, B. C., Gunz, D. and Hoffmann, M. R.: Intensive studies of Sierra  
488 Nevada cloudwater chemistry and its relationship to precursor aerosol and gas  
489 concentrations, *Atmos. Environ. Part A. Gen. Top.*, 24(7), 1741–1757, 1990.

490 Decarlo, P. F., Kimmel, J. R., Trimborn, A., Northway, M. J., Jayne, J. T., Aiken, A. C.,  
491 Gonin, M., Fuhrer, K., Horvath, T., Docherty, K. S., Worsnop, D. R. and Jimenez, J. L.:  
492 Field-deployable, high-resolution, time-of-flight aerosol mass spectrometer., *Anal. Chem.*,  
493 78(24), 8281–9, doi:10.1021/ac061249n, 2006.

494 Deguillaume, L., Leriche, M. and Chaumerliac, N.: Impact of radical versus non-radical  
495 pathway in the Fenton chemistry on the iron redox cycle in clouds., *Chemosphere*, 60(5),  
496 718–24, doi:10.1016/j.chemosphere.2005.03.052, 2005.

497 Ervens, B., George, C., Williams, J. E., Buxton, G. V., Salmon, G. A., Bydder, M.,  
498 Wilkinson, F., Dentener, F., Mirabel, P., Wolke, R. and Herrmann, H.: CAPRAM 2.4  
499 (MODAC mechanism): An extended and condensed tropospheric aqueous phase  
500 mechanism and its application, *J. Geophys. Res.*, 108(D14), 4426,  
501 doi:10.1029/2002JD002202, 2003a.

502 Ervens, B., Gligorovski, S. and Herrmann, H.: Temperature-dependent rate constants for  
503 hydroxyl radical reactions with organic compounds in aqueous solutions, *Phys. Chem.*  
504 *Chem. Phys.*, 5(9), 1811–1824, doi:10.1039/b300072a, 2003b.

505 Ervens, B. and Volkamer, R.: Glyoxal processing by aerosol multiphase chemistry: towards a  
506 kinetic modeling framework of secondary organic aerosol formation in aqueous particles,  
507 *Atmos. Chem. Phys.*, 10(17), 8219–8244, doi:10.5194/acp-10-8219-2010, 2010.

508 Fischer, M. and Warneck, P.: Photodecomposition of Nitrite and Undissociated Nitrous Acid  
509 in Aqueous Solution, *J. Phys. Chem.*, 100(48), 18749–18756, doi:10.1021/jp961692+,  
510 1996.

511 Fornaro, A. and Gutz, I. G. R.: Wet deposition and related atmospheric chemistry in the São  
512 Paulo metropolis, Brazil: Part 2—contribution of formic and acetic acids, *Atmos. Environ.*,  
513 37(1), 117–128, 2003.

514 Fu, T. M., Jacob, D. J., Wittrock, F., Burrows, J. P., Vrekoussis, M. and Henze, D. K.: Global  
515 budgets of atmospheric glyoxal and methylglyoxal, and implications for formation of  
516 secondary organic aerosols, *J. Geophys. Res.*, 113(D15), D15303,  
517 doi:10.1029/2007JD009505, 2008.

518 Galloway, M. M., Loza, C. L., Chhabra, P. S., Chan, A. W. H., Yee, L. D., Seinfeld, J. H. and  
519 Keutsch, F. N.: Analysis of photochemical and dark glyoxal uptake: Implications for SOA  
520 formation, *Geophys. Res. Lett.*, 38(17), n/a–n/a, doi:10.1029/2011GL048514, 2011.

521 George, I. J. and Abbatt, J. P. D.: Chemical evolution of secondary organic aerosol from OH-  
522 initiated heterogeneous oxidation, *Atmos. Chem. Phys.*, 10(12), 5551–5563,  
523 doi:10.5194/acp-10-5551-2010, 2010.

524 Graedel, T. E. and Goldberg, K. I.: Kinetic studies of raindrop chemistry: 1. Inorganic and

525 organic processes, *J. Geophys. Res.*, 88(C15), 10865, doi:10.1029/JC088iC15p10865,  
526 1983.

527 Graedel, T. E. and Weschler, C. J.: Chemistry within aqueous atmospheric aerosols and  
528 raindrops, *Rev. Geophys.*, 19(4), 505, doi:10.1029/RG019i004p00505, 1981.

529 Guzman, M. I., Colussi, A. J. and Hoffmann, M. R.: Photoinduced oligomerization of  
530 aqueous pyruvic acid., *J. Phys. Chem. A*, 110(10), 3619–26, doi:10.1021/jp056097z, 2006.

531 De Haan, D. O., Corrigan, A. L., Tolbert, M. A., Jimenez, J. L., Wood, S. E. and Turley, J. J.:  
532 Secondary organic aerosol formation by self-reactions of methylglyoxal and glyoxal in  
533 evaporating droplets., *Environ. Sci. Technol.*, 43(21), 8184–90, doi:10.1021/es902152t,  
534 2009.

535 El Haddad, I., Nieto-Gligorovski, L., Michaud, V., Temime-Roussel, B., Quivet, E.,  
536 Marchand, N., Sellegri, K. and Monod, A.: In-cloud processes of methacrolein under  
537 simulated conditions – Part 2: Formation of secondary organic aerosol, *Atmos. Chem.*  
538 *Phys.*, 9(14), 5107–5117, doi:10.5194/acp-9-5107-2009, 2009.

539 Heald, C. L., Jacob, D. J., Park, R. J., Russell, L. M., Huebert, B. J., Seinfeld, J. H., Liao, H.  
540 and Weber, R. J.: A Large Organic Aerosol Source in the Free Troposphere Missing from  
541 Current Models, 2005.

542 Heald, C. L., Kroll, J. H., Jimenez, J. L., Docherty, K. S., DeCarlo, P. F., Aiken, A. C., Chen,  
543 Q., Martin, S. T., Farmer, D. K. and Artaxo, P.: A simplified description of the evolution  
544 of organic aerosol composition in the atmosphere, *Geophys. Res. Lett.*, 37(8), L08803,  
545 doi:10.1029/2010GL042737, 2010.

546 Herrmann, H., Hoffmann, D., Schaefer, T., Brüner, P. and Tilgner, A.: Tropospheric aqueous-  
547 phase free-radical chemistry: radical sources, spectra, reaction kinetics and prediction

548 tools., *Chemphyschem*, 11(18), 3796–822, doi:10.1002/cphc.201000533, 2010.

549 Jacob, D. J.: Chemistry of OH in remote clouds and its role in the production of formic acid  
550 and peroxymonosulfate, *J. Geophys. Res.*, 91(D9), 9807, doi:10.1029/JD091iD09p09807,  
551 1986.

552 Jang, M.: Newly characterized products and composition of secondary aerosols from the  
553 reaction of  $\alpha$ -pinene with ozone, *Atmos. Environ.*, 33(3), 459–474, doi:10.1016/S1352-  
554 2310(98)00222-2, 1999.

555 Kanakidou, M., Seinfeld, J. H., Pandis, S. N., Barnes, I., Dentener, F. J., Facchini, M. C., Van  
556 Dingenen, R., Ervens, B., Nenes, a., Nielsen, C. J., Swietlicki, E., Putaud, J. P., Balkanski,  
557 Y., Fuzzi, S., Horth, J., Moortgat, G. K., Winterhalter, R., Myhre, C. E. L., Tsigaridis, K.,  
558 Vignati, E., Stephanou, E. G. and Wilson, J.: Organic aerosol and global climate  
559 modelling: a review, *Atmos. Chem. Phys.*, 5(4), 1053–1123, doi:10.5194/acp-5-1053-2005,  
560 2005.

561 Kawamura, K., IMAI, Y. and BARRIE, L.: Photochemical production and loss of organic  
562 acids in high Arctic aerosols during long-range transport and polar sunrise ozone depletion  
563 events, *Atmos. Environ.*, 39(4), 599–614, doi:10.1016/j.atmosenv.2004.10.020, 2005.

564 Kawamura, K., Matsumoto, K., Tachibana, E. and Aoki, K.: Low molecular weight (C1–C10)  
565 monocarboxylic acids, dissolved organic carbon and major inorganic ions in alpine snow  
566 pit sequence from a high mountain site, central Japan, *Atmos. Environ.*, 62, 272–280, 2012.

567 Kroll, J. H., Donahue, N. M., Jimenez, J. L., Kessler, S. H., Canagaratna, M. R., Wilson, K.  
568 R., Altieri, K. E., Mazzoleni, L. R., Wozniak, A. S., Bluhm, H., Mysak, E. R., Smith, J. D.,  
569 Kolb, C. E. and Worsnop, D. R.: Carbon oxidation state as a metric for describing the  
570 chemistry of atmospheric organic aerosol., *Nat. Chem.*, 3(2), 133–9,

571 doi:10.1038/nchem.948, 2011.

572 Lambe, A. T., Onasch, T. B., Massoli, P., Croasdale, D. R., Wright, J. P., Ahern, A. T.,  
573 Williams, L. R., Worsnop, D. R., Brune, W. H. and Davidovits, P.: Laboratory studies of  
574 the chemical composition and cloud condensation nuclei (CCN) activity of secondary  
575 organic aerosol (SOA) and oxidized primary organic aerosol (OPOA), *Atmos. Chem.*  
576 *Phys.*, 11(17), 8913–8928, doi:10.5194/acp-11-8913-2011, 2011.

577 Lee, A. K. Y., Hayden, K. L., Herckes, P., Leaitch, W. R., Liggio, J., Macdonald, A. M. and  
578 Abbatt, J. P. D.: Characterization of aerosol and cloud water at a mountain site during  
579 WACS 2010: secondary organic aerosol formation through oxidative cloud processing,  
580 *Atmos. Chem. Phys.*, 12(15), 7103–7116, doi:10.5194/acp-12-7103-2012, 2012.

581 Lee, A. K. Y., Herckes, P., Leaitch, W. R., Macdonald, A. M. and Abbatt, J. P. D.: Aqueous  
582 OH oxidation of ambient organic aerosol and cloud water organics: Formation of highly  
583 oxidized products, *Geophys. Res. Lett.*, 38(11), n/a–n/a, doi:10.1029/2011GL047439,  
584 2011a.

585 Lee, A. K. Y., Zhao, R., Gao, S. S. and Abbatt, J. P. D.: Aqueous-phase OH oxidation of  
586 glyoxal: application of a novel analytical approach employing aerosol mass spectrometry  
587 and complementary off-line techniques., *J. Phys. Chem. A*, 115(38), 10517–26,  
588 doi:10.1021/jp204099g, 2011b.

589 Lignell, H., Epstein, S., Marvin, M. and et al. 2013. Experimental and theoretical study of  
590 aqueous cis-pinonic acid photolysis. *The Journal of Physical Chemistry A* 2013 117 (48),  
591 12930-12945 DOI: 10.1021/jp4093018

592 Lim, Y. B., Tan, Y., Perri, M. J., Seitzinger, S. P. and Turpin, B. J.: Aqueous chemistry and  
593 its role in secondary organic aerosol (SOA) formation, *Atmos. Chem. Phys.*, 10(21),

594 10521–10539, doi:10.5194/acp-10-10521-2010, 2010.

595 Liu, J., Horowitz, L. W., Fan, S., Carlton, A. G. and Levy, H.: Global in-cloud production of  
596 secondary organic aerosols: Implementation of a detailed chemical mechanism in the  
597 GFDL atmospheric model AM3, *J. Geophys. Res.*, 117(D15), D15303,  
598 doi:10.1029/2012JD017838, 2012a.

599 Liu, J., Zhang, X., Parker, E. T., Veres, P. R., Roberts, J. M., de Gouw, J. A., Hayes, P. L.,  
600 Jimenez, J. L., Murphy, J. G., Ellis, R. A., Huey, L. G. and Weber, R. J.: On the gas-  
601 particle partitioning of soluble organic aerosol in two urban atmospheres with contrasting  
602 emissions: 2. Gas and particle phase formic acid, *J. Geophys. Res. Atmos.*, 117(D21), n/a–  
603 n/a, doi:10.1029/2012JD017912, 2012b.

604 Liu, Y., El Haddad, I., Scarfogliero, M., Nieto-Gligorovski, L., Temime-Roussel, B., Quivet,  
605 E., Marchand, N., Picquet-Varrault, B. and Monod, A.: In-cloud processes of methacrolein  
606 under simulated conditions – Part 1: Aqueous phase photooxidation, *Atmos. Chem. Phys.*,  
607 9(14), 5093–5105, doi:10.5194/acp-9-5093-2009, 2009.

608 Loeffler, K. W., Koehler, C. A., Paul, N. M. and De Haan, D. O.: Oligomer Formation in  
609 Evaporating Aqueous Glyoxal and Methyl Glyoxal Solutions, *Environ. Sci. Technol.*,  
610 40(20), 6318–6323, doi:10.1021/es060810w, 2006.

611 Löflund, M., Kasper-Giebl, A., Schuster, B., Giebl, H., Hitenberger, R. and Puxbaum, H.:  
612 Formic, acetic, oxalic, malonic and succinic acid concentrations and their contribution to  
613 organic carbon in cloud water., 2002.

614 Malm, W. C., Barna, M. G., Beem, K. B., Carrico, C. M., Collett Jr., J. L., Day, D. E.,  
615 Gebhart, K. A., Hand, J. L., Kreidenweis, S. M., Lee, T., Levin, E. J. T., McDade, C. E.,  
616 McMeeking, G. R., Molenaar, J. V., Raja, S., Rodriguez, M. A., Schichtel, B. A.,



617 Schwandner, F. M., Sullivan, A. P. and Taylor, C.: Rocky Mountain Atmospheric  
618 Nitrogen and Sulfur Study, Fort Collins, CO., 2009.

619 McLafferty, F. W. and Turecek, F.: Interpretation of Mass Spectra, 4th ed., University  
620 Science Books, Herndon, VA., 1993.

621 Monod, A., Chebbi, A., Durand-Jolibois, R. and Carlier, P.: Oxidation of methanol by  
622 hydroxyl radicals in aqueous solution under simulated cloud droplet conditions, *Atmos.*  
623 *Environ.*, 34(29-30), 5283–5294, doi:10.1016/S1352-2310(00)00191-6, 2000.

624 Monod, A., Chevallier, E., Durandjolibois, R., Doussin, J., Picquet-Varrault, B. and Carlier, P.:  
625 Photooxidation of methylhydroperoxide and ethylhydroperoxide in the aqueous phase  
626 under simulated cloud droplet conditions, *Atmos. Environ.*, 41(11), 2412–2426,  
627 doi:10.1016/j.atmosenv.2006.10.006, 2007.

628 Müller, L., Reinnig, M.-C., Naumann, K. H., Saathoff, H., Mentel, T. F., Donahue, N. M. and  
629 Hoffmann, T.: Formation of 3-methyl-1,2,3-butanetricarboxylic acid via gas phase  
630 oxidation of pinonic acid – a mass spectrometric study of SOA aging, *Atmos. Chem. Phys.*,  
631 12(3), 1483–1496, doi:10.5194/acp-12-1483-2012, 2012.

632 Ng, N. L., Canagaratna, M. R., Jimenez, J. L., Chhabra, P. S., Seinfeld, J. H. and Worsnop, D.  
633 R.: Changes in organic aerosol composition with aging inferred from aerosol mass spectra,  
634 *Atmos. Chem. Phys.*, 11(13), 6465–6474, doi:10.5194/acp-11-6465-2011, 2011.

635 Ng, N. L., Canagaratna, M. R., Zhang, Q., Jimenez, J. L., Tian, J., Ulbrich, I. M., Kroll, J. H.,  
636 Docherty, K. S., Chhabra, P. S., Bahreini, R., Murphy, S. M., Seinfeld, J. H., Hildebrandt,  
637 L., Donahue, N. M., DeCarlo, P. F., Lanz, V. a., Prévôt, a. S. H., Dinar, E., Rudich, Y.  
638 and Worsnop, D. R.: Organic aerosol components observed in Northern Hemispheric  
639 datasets from Aerosol Mass Spectrometry, *Atmos. Chem. Phys.*, 10(10), 4625–4641,

640 doi:10.5194/acp-10-4625-2010, 2010.

641 Nozière, B., Barnes, I. and Becker, K. H.: Product study and mechanisms of the reactions of  
642  $\alpha$ -pinene and of pinonaldehyde with OH radicals, *J. Geophys. Res.*, 104(D19), 23645,  
643 doi:10.1029/1999JD900778, 1999.

644 Oturan, M. A., Peiroten, J., Chartrin, P. and Acher, A. J.: Complete Destruction of p -  
645 Nitrophenol in Aqueous Medium by Electro-Fenton Method, *Environ. Sci. Technol.*,  
646 34(16), 3474–3479, doi:10.1021/es990901b, 2000.

647 van Pinxteren, D., Plewka, A., Hofmann, D., Müller, K., Kramberger, H., Svrčina, B.,  
648 Bächmann, K., Jaeschke, W., Mertes, S., Collett, J. L. and Herrmann, H.: Schmücke hill  
649 cap cloud and valley stations aerosol characterisation during FEBUKO (II): Organic  
650 compounds, *Atmos. Environ.*, 39(23-24), 4305–4320, doi:10.1016/j.atmosenv.2005.02.014,  
651 2005.

652 Schurman, M. I.: Characteristics, Sources, and Formation of Organic Aerosols in the Rocky  
653 Mountains, Colorado State University, Fort Collins, Colorado., 2014.

654 Shen, X.: Aqueous Phase Sulfate Production in Clouds at Mt. Tai in Eastern China, Colorado  
655 State University., 2011.

656 Takegawa, N., Miyakawa, T., Kawamura, K. and Kondo, Y.: Contribution of selected  
657 dicarboxylic and omega-oxocarboxylic acids in ambient aerosol to the m/z 44 signal of an  
658 Aerodyne aerosol mass spectrometer, *Aerosol Sci. Technol.*, 41, 418–437, 2007.

659 Tan, Y., Carlton, A. G., Seitzinger, S. P. and Turpin, B. J.: SOA from methylglyoxal in  
660 clouds and wet aerosols: Measurement and prediction of key products, *Atmos. Environ.*,  
661 44(39), 5218–5226, doi:10.1016/j.atmosenv.2010.08.045, 2010.

662 Tan, Y., Lim, Y. B., Altieri, K. E., Seitzinger, S. P. and Turpin, B. J.: Mechanisms leading to  
663 oligomers and SOA through aqueous photooxidation: insights from OH radical oxidation  
664 of acetic acid and methylglyoxal, *Atmos. Chem. Phys.*, 12(2), 801–813, doi:10.5194/acp-  
665 12-801-2012, 2012.

666 Tan, Y., Perri, M. J., Seitzinger, S. P. and Turpin, B. J.: Effects of precursor concentration  
667 and acidic sulfate in aqueous glyoxal-OH radical oxidation and implications for secondary  
668 organic aerosol., *Environ. Sci. Technol.*, 43(21), 8105–12, doi:10.1021/es901742f, 2009.

669 Tsigaridis, K. and Kanakidou, M.: Global modelling of secondary organic aerosol in the  
670 troposphere: a sensitivity analysis, *Atmos. Chem. Phys.*, 3(5), 1849–1869,  
671 doi:10.5194/acp-3-1849-2003, 2003.

672 Valverde-Canossa, J., Wieprecht, W., Acker, K. and Moortgat, G. K.: H<sub>2</sub>O<sub>2</sub> and organic  
673 peroxide measurements in an orographic cloud: The FEBUKO experiment, *Atmos.*  
674 *Environ.*, 39(23-24), 4279–4290, doi:10.1016/j.atmosenv.2005.02.040, 2005.

675 Volkamer, R. M., Ziemann, P. J. and Molina, M. J.: Secondary Organic Aerosol Formation  
676 from Acetylene (C<sub>2</sub>H<sub>2</sub>): seed effect on SOA yields due to organic photochemistry in the  
677 aerosol aqueous phase, *Atmos. Chem. Phys.*, 9(6), 1907–1928, doi:10.5194/acp-9-1907-  
678 2009, 2009.

679 Wang, Y., Zhuang, G., Chen, S., An, Z. and Zheng, A.: Characteristics and sources of formic,  
680 acetic and oxalic acids in PM<sub>2.5</sub> and PM<sub>10</sub> aerosols in Beijing, China, *Atmos. Res.*, 84(2),  
681 169–181, doi:10.1016/j.atmosres.2006.07.001, 2007.

682 Weschler, C. J., Mandich, M. L. and Graedel, T. E.: Speciation, photosensitivity, and  
683 reactions of transition metal ions in atmospheric droplets, *J. Geophys. Res.*, 91(D4), 5189,  
684 doi:10.1029/JD091iD04p05189, 1986.

685 Yu, J., III, D. R. C., Griffin, R. J., Flagan, R. C. and Seinfeld, J. H.: Gas-Phase Ozone  
686 Oxidation of Monoterpenes: Gaseous and Particulate Products, *J. Atmos. Chem.*, 34(2),  
687 207–258, doi:10.1023/A:1006254930583, 1999.

688 Zellner, R., Exner, M. and Herrmann, H.: Absolute OH quantum yields in the laser photolysis  
689 of nitrate, nitrite and dissolved H<sub>2</sub>O<sub>2</sub> at 308 and 351 nm in the temperature range 278–353  
690 K, *J. Atmos. Chem.*, 10(4), 411–425, doi:10.1007/BF00115783, 1990.

691

692

693

694

695

696

697

698

699

700

701

702 **Table 1:** Initial total organic carbon (TOC,  $\mu\text{g/L}$ ) in the given ambient cloud water sample  
703 (T#). \*Due to low sample volume, T1 is a combination of three cloud samples; TOC and pH  
704 are estimated using averages of the three samples and assuming equal volumes.

|     | T1*   | T2   | T3   | T4   | T6PA | T7/7PA     |
|-----|-------|------|------|------|------|------------|
| TOC | ~20   | 24.1 | 2.86 | 1.73 | 9.17 | 6.40/10.06 |
| pH  | ~5.49 | 3.32 | 6.37 | 4.74 | ~3.7 | 4.58/~3.7  |

ACCEPTED MANUSCRIPT

705

### Figure Captions

706 **Figure 1:** Schematic of photoreactor experimental setup. Circles with black triangles indicate  
707 pumps.

708 **Figure 2:** Timelines of  $f_{44}$  (red), normalized (to  $t=0$ ) Org/SO<sub>4</sub> (green), and O:C (blue) for T1,  
709 T2, T3, T4-1, and T7 aqueous photo-oxidations. Uncertainties (not plotted): one standard  
710 deviation equals 28% of the value, from propagation of 20% AMS quantification error  
711 through the indicated calculations. Experiment time approximates ambient oxidation time as  
712 discussed in Section 2.3.

713 **Figure 3:** Timelines of (left axes) fragment ratios normalized to the  $t=0$  value and (right axes)  
714 the mass concentration of the ‘formic acid’ fragment for T1-4, T7, and T7PA. Uncertainties  
715 (not plotted): one standard deviation equals 28% of the value, from propagation of 20% AMS  
716 quantification error through the indicated calculations.

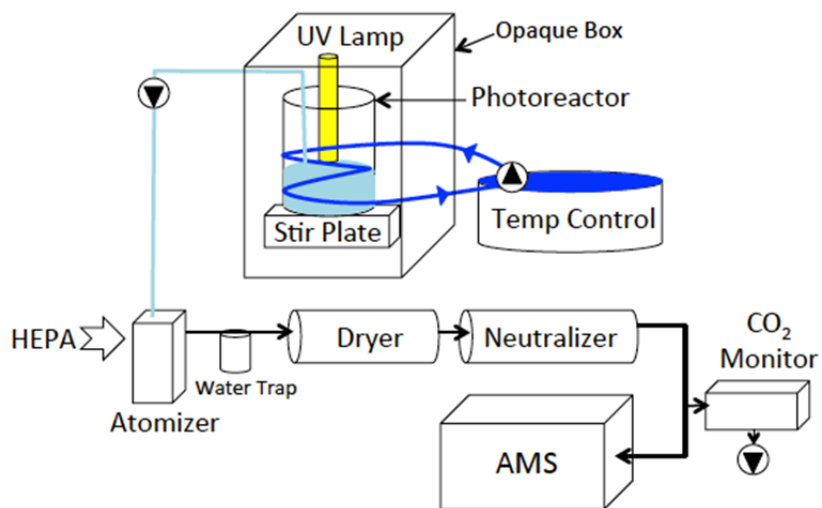
717 **Figure 4:** van Krevelen diagram showing (left) T1, 3, & 4 full photo-oxidations and (right)  
718 T2, 7, & 7PA, colored by time during experiment. Dashed red lines: Estimated oxidation  
719 state,  $OS_c=2*O:C-H:C$  (Kroll et al., 2011). Red and blue lines indicate the region usually  
720 inhabited by ambient data in the  $f_{43}$  vs  $f_{44}$  plot, as transposed to van Krevelen space using  
721 the empirical parameterization of  $f_{43}$  vs H:C and  $f_{44}$  vs O:C for ambient and chamber OA  
722 (Ng et al., 2010); grey lines represent 10% error. Linear fits: T1:  $m = -0.28$ ,  $r^2 = 0.89$ ; T2:  $m$   
723  $= 0.25$ ,  $r^2 = 0.32$ ; T3:  $m = -0.28$ ,  $r^2 = 0.64$ ; T4:  $m = -0.51$  (see text),  $r^2 = 0.46$ ; T7:  $m = -0.32$   
724 (see text),  $r^2 = 0.14$ ; T7PA:  $m = -0.45$ ,  $r^2 = 0.26$ .

725 **Figure 5:** Initial concentrations of the given cloud water constituents for given sample  
726 regimes. Units are, from left to right: Cl-Ca, HCHO, & S(IV) ( $\mu\text{N}$ ); TC-TOC (ppm); Fe &  
727 Mn ( $\mu\text{g/L}$ ). Error bars are standard deviations where data from multiple samples of the same  
728 regime were averaged. T2 consists of only one sample, so error bars are omitted.

729

730

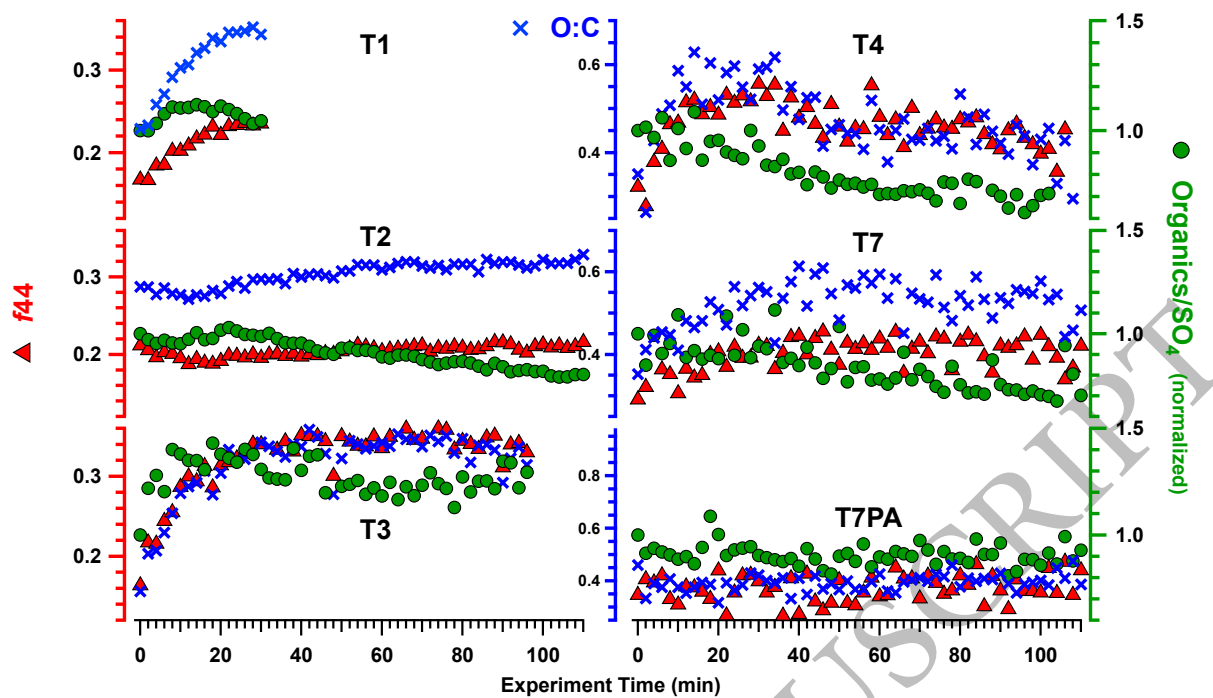
731



732  
733  
734  
735  
736  
737  
738  
739  
740  
741  
742  
743  
744  
745  
746  
747  
748  
749  
750

ACCEPTED MANUSCRIPT

Fig.1



751

752

753

754

755

756

757

758

759

760

761

762

763

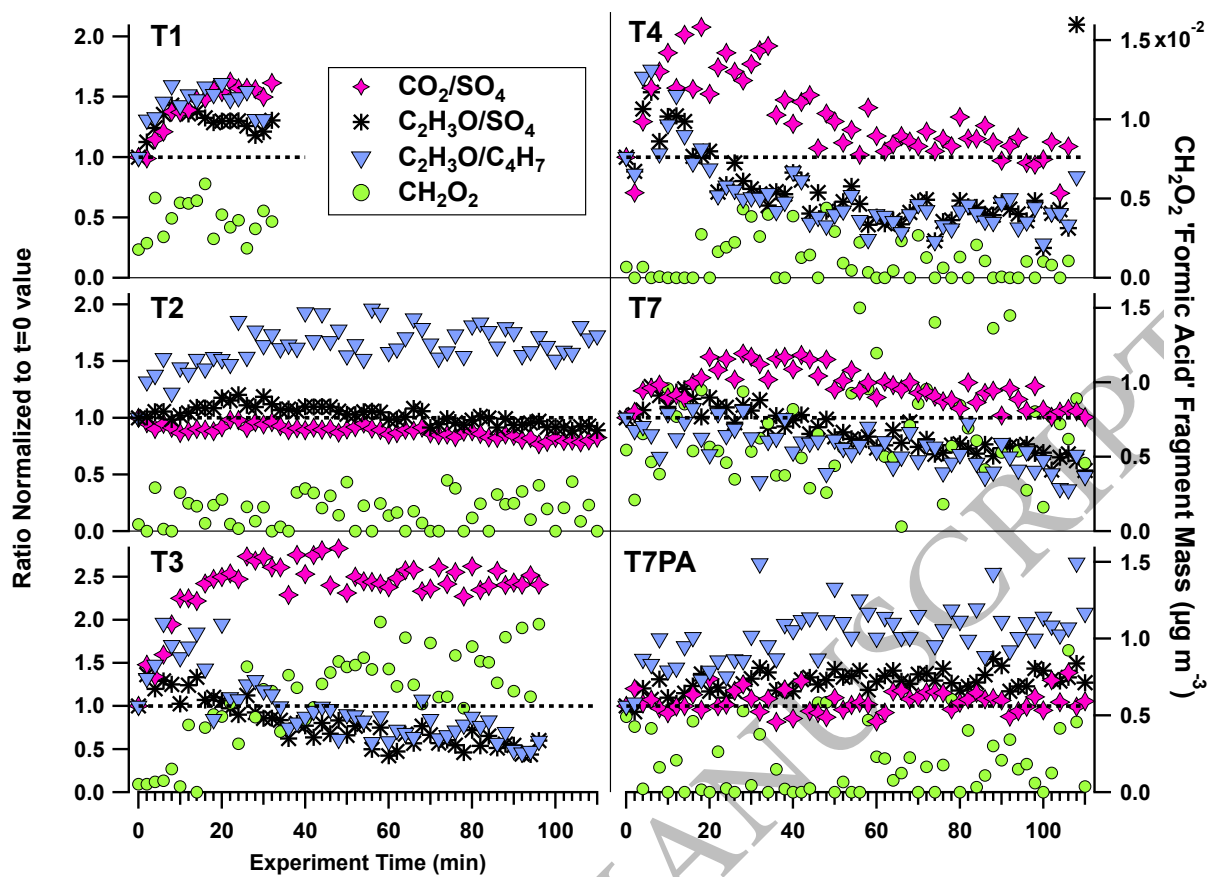
764

765

766

Fig. 2





767

768

769

770

771

772

773

774

775

776

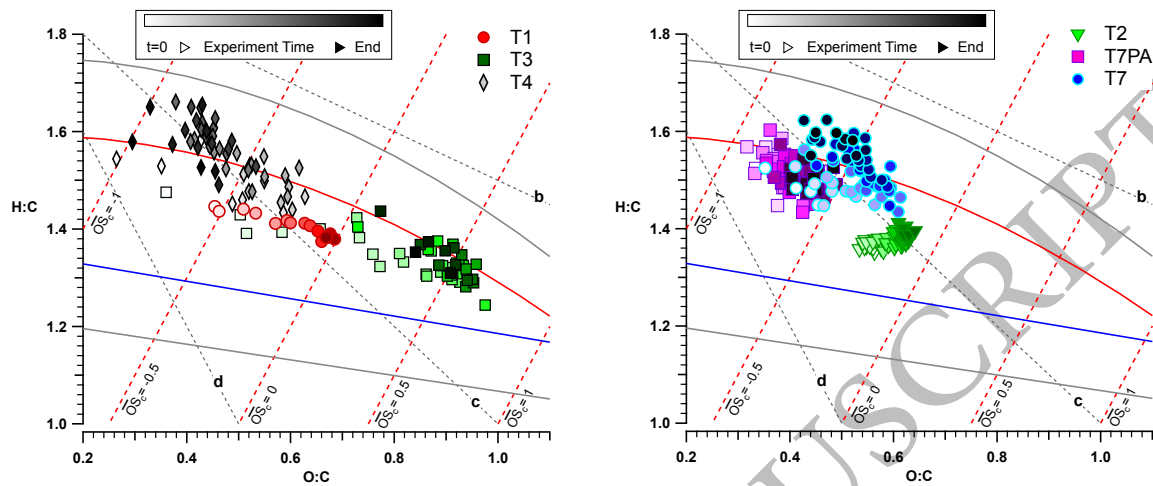
777

Fig. 3

778

779

780



781

782

783

784

785

786

787

788

789

790

791

792

793

794

795

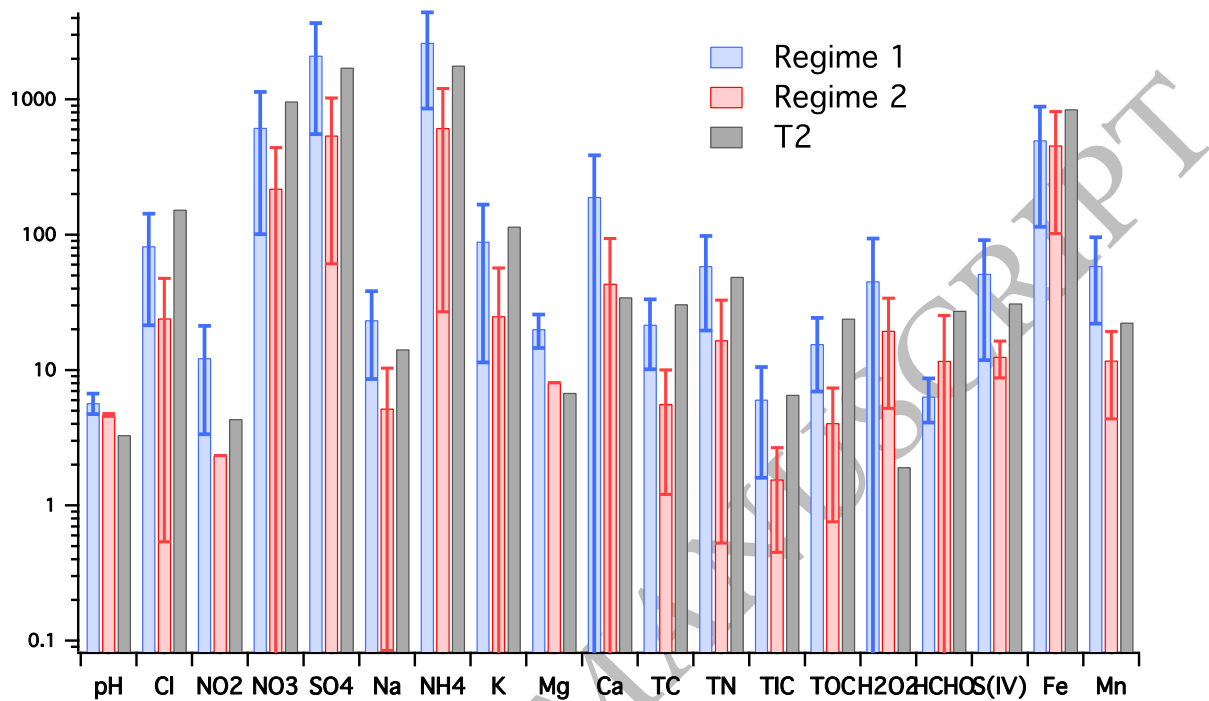
796

Fig. 4

797

798

799



800

801

802

803

804

805

806

807

808

809

810

811

Fig. 5